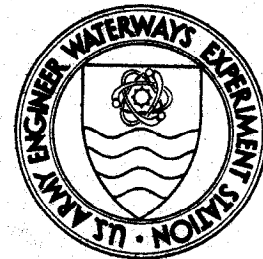


# DREDGED MATERIAL RESEARCH PROGRAM



TECHNICAL REPORT D-78-23

## CONSIDERATIONS IN CONDUCTING BIOASSAYS

by

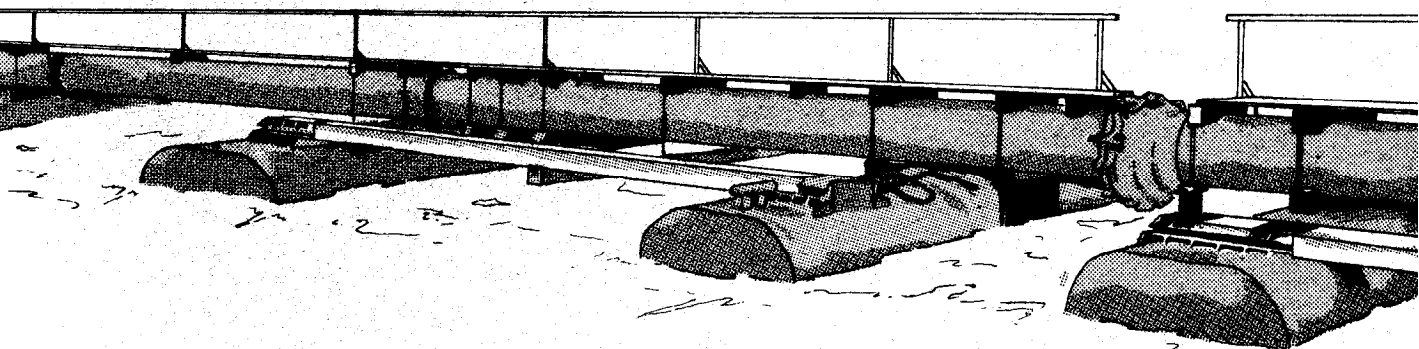
David R. Rosenberger, Edward Long, Raymond Bogardus  
Elaine Farbenbloom, Robert Hitch, Susan Hitch

Bioassay Laboratory  
WAPORA, Inc.  
Charleston, Illinois 61920

June 1978

Final Report

Approved For Public Release; Distribution Unlimited



Prepared for Office, Chief of Engineers, U. S. Army  
Washington, D. C. 20314

Under Contract No. DACW39-73-C-0134  
(DMRP Work Unit No. ID02)

Monitored by Environmental Laboratory  
U. S. Army Engineer Waterways Experiment Station  
P. O. Box 631, Vicksburg, Miss. 39180

**Destroy this report when no longer needed. Do not return  
it to the originator.**



DEPARTMENT OF THE ARMY  
WATERWAYS EXPERIMENT STATION, CORPS OF ENGINEERS  
P. O. BOX 631  
VICKSBURG, MISSISSIPPI 39180

IN REPLY REFER TO: WESYV

31 July 1978

SUBJECT: Transmittal of Technical Report D-78-23

TO: All Report Recipients

1. The work reported herein was undertaken as Work Unit 1D02 of Task 1D, Effects of Dredging and Disposal on Aquatic Organisms, of the Corps of Engineers' Dredged Material Research Program. Task 1D was a part of the Environmental Impacts and Criteria Development Project (EICDP), which had a general objective of determining on a regional basis the direct and indirect effects on aquatic organisms due to dredging and disposal operations. The study reported herein was part of a series of research contracts developed to achieve the EICDP general objective.
2. This literature review is intended as an introductory overview of aquatic bioassays, with emphasis on dredged material bioassay techniques. It provides a general discussion of bioassay theory and practice, equipment, and selection of appropriate test species.
3. Species selection is discussed at length as it is one of the most important variables influencing the usefulness of bioassays for regulatory purposes. Similar criteria have been used by many authors in species selection with the basic consideration often being adaptability of the species to laboratory conditions and compatibility with testing procedures.
4. The information and discussion published in this report provide technical personnel unfamiliar with bioassays, managers, and administrators with a general introduction to the field of bioassay. This introduction to the bioassay literature will serve as a basis and guide in developing and implementing a regulatory program based on bioassays. It is expected that the information contained herein will be of significant value to those concerned with CE dredging material permit programs.

A handwritten signature in cursive script, reading "John Cannon", is positioned above the typed name.

JOHN L. CANNON  
Colonel, Corps of Engineers  
Commander and Director

REPORT DOCUMENTATION PAGE		READ INSTRUCTIONS BEFORE COMPLETING FORM
1. REPORT NUMBER Technical Report D-78-23	2. GOVT ACCESSION NO.	3. RECIPIENT'S CATALOG NUMBER
4. TITLE (and Subtitle)  CONSIDERATIONS IN CONDUCTING BIOASSAYS		5. TYPE OF REPORT & PERIOD COVERED Final report
		6. PERFORMING ORG. REPORT NUMBER
7. AUTHOR(s) D. R. Rosenberger                      E. Farbenbloom E. Long                                      R. Hitch R. Bogardus                                S. Hitch		8. CONTRACT OR GRANT NUMBER(s)  DACW 39-73-C-0134
9. PERFORMING ORGANIZATION NAME AND ADDRESS Bioassay Laboratory WAPORA, Inc. Charleston, Illinois 61920		10. PROGRAM ELEMENT, PROJECT, TASK AREA & WORK UNIT NUMBERS  DMRP Work Unit No. 1D02
11. CONTROLLING OFFICE NAME AND ADDRESS  Office, Chief of Engineers, U. S. Army Washington, D. C. 20314		12. REPORT DATE June 1978
		13. NUMBER OF PAGES 144
14. MONITORING AGENCY NAME & ADDRESS (if different from Controlling Office)  U. S. Army Engineer Waterways Experiment Station Environmental Laboratory P. O. Box 631, Vicksburg, Miss. 39180		15. SECURITY CLASS. (of this report)  Unclassified
		15a. DECLASSIFICATION/DOWNGRADING SCHEDULE
16. DISTRIBUTION STATEMENT (of this Report)  Approved for public release; distribution unlimited.		
17. DISTRIBUTION STATEMENT (of the abstract entered in Block 20, if different from Report)		
18. SUPPLEMENTARY NOTES		
19. KEY WORDS (Continue on reverse side if necessary and identify by block number)  Bioassay Freshwater fishes Toxicity		
20. ABSTRACT (Continue on reverse side if necessary and identify by block number)  Many types of tests can be considered as forms of bioassay. The broad range of substrate types, chemical composition, geographical locations, and equipment used in dredging adds to the complexity of formulating a single bioassay procedure applicable to all dredging programs. A discussion of the bioassay principles and techniques in the literature are presented to provide a background for an understanding of aquatic bioassay testing.  (Continued)		



## 20. ABSTRACT (Continued).

Aquatic bioassays, which were developed from pharmacological drug testing techniques, often encounter complex toxicity problems. Two or more toxic compounds may be acting together or the toxicant concentration may fluctuate. The techniques used in aquatic testing may vary depending on the type of toxicant tested and the environmental parameters, such as temperature or dissolved oxygen, that need be controlled. Toxicant dosing equipment varies with the type of test species, source of water, and type of toxicant in question. Static bioassays are generally performed in glass jars or commercial aquarium tanks. Continuous-flow equipment consists of a reservoir of toxicant water which is metered into the test vessel. Extensive elaboration of the equipment has led to multiple toxicant delivery systems capable of providing a range of concentrations. The two most widely used delivery systems are the serial diluter and proportional diluter.

The selection of bioassay test species warrants careful consideration. The criteria for selection of species cited by most authors include the following not listed in order of importance: type of test; economic importance; ecological significance; geographical distribution; ease of capturing, handling, holding, and culturing; availability and local abundance; toxicity responsiveness; consistency of response to toxicity; and reproductive success under assay conditions.

The vast majority of species discussed in the bioassay literature are freshwater fish, primarily bluegill, rainbow trout, fathead minnow, goldfish, and largemouth bass. Infrequently used species generally did not fit all nine of the above selection criteria.

Recommended methods for capturing, handling, and maintaining many plants and animals for use in bioassays are presented in the report. Selection of wild species versus cultured species is the subject of continued debate. Wild specimens represent the fauna likely to be affected, but may be very difficult to maintain in the laboratory, while cultured specimens may be much more adaptable to bioassay testing.

## SUMMARY

Pollutant materials become incorporated into the bottom substrates through the process of adsorption and sedimentation. Concern has developed in recent years over the possible deleterious effects to the water quality and aquatic organisms from the dredging and disposal of these contaminated sediments.

In response to Section 404 of Public Law (PL) 92-500 and Section 103 of PL 92-532, criteria have been formulated for determining the acceptability of dredged material disposal to the Nation's waters. The criteria allow, or in some cases require, bioassays to be performed on dredged material, but do not specify what type of study constitutes a bioassay.

A bioassay for the purpose of this report is defined as: a method of testing the potency or activity of a material (drug, naturally-occurring chemical, or pollutant) through the elicitation of a response (biochemical, physiological, or mortality) by a living organism, tissue, or cell. Behavioral responses are also considered in the report because of their significance in contributing to the ecological success of the species.

The screening of drugs for potency and specificity led to the development of classical bioassay principles and techniques, although many types of tests are presently considered as forms of bioassay. Aquatic bioassays are composed of three parts: stimulus, subject, and response. The application of these components to aquatic toxicity testing are complicated by the presence of multiple toxic elements more than one potentially toxic substance in many wastes and the presence of particulates which may sorb the pollutants, altering their toxicity. Calculation of the toxicity of these multiple systems has resulted in the development of mathematical formulae to estimate the concentrations which would cause mortality to the test species. Fluctuating concentrations of the toxicants have also been evaluated using mathematical formulations.

Numerous bioassay techniques have been developed for specific toxicant materials and test species. Variability of test results are

influenced by the chemical characteristics of the experimental water. Warm water decreased the survival time of fish exposed to heavy metals while cold water increased the susceptibility of fish to anesthetics. Effects of water hardness on survival times have also been reported.

There is general agreement that small organisms should be selected with at least ten used for each test. The organisms should be acclimated and tested in tanks constructed from nontoxic materials. Stainless steel or glass are most often recommended. Test water should be the same as the acclimation water and free from chemical or bacterial contaminants. The animals should be discarded if excessive loss occurs from disease or unexplained causes.

Static bioassay test chambers are almost universally constructed of either solid glass jars or commercial aquaria. Adaptation of the static tanks are reported for use with suspended sediments.

Continuous-flow bioassay equipment consists of an exposure vessel which receives a metered test material. The two most widely used multiple toxicant delivery systems are the serial and proportional diluters. The systems should require low maintenance and cease to deliver toxicant upon loss of dilution water.

In recent years there has been a tendency for toxicologists to incorporate on-site bioassay studies into their analysis of environmental toxicity. On-site studies allow for the natural variability of the toxicant in the receiving water. The equipment in simplest form consists of a holding box which is placed in the effluent water. Other systems consist of substrate material which is colonized by benthic organisms. A further advancement in on-site testing is to bring a portable bioassay laboratory to the test site.

A major requirement for conducting bioassay is an adequate supply of test organisms. Most authors agree that bioassay test species should possess certain qualities. The criteria cited by most authors for selection of test species include:

1. Type of test
2. Economic importance
3. Ecological significance

4. Geographical distribution
5. Ease of capturing, handling, holding, and culturing
6. Availability and local abundance
7. Toxicity responsiveness
8. Consistency of response to toxicity
9. Reproductive success under assay conditions

In tests evaluating site-specific impacts, species commonly found in the receiving water are often recommended. If a test is to be performed to establish standard concentration limits, common laboratory test organisms are recommended.

Organisms of economic importance should be considered as test species. The economically important freshwater species are almost entirely the Salmonid, Centrarchid, Cyprinid, and Catostomid fishes. Those species which are of economic importance also play a major role in the ecosystems to which they belong. Species which are important food chain species should also be considered as test species since they provide the biomass for higher organisms and they are known to bioconcentrate selected environmental contaminants. Test species which are healthy, unstressed and disease-free are essential in conducting a bioassay. Health problems may accentuate or mask the response of the organism, producing misleading results.

Species vary widely in their response to specific pollutants. The most meaningful ecological information is obtained by using a species of median toxicant sensitivity. Concurrently, the species should display a constant response to the test solution over time and between males and females.

Numerous authors noted some of the difficulties of using certain species as test organisms. The most common problems mentioned include: high mortalities, cannibalism, aggression, parasitism and feeding problems. Crustaceans commonly were reported to experience problems with cannibalism. Arthropods were very sensitive during molting.

Numerous species of plants and animals require special care in handling. Excessive handling can result in physiological stress which may influence the bioassay results. Methods and apparatus for culturing

some species of edible plants and animals have been available for many years. Others have developed only recently. The use of specimens taken from cultured stock ensures that the test organisms are free from pollution contamination and are descendants of the same genetic stock. Many of the economically important fish, crustaceans, and mollusks have been raised in the laboratory.

Wild-caught and cultured species can vary in their sensitivity to many environmental contaminants due to their differing genetic history. Species populations in a continually polluted stream are more highly tolerant to these contaminants than are species of an unpolluted source. Laboratory cultured specimens have been reported to develop an abnormal sensitivity to many pollutants. Selection of wild-caught versus cultured specimens should be judged by these factors.

## PREFACE

The work described in this report was performed under Contract No. DACW 39-C-0134, titled "Considerations in Conducting Bioassays." The research was sponsored by the Office, Chief of Engineers (DAEN-CWO-M) under the civil works research program, "Dredged Material Research Program."

The research was conducted under the supervision of Mr. David R. Rosenberger, Manager, Bioassay Laboratory, WAPORA, Inc., Charleston, Illinois. Mr. Edward Long, Raymond Bogardus, Elaine Farbenbloom, Robert Hitch, and Susan Hitch assisted in the report preparation.

The contract was monitored by Dr. Paul Becker, assisted by Ms. Susan Palmer, under the direct supervision of Dr. John Keeley, Project Manager. General supervision was provided by Dr. John Harrison, Chief, Environmental Laboratory, U. S. Army Engineer Waterways Experiment Station (WES).

Director of WES during the period of the contract and the preparation of the report was COL J. L. Cannon, CE. Technical Director was Mr. F. R. Brown.

# CONTENTS

	<u>Page</u>
SUMMARY . . . . .	1
PREFACE . . . . .	5
LIST OF TABLES . . . . .	7
CONVERSION FACTORS, U. S. CUSTOMARY TO METRIC (SI)	
UNITS OF MEASUREMENT . . . . .	8
PART I: INTRODUCTION . . . . .	9
Background . . . . .	9
State-of-the-Art . . . . .	10
Source of Material . . . . .	10
Bioassay Definition . . . . .	10
Scope of Report . . . . .	11
PART II: PRINCIPLES OF AQUATIC BIOASSAY . . . . .	12
Principles of Bioassay . . . . .	12
History of Aquatic Toxicity . . . . .	15
Scope of Aquatic Bioassay . . . . .	16
Variability of Aquatic Bioassay Testing . . . . .	18
Current Proposed Standard Procedures . . . . .	23
Bioassay Equipment . . . . .	26
PART III: TEST ORGANISMS . . . . .	36
Criteria for Selection of Test Organisms . . . . .	36
Species Most Commonly Selected . . . . .	43
Species Infrequently Used . . . . .	44
Test Species Previously Used in Turbidity Bioassays . . . . .	47
Test Species Which Present Special Problems . . . . .	48
Test Species Unsuitable for Bioassays . . . . .	49
Recommendations for Capturing, Handling and Maintaining Wild-Caught Organisms . . . . .	49
Recommendations for Culture of Test Organisms and Stock Populations . . . . .	50
Wild-Caught Versus Stock-Cultured Test Organisms . . . . .	58
Recommended Test Species . . . . .	59
PART IV: DREDGED MATERIAL BIOASSAY DEVELOPMENT . . . . .	63
REFERENCES . . . . .	69
TABLES 1-7	

## LIST OF TABLES

<u>No.</u>	<u>Title</u>
1	Commonly Used Species, Reported in Bioassay Literature
2	Infrequently Used Species
3	Test Species Previously Used in Turbidity and Turbidity-Related Bioassays
4	Test Species Which Have Been Used Successfully in Bioassays but Present Some Special Problems
5	Rankings of Fish Species on a Scale of 1-8 as Suitable Bioassay Animals
6	Recommended Fish Species Previously Used in Bioassay Research
7	Species Recommended for Use in Bioassays



CONVERSION FACTORS, U. S. CUSTOMARY TO METRIC (SI)  
UNITS OF MEASUREMENT

Units of measurement used in this report can be converted as follows:

<u>Multiply</u>	<u>By</u>	<u>To Obtain</u>
inches	25.4	millimetres
feet	0.3048	metres
gallons (U. S. liquid)	0.003785412	cubic metres

# CONSIDERATIONS IN CONDUCTING BIOASSAYS

## PART I: INTRODUCTION

### Background

1. Industrialization and technical advancement have resulted in the production of new and exotic chemical compounds and increased amounts of waste products. Chemical control of undesirable plants and animals is now almost universal. Petrochemical wastes, mine-drainage waters, and industrial effluents are transported to the waterways through erosion, air deposition, and direct discharge. These pollutant materials become incorporated into the bottom substrates through the process of adsorption and sedimentation. Concern has developed in recent years over the possible deleterious effects to the water quality and aquatic organisms from the dredging and disposal of these contaminated sediments.

2. Contingent to the River and Harbor Act of 1970, the Environmental Protection Agency (EPA) formulated criteria for determining the acceptability of dredged material disposal to the nation's waters. However, a review of these early criteria by the Corps of Engineers<sup>1</sup> found that direct application to a particular dredging project was extremely difficult. A critical review of the literature relevant to the factors contained in the criteria provided insufficient data on which to base a decision as to the acceptability of dredged material for open-water disposal.

3. Bioassay is among the tests recommended in the criteria. Unlike phosphorus, sulfides, or others in the list, a bioassay is not a physical parameter which can be measured analytically, but is itself a technique for measuring the environmental effect of the other parameters. The criteria did not specify what physical and environmental conditions of dredged material would warrant the inclusion of bioassay tests as "appropriate and pertinent." Technical questions relevant to

implementation of regulatory bioassays include:

- a. What type of biological test constitutes a bioassay?
- b. Which species should be studied?
- c. What types of biological responses should be observed?
- d. What duration of time should the organisms be subjected to the dredged material?
- e. What degree of biological response would determine the unacceptability of open-water disposal of dredged material?

#### State-of-the-Art

4. The preceeding questions do not imply a lack of knowledge for conducting aquatic toxicity bioassays. Extensive research in this field has resulted in the establishment of field and laboratory methods of conducting aquatic bioassays. These questions are, however, indicative of the "state-of-the-art" of conducting bioassays on dredged material. The general lack of published dredged material bioassay literature is evidence to this fact.

#### Source of Material

5. The material contained in this report is the result of an in-depth literature search and site visitation to recognized laboratories engaged in bioassay research where the problems and techniques of performing biological assays on dredged material were discussed. The equipment, testing procedures, and test species utilized for bioassay research were reviewed for their application to dredge material bioassay. Over 1500 research and literature review publications were analyzed for their contribution to this assessment. Additional unpublished literature was gathered from members of the scientific community engaged in bioassay and/or dredged material research.

#### Bioassay Definition

6. Pharmacologists have for many years used biological systems to test the therapeutic, effective dosage of drugs and the response

elicited by a chemical. In more recent years, bioassays have been expanded to include environmental pollution biology. This expansion of use has often resulted in a misconception of what constitutes a bioassay. A bioassay, for the purpose of this report, is defined as a method of testing the potency or activity of a material (drug, naturally-occurring chemical, or pollutant) through the elicitation of a response (biochemical, physiological, or mortality) by a living organism, tissue, or cell.

7. Although not generally included in the definition, behavioral responses are also considered in this report because of their significance in contributing to the ecological success of the organism. Examples of behavioral responses include those which produce an avoidance reaction to a naturally-occurring chemical or pollutant; failure of the organisms to initiate or respond to mating behavioral activity; and changes in predator-prey interactions.

8. Beyond the scope of the above definition are those responses which produce a carcinogenic, mutagenic, or teratogenic reaction within an organism. Deletion of these responses from the scope of the definition should not be construed to imply a lack of their importance. To the contrary, their contribution to the ecological success of a species may prove as significant as an acute toxic mortality. Discussion of these topics in relation to dredged material must await development of an adequate data base.

#### Scope of Report

9. Many types of tests may be considered as forms of bioassay. Further, dredging activities encompass a wide range of substrate types, chemical composition, geographical locations, and equipment used. Selected environmental contaminants incorporated into dredged material may present special problems in bioassay investigations. Specialized techniques and equipment used in conjunction with these contaminants are presented. A discussion of bioassay principles and techniques are presented to provide an appreciation of the complexity of performing bioassays on dredged material.

## PART II: PRINCIPLES OF AQUATIC BIOASSAY

### Principles of Bioassay

#### Reason for study

10. The field of toxicology embraces a broad range of testing, interpretation, and application. The screening of drugs for potency and specificity has led to the development of classic bioassay principles and techniques.<sup>2</sup> A basic familiarity with the basic toxicological principles and methods of study is helpful for the interpretation and evaluation of bioassay results in aquatic toxicity testing.

#### Test for potency

11. Experimental pharmaceutical compounds undergo a series of bioassay tests to evaluate their biological potency. Reference compounds are selected as standards in the evaluation of potency. The experimental compound is tested against the standard in three types of studies.<sup>3</sup> The first is a measure of minimal potency. When given in low doses, the test compound must demonstrate an equivalent potency to that of the standard. A second study entails adjusting the dosage of the test compound to produce a submaximal response equivalent to the standard. Combination of the test results of these first two studies provides a range over which the drug may effectively be administered pharmaceutically (i.e., a minimum and maximum safe dosage). The dose-response curve is determined from the third study. The response observed at various dosage levels are plotted to evaluate the median effective dose. (The dosage level at which a 50 percent maximum response is observed.) The median effective dose curve is then used to evaluate the dosage schedule of the experimental drug. A modification of the latter type of test is the one most commonly used in aquatic toxicity testing.

#### Types of responses

12. Dose-response curves in pharmacological assay tests are based on two classical types of responses -- all-or-none and graded.<sup>4</sup> The all-or-none type is used to measure the threshold dose of the material

being tested. As implied by its name, the response observed is specific and distinct. Examples of all-or-none responses include mortality, avoidance, and nerve stimulation.

13. In all-or-none response tests, potency is inversely proportional to the smallest amount of drug that will cause the reaction. This level can be determined from curves based on a series of tests of separate subgroups exposed to different doses.<sup>3,4</sup>

14. In graded response, the amount of the reaction is a function of the amount of drug administered. When the response is plotted against the log-dose, a sigmoid curve is generally produced. Goldstein et al.<sup>2</sup> discuss exceptions to this rule. The area of the curve usually studied is the central linear portion of the graph. If two chemically similar drugs of different potency are administered in a series of tests, one may expect parallel log-dose response curves to be produced. This parallelism suggests that the mode of action of the drugs are similar. Nonparallel response curves indicate that the drugs affect different biological mechanisms or receptors (e.g., different enzymes or different sites on an enzyme).

#### Interpretation of results

15. The drug or toxicant undergoes a complex series of physical and chemical steps before its action is manifested as a biological response. If the test is not carefully designed, ambiguous or misleading results may develop.

16. The indiscriminate use of mathematical formulae, especially at the limits of their applicability, may yield improbable results. For example, in the formula:  $\text{Concentration} \times \text{Time} = \text{Constant}$ , an infinitely large dose may be interpreted to produce an instantaneous response. However, this has been demonstrated not to be the case.<sup>5</sup>

17. The response that is being measured in the experiment must also be carefully defined. For example, in biological systems when a population of algae cells exhibits a decrease in growth rate after being exposed to a toxic substance, the decline might be due to the lowered mitotic rate of all the cells or to the death of some of the cells with a natural growth rate in the remaining cells. In the former,

the reaction is a graded response, but in the latter, an all-or-none response is indicated for each individual cell.

#### Application of results

18. Bioassay experimentation has lead to the development of some specialized forms of statistical treatments. Often two parameters of biological response are of interest: the median response and the threshold response. These responses can be expressed in either units of time or concentration of test chemical. The median response is of importance in representing the average response one may observe in a population exposed to the drug or toxic material. The threshold response signifies the point where an organism is first affected by the material. Threshold responses are most often used to determine a safe dose.

19. Gaddum<sup>6,7</sup> reviewed the state-of-the-art of bioassays, their mathematical relationship in pharmacology, and discussed the problems associated with determining a safe dose of toxicant. He felt that a safe dose should be established by setting arbitrary limits or application factors of 1/10 or 1/100 of a known threshold concentration. He recognized the variability of biological response and stated that these limits may not be applicable in cases of high biological variability. An alternate proposed method of determining a safe dose was to arbitrarily select three or six standard deviations below the median log-dose response concentration. The latter method was felt to be less costly and time consuming as both the median response and safe concentration could be ascertained from a single study.

20. Gaddum<sup>7</sup> believed that changes in such parameters as weight, blood pressure, and amount of food consumed may also be used as a measure of toxic effect. The use of the three to six standard deviations from the median dose necessary to affect this change may be more meaningful than acute mortality.

#### Statistical treatment

21. Statistical treatment of the median response is explored by Bliss<sup>4</sup> for time-mortality response curves. He describes the use of probit statistical analysis to convert the sigmoid time-response curve to a linear function. The tedious mathematical computations required

for this conversion have limited the usage of this technique. Most researchers have not been able to justify the time necessary to attain the precision possible using Bliss' methods.

22. Litchfield and Wilcoxon<sup>8</sup> proposed a graphic method for approximating the median effective dose. Data in its original form is plotted on probability graph paper and the confidence interval calculated by means of nomograph tables. Litchfield<sup>9</sup> presented a similar technique for calculating time-response curves. Recently, computer programming has offered more rapid usage of the probit computations.

23. Alternate methods to probit analysis are also available. Burdick<sup>10</sup> describes a graphical method for translating the sigmoid dose-response curve to a straight line. Harris<sup>11</sup> presents confidence limits and significance values for dose-response curves using a moving average-angle method. Further insight into the statistical treatment of bioassay data is available in the excellent work by Sprague.<sup>12</sup>

## History of Aquatic Toxicity

### General reviews

24. The development of aquatic bioassay began in the 1860's in England and in the 1910's in the United States. Penny and Adams (cited in Erichsen-Jones<sup>13</sup>) worked with the complex effluent from a dye factory and the individual components of the effluent. In the early 20th century in the United States Shelford and Wells (cited in Erichsen-Jones<sup>13</sup>) explored the suitability of different species for bioassay.

25. There have been several extensive general reviews of the bioassay literature produced over the years.<sup>14-16</sup> In 1964 Erichsen-Jones<sup>13</sup> reviewed the literature and techniques of bioassay up to the early 1960's. His work separated techniques by the type of substance being tested, e.g., metal salts, respiratory depressants, detergents, pesticides, and described equipment as well as results.

### Specific aspects

26. Specific aspects of bioassay have also been reviewed over the years. Katz<sup>14</sup> covered the standard procedures of bioassay as they have



developed. Weiss<sup>17,18</sup> discussed the effects of organic pesticides on fish and Muirhead-Thomson<sup>19</sup> reviewed public health aspects of the use of pesticides. The toxic properties of metals in aquatic environments have also received widespread attention in review.<sup>20-23</sup> Doudoroff and Katz<sup>24</sup> reviewed bioassays of alkalis, acids, and inorganic gases, and Ball<sup>25</sup> reviewed fish bioassays and ammonia. The chemical aspects of the formulations reported in bioassay research have also been assessed.<sup>26</sup>

#### Water quality criteria

27. The problem of water quality criteria has been the subject of several papers.<sup>27,28</sup> In 1968 the Federal Water Pollution Control Administration<sup>29</sup> published Water Quality Criteria, which defined the standards and reviewed the earlier work on toxic substances and environmental variables. In 1956 the Ohio River Valley Water Sanitary Commission (ORSANCO) reviewed temperature limits, dissolved solids, chlorine, fluoride, and dissolved oxygen.<sup>30</sup> The direct and indirect effects of pH on European freshwater fish have been reviewed.<sup>27,31</sup> Lloyd<sup>16</sup> reviewed water quality problems for freshwater fisheries and recommended bioassays and field observation as the methods for achieving water quality standards.

28. The setting of criteria for estuarine organisms is a complex problem due to the physical diversity of the ecosystem as well as the diversity of life involved. In an estuary, invertebrates such as mussels and oysters are as important to sportsmen and commercial fishermen as are fish species.<sup>32</sup>

#### Scope of Aquatic Bioassay

##### Application of toxicological principles

29. Bioassay components. The application of toxicological principles to aquatic bioassay testing has been the subject of several reviews. Alderdice<sup>33</sup> noted that aquatic bioassays consist of three parts: 1) stimulus--such as drug, insecticide, or industrial waste; 2) subject--such as tissue, organ, or whole animal; and 3) subject response--such as death, change in weight, blood pressure, oxygen

consumption, or some other biological activity that has been evoked by the stimulus. The stimulus-response relationship for aquatic bioassay were reported to be of two separate types: lethal or acute, and sublethal or chronic.

30. Joint toxicity. Sprague<sup>12</sup> described a method for evaluating the joint toxicity of a mixture of compounds from their respective lethal threshold concentrations. Toxic units are presented which are equal to the actual concentration in the test solution divided by the lethal threshold concentration of the components of the solution. The toxic units of the respective components are added together, and if the total is greater than 1.0, the mixture is said to be lethal.

31. Field trials have also been made on predicting the toxicity of polluted rivers by adding toxic units. Sprague<sup>12</sup> reports that the toxic units may not be additive at low concentration, and the predicted total toxicity would be overestimated. Other cases of joint toxicity not covered by simple addition include those where two toxicants of the solution combine chemically.

32. Toxicant fluctuation. Fluctuating concentrations of effluent discharges present another special problem in aquatic bioassay testing.<sup>34</sup> Laboratory tests usually provide a constant dosage of test solution; whereas, in natural situations the concentration of contaminant and the volume of discharge varies from day to day or during the course of a day. Several techniques have been proposed for analyzing the toxicity of fluctuating effluents.

33. The first technique is merely to roughly associate the peaks of effluent discharges with observed fish mortality. A second technique involves calculating the survival times to concentrations of toxicants in laboratory studies. A curve of survival time versus exposure concentration is then produced. If environmental levels exceed those of the laboratory time-exposure curve, toxicity may be expected. A third method is based on the assumption that the rate of death is constant for a toxicant concentration. The rate of death per unit of time is then calculated at each level of toxicant. A peak of high toxicant discharge was then evaluated by plotting it as a graph of death versus

time. Counting the squares under the curve showed that fish died when accumulated death rates equalled unity. A fourth approach suggests that survival time may be estimated by calculating the average concentration of toxicant. If the average concentration is above lethal levels, fish may not be expected to survive. This approach may not be a true predictor of toxicity if exposure levels are at or near threshold levels.

#### Sublethal effects

34. One problem of acute toxicity testing is that the test cannot explain many of the ecological changes which occur in nature. Sublethal responses often occur when test organisms are exposed to concentrations much lower than necessary to cause an acute lethal response, but for much longer exposure periods. However, while certain sublethal changes, such as reproduction, are important, the ecological significance of other sublethal changes is highly questionable.<sup>35</sup> The physiological action of a toxicant is the key to predicting sublethal effects and depends on knowledge about the normal internal structure and functioning of the animal. There are several approaches that can be made. Histology is a very important tool for determining mode of action, but the primitive state of knowledge about aquatic organisms hampers this subject. There is no histological atlas for even a given species of fish. Physiology and biochemistry are also promising areas. Urine production, hematology, and blood chemistry have been tested in bioassay.

### Variability of Aquatic Bioassay Testing

#### Variety of toxicants tested

35. Techniques for specific compounds. Aquatic toxicity investigations have produced numerous bioassay techniques for the study of specific types of compounds and test species. Because of these numerous techniques for different compounds and species, a single standardized bioassay technique has not been forthcoming.

36. Selected representatives of these specialized methodologies are presented in the following section to demonstrate the specialized problems which face the bioanalyst. Citation of specific techniques

does not constitute endorsement of the methods presented, but merely an attempt to provide total understanding of the bioassay literature.

37. Pulp mill waste. Warren and Doudoroff<sup>36</sup> discuss the development of methods for using bioassays in the control of pulp mill waste disposal. Discharges of these materials are of concern because they may create potentially toxic conditions for aquatic life, they may deplete the dissolved oxygen concentrations in the aquatic environment (possibly enhancing the toxicity problem), and/or they may promote the growth of slime or bacterial periphyton. The causative agent(s) responsible for any observed toxicity is largely unknown due to the complex composition of the effluents and the numerous steps in the manufacturing process. However, the complexity of pulp mill effluents has produced a wide range of reported toxicity characteristics that may fluctuate daily or even hourly. One explanation for the observed results is that potentially toxic, volatile materials may be lost to the atmosphere under laboratory conditions. Because of the potential loss of volatile solids, the high oxygen demand and potential changes in the chemical nature of pulp mill effluents, static bioassay evaluations may provide variable and not very meaningful results. Betts et al.<sup>37</sup> reviewed the test requirements of pulp mill effluent studies and concluded that continuous-flow studies were necessary to properly maintain test conditions and evaluate such discharges.

38. Betts et al.<sup>37</sup> attempted to evaluate the various components of the pulp mill waste treatment process. Miniature constant-flow screening was performed on Atlantic salmon (Salmo salar), followed by full-scale test conditions. Detailed descriptions are given for the experimental test tanks, cooling systems, stirrers, retention troughs, circulating pumps, temperature controls, and oxygen analyzers.

39. Oils and oil dispersants. Oil spills occurring along the coastlines of the continents of North America and Europe have led to the localized disruption of marine systems. Hazards to marine and estuarine fauna from the transport of oil, oil spills, and the dispersants used in cleaning up these oil spills have been studied using numerous bioassay techniques.

40. Crude and refined oils contain thousands of compounds with differing solubilities, volatilities, and toxicity. Toxicity data are available for various petroleum products, but usually the physical behavior of the oil in water mixture is not fully examined.

41. LaRoche et al.<sup>38</sup> conducted static bioassay investigations of the effects of oils and oil dispersant mixtures on the mummichog, a sandworm (Nereis vireus), and grass shrimp (Palaemonetes vulgaris). Homogenization of the solutions in water resulted in an increased toxicity to the species tested. The dispersants were found to vary in their primary affinity; some more soluble in seawater and others more soluble in the petroleum derivatives.

42. Lichatowich et al.<sup>39</sup> contend that crude and refined oils should be studied under continuous-flow conditions. They felt that the aeration required for static tests would accelerate the loss of volatile aromatics believed to be the prime contributors to the oil toxicity.

43. One of the most important considerations of the bioassay apparatus developed by Lichatowich et al.<sup>39</sup> was the size and shape of the test vessels, due to the changes in toxicity in shallow tanks or tanks with a large surface area. Two phases developed in the test aquaria. A slick containing most of the oil developed on the surface of the water, while some oil was present in the water column as soluble oil fractions or dispersed droplets.

44. Ecological behavior of the test fish influenced susceptibility to the oil. Benthic fish, such as sculpin and flounder, remained on the tank bottoms and were least susceptible to the toxic effects; whereas pikefish and capelin positioned themselves close to the surface and were exposed to the more toxic effects of the oil. Those which remained in midtank demonstrated an intermediate sensitivity.

#### Variations of water quality

45. Temperature. One of the more important and most easily measurable factors that influence the outcome of a bioassay test is temperature. Because temperature can be readily controlled, more is known about this one factor than any other in the environment. Temperature changes can affect the activity, metabolism, and behavior of the

organism. The chemical and physical state of the toxicant, and the rate of uptake of the toxicant are also influenced by temperature.

46. It is generally recognized that there is a range of temperatures in which an organism can exist. The upper and lower lethal temperatures are dependent on the acclimation temperature. When the lethal temperature, both upper and lower, are put in the form of an area of response, it is called a zone of tolerance. Factors which affect these lethal temperatures include dissolved oxygen, pH, species, age, and size. Alderdice<sup>40</sup> exposed coho salmon to sodium pentachlorophenate, with a change of temperature, salinity, and dissolved oxygen. Analysis of the results indicated multiple interactions of the test variables.

47. The effect of temperature on the tolerance of fish to heavy metal poisoning has been reviewed.<sup>41</sup> He concluded, in general, that a decrease in temperature usually increases the survival time of fish in toxic solutions. Since low temperatures generally increase survival time of fish in toxic solutions, a lethal threshold concentration may not be applicable.

48. Durham<sup>42</sup> reviewed the influence of temperature on the toxicity of pesticides. He concluded that temperature interaction with biological systems are complex and no general rules can be applied.

49. Anesthetic gases dissolved in water are known to affect goldfish.<sup>43</sup> Fish acclimated to a range of temperatures received known concentrations of anesthetics. It was found that a rise in temperature resulted in a fall in anesthetic potency.

50. Relatively little information is available on the effect of heat upon microorganisms. Certain microorganisms do vary greatly in their temperature tolerance. In general, diatoms grow best between 15 and 40°C.<sup>44</sup> Protozoa are also able to live only within a rather narrow temperature range although the encysted or inactive stages extend this range considerably. Gradual acclimation also shows that microorganisms can adapt to higher temperatures if the changes are small and gradual.

51. Other water quality parameters. Other water quality parameters, such as pH, alkalinity, and hardness, have been shown to cause variability in the results of toxicity tests. Lloyd<sup>41</sup> cautions that

the extent of the difference in toxicity of heavy metals appears to vary for different levels of calcium concentrations. Some differences might be due to the decreasing solubility of the metals as the water becomes more alkaline, or the presence of calcium ions themselves reducing toxicity of the metal ions. There is the possibility that intracellular calcium content may also increase with the calcium content of the water, thereby reducing the toxicity of the heavy metals.

52. Zinc was found to be most toxic at a hardness of 50 ppm pH of 8 and least toxic at a hardness of 200 ppm and pH of 6. At a pH of 8 mechanical accumulation of zinc between gill filaments was noted followed by transfer to the epithelium and mucous. This may have caused the higher toxicity.

53. Water of low dissolved oxygen concentration has been found to increase the sensitivity of aquatic animals to toxicants. Hypothesis for this action is that respiratory rate increases with a lower oxygen content of the water, and the pumping of water over the gills increases the rate at which ions and molecules are adsorbed on the gills. Thus, the rate at which toxicants would enter a fish or aquatic animal would increase as the pumping rate of the water increases.<sup>41</sup>

54. Cairns and Schierer<sup>45</sup> observed the effects of periodic low oxygen upon the toxicity of various chemicals to aquatic organisms. Low oxygen concentrations were found to decrease the tolerance of bluegill to zinc chloride, naphthenic acid, and potassium cyanide, but not to potassium dichromate. The authors conclude that a series of threshold values exist for each chemical dependent on the dissolved oxygen content of the test water.

55. The factors affecting the resistance of rainbow trout to acid waters was investigated by Lloyd and Jordan.<sup>46</sup> Tolerance of acid waters was observed to increase with increased water hardness. Trout were also more resistant to acid drainage when 20 ppm CO<sub>2</sub> was added to the water.

56. Investigators rarely control the carbonate system and the pH system independently, even though the chemical ionic state of the metal is regulated. Roberts and Allen<sup>47</sup> give equations and procedures for

controlling total alkalinity and pH or total carbonate and pH in synthetic waters and in natural waters.

### Current Proposed Standard Procedures

#### Sources of recommended methodologies

57. There are several sources of standard testing procedures for bioassay investigations. The more widely used procedures are those proposed by the American Society for Testing Materials (ASTM),<sup>48</sup> the Environmental Protection Agency,<sup>49</sup> and Standard Methods for the Examination of Water and Wastewater (Standard Methods).<sup>50</sup> Government agencies such as the Bureau of the Sport Fisheries and Wildlife,<sup>51</sup> the California State Water Pollution Control Board,<sup>52</sup> and ORSANCO<sup>53</sup> have published additional recommended methodologies. The EPA National Water Quality Laboratory<sup>54,55</sup> in Duluth, Minnesota, has published detailed methods for individual species.

58. The lack of a single uniform bioassay procedure to be used throughout the scientific community is due to several reasons. The most obvious reason is that individual authors differ in their opinion of the best research approach. Also, the testing of selected toxicants may require specialized procedures or conditions. In addition, the fish species selected for the test may place specific requirements on the size of the test tanks and rate of water replacement.

#### Fish

59. The ASTM<sup>48</sup> procedures specify that the fish preferably should be less than 8 cm long and weigh less than 5 g, and as a general rule that the larger fish in a set of tests in a tank should not be more than 1.5 times the length of the smallest one. Standard Methods<sup>50</sup> propose very similar standards, but define small fish as not more than 7.5 cm. The Fish Control Laboratory<sup>51</sup> uses fish between one-half and 2 g each. They recommend the range should not exceed 15 percent from the smallest to the largest fish. For testing the brook trout, EPA<sup>54</sup> recommends beginning with yearling fish with a total length of all the fish distributed between the tanks measured by a photographic method.



60. The EPA<sup>55</sup> gets their recommended fathead minnow, Pimephales promelas, eggs and larvae from stock cultures maintained in Duluth, Minnesota. At the beginning of the test 40 to 50 eggs or 1 to 5 day old larvae are distributed between the tanks. The fish are also measured by the photographic method at 30 to 60 days. Most authors are in agreement that at least ten individuals should be held at each concentration. The ASTM<sup>48</sup> standards recommend an acclimation time of 2 weeks, or up to one month if the fish is a cold water species. Standard Methods<sup>50</sup> recommends a ten-day acclimation period, while EPA<sup>49</sup> suggests at least two weeks.

#### Test tank

61. ASTM<sup>48</sup> suggested nontoxic tanks and other accessories be used in bioassay. It is important that the animal holding tanks are located away from any mechanical or physical disturbance, and they suggest a central drain so that wastes and feces can be removed. Standard Methods<sup>50</sup> mentioned that the test tanks should be large enough that the fish are not crowded. The EPA<sup>55</sup> regulations for the fathead minnow indicate the use of glass or stainless steel tanks with glass ends. For spawning purposes, there are two types of tanks which can be used. The first tank is 1 x 1 x 3 feet\* with one of the ends of the tank screened off in a 1 x 1 x 1 foot chamber for separation of the young. The second type is 1 x 1 x 2 foot tank with separate tanks for the young totaling at least 1 cubic foot. The EPA<sup>54</sup> regulations for brook trout recommend 1.3 x 3 x 1 foot spawning tank with the water at least 1 foot deep, and tanks for rearing the juveniles 7 x 15 x 5 inches wide with a depth of 5 inches of water. The ASTM<sup>48</sup> recommend at least 6 inches of water, and Standard Methods<sup>50</sup> recommends the water be 15 cm deep. For their laboratory experiments, the Bureau of Sport Fisheries and Wildlife<sup>51</sup> uses almost exclusively 1-gallon pickle jars, which are discarded after use. They also have two types of containers for outdoor tests. One type is a wading pool with about a 7-foot diameter vinyl-lined tank, 2.5 feet

---

\* A table of factors for converting U. S. customary to metric (SI) units of measurement is presented on page 8.

deep, with a total capacity of 1000 gallons. Aquatic plants are added to these tanks before testing. The second type of test contains the Bureau of Sport Fisheries and Wildlife<sup>51</sup> use are 20-foot raceways for flowing water. These have disposal vinyl liners.

#### Test water

62. ASTM<sup>48</sup> recommends unchlorinated well water or a clean surface water source for the test diluents. Sprague<sup>35</sup> reminds the reader that even carbon-filtered tap water may be poisonous because small quantities of chlorine may remain. Synthetic waters are also available for both hard and soft freshwater and seawater. The Bureau of Sport Fisheries and Wildlife<sup>51</sup> uses reconstituted water for their laboratory tests. Standard Methods<sup>50</sup> recommends either unpolluted receiving water from upstream of the effluent or a reconstituted water for static tests. Filtered tapwater or natural water would also be acceptable in these tests. When a synthetic water is used, acclimation of the test fish is required for 10 to 30 days. EPA requirements for brook trout<sup>54</sup> and fat-head minnow<sup>55</sup> suggest using water from a well or a spring or very clean surface water. As a very last resort, dechlorinated carbon-filtered tap water is acceptable when ultraviolet sterilization procedures are employed.

63. Water flow recommended by ASTM<sup>48</sup> is 1.44 liters of water per gram of fish per day. Aeration is considered necessary in the test tank. Standard Methods<sup>50</sup> and the Bureau of Sport Fisheries and Wildlife<sup>51</sup> suggested not over 1 g of fish per liter of water. For the EPA tests on the fathead minnow<sup>55</sup> and the brook trout,<sup>54</sup> respectively, the flow rates were recommended to be 6 and 10 tank volumes per 24 hours. Standard Methods<sup>50</sup> recommends withholding food for 48 hours before the test for short-term and static tests. The Bureau of Sport Fisheries and Wildlife<sup>51</sup> recommends withholding food for 96 hours before testing. They also recommend acclimating the fish to the test water without the toxicant for 24 hours prior to the test.

#### Chemical measurements

64. ASTM standards<sup>48</sup> recommend testing temperature, pH, and dissolved oxygen (DO) in the test tanks on the same schedule as the level

of toxicant is measured. A pH between 7.8 and 8.3 was recommended as a standard. Temperature for cold-water fish should be 15°C with 25°C  $\pm$  2°C recommended for warm-water fish. Dissolved oxygen should be held at saturation level. EPA requirements<sup>54,55</sup> suggest continuous recording of temperature, and DO measurements in at least one tank daily and in every tank at least once a week.

### Disease

65. All independent sources promoting standard bioassays are agreed that disease and the percentage mortality in stock organisms should be kept to a minimum. ASTM<sup>48</sup> recommends that mortality actually be absent, or no greater than 5 percent during the 96 hours preceding the bioassay. The EPA fathead minnow<sup>55</sup> and brook trout<sup>54</sup> recommendations are that in case of an outbreak of a disease in the tanks, all the tanks should be treated for the disease at the same time and with the same level of treatment. Standard Methods<sup>48</sup> recommends not more than 10 percent mortality for 4 days before a test and no shipment of organisms should be accepted that show symptoms of disease, abnormal behavior, etc. The Bureau of Sport Fisheries and Wildlife<sup>51</sup> also recommends that when the organisms show a mortality greater than 10 percent prior to testing, none be used in the test.

## Bioassay Equipment

### Development of bioassay equipment

66. Aquatic bioassay equipment was first used in England during the 1860's. During the first decade of the twentieth century, English toxicology equipment was modified in many ways by American biologists.

67. Unfortunately, the influence of bioassay apparatus on the results obtained in toxicology tests was not critically evaluated until the second decade of the twentieth century. It was at this time that Belding<sup>56</sup> published his review of bioassay technology. The effect of such equipment characteristics as test container volume, construction material leachability, test water temperature, and test water oxygen content were all discussed in this early publication.

68. Before Belding's<sup>56</sup> research was published, there had been little interest in maintaining aquatic organisms in the laboratory for long periods of time. Researchers, for the most part, used nonquantitative field-sampling equipment to describe the general nature of polluted environments. Those investigators who did subject aquatic organisms to laboratory toxicological investigations were unable to determine chronic effects because no provisions were incorporated into the study for reducing the buildup of metabolic waste products. The test organisms themselves add toxic metabolic wastes to the control, test, and holding tanks by their normal physiological activities. Inadequate holding-tank filtration or the use of small holding-tank volumes allowed the buildup of the test organisms own harmful metabolites. The accumulation of nitrogenous wastes, especially ammonia, has been particularly troublesome.

69. Other shortcomings of early toxicological equipment have been determined since Belding's<sup>56</sup> study. The methods of maintaining constant toxicant concentrations for extended periods of time were not widely known during the early phase of toxicological studies. Early investigators were not equipped to maintain constant dosage levels of toxicants. The loss of test material in these early studies has been attributed to adsorption on the walls of the test chambers. Loss of toxicants also occurred through volatilization from open test tanks and natural chemical decomposition.

70. The maintenance of similar water conditions in the test chambers for the duration of an experiment was difficult for early investigators. Differences in water chemistry often occurred. The biological and chemical oxygen demand of a toxicant selectively reduced the dissolved oxygen levels at different exposure concentrations. Changes in pH and water hardness also were experienced through chemical reactions of the toxicant with the exposure dilution water. Variances in water chemistry during an experimental study may have caused altered chemical toxicity.

#### Modern dosing equipment

71. Static bioassay. The configuration and size of the equipment used in static or standing water bioassays has generally not changed

since the techniques were first developed. More information currently is known about the working volumes to be used with certain organisms and the types of materials recommended for construction of the holding tanks. Many tank construction materials have shown promise but have later been shown to have disagreeable effects on bioassay experiments. Fiberglass, poethylene, and polypropylene may absorb some organic toxicants causing an underestimation of the chemicals' toxic properties. Plastics and synthetic rubber products tend to release organics and plasticizers which may produce undesirable side effects to the test organism or toxicant. The material most often recommended for static bioassay tanks is glass. The tanks may be solid glass jars or aquarium tanks either purchased commercially or constructed to specification from double-strength plate glass cemented with a silicone rubber adhesive.

72. A variety of equipment has been used to maintain suspensions of sediment material in static bioassay experiments. Hubble and Reiff<sup>57</sup> successfully maintained a uniform suspension of particulate matter in a 30-gallon reservoir by an electric motor which raised and lowered perforated plates within the container. They were able to maintain 15 mg/l of suspended matter within 15 percent of its mean concentration for up to 7 days. Davis and Hidu<sup>58</sup> used a heavy-duty stirrer to keep sediment in suspension while a cup wheel delivered the suspended sediment to test aquaria. A more elaborate mechanism for maintaining suspending sediments was used by Davis.<sup>59</sup> The device was composed of polyethylene bottles placed on a revolving wheel. During toxicity experiments the wheel was continually turned to constantly agitate all the sediment to the same degree.

73. Continuous-flow bioassay. The equipment difficulties discussed thus far can be overcome or greatly lessened by constant replacement of the toxicant and water to which the organisms are exposed. Continuous-flow bioassay equipment has received much attention as to its design and function. While very elaborate designs have been reported in the literature, many do not possess functional reliability under continued operation.

74. All continuous-flow equipment consists of a reservoir of

exposure water which is metered into the test vessel. The toxicant may be contained in the water reservoir or may be metered into the test vessel separate from the diluent water. The toxicant can also be introduced on-line with the diluent water. Elaboration of this basic design provides for multiple dilution ratios and/or chemical modification of the test water.

75. The selection of a continuous-flow bioassay diluter should consider specific equipment features. Construction should be simple, requiring minimal effort and equipment expenditure. Calibration of the diluter to desired exposure concentrations must not be excessively complicated. Once calibrated, it should be capable of continued operation for extended periods of time. Maintenance needs to be minimal with only slight adjustments necessary to recalibrate the exposure concentration. When the diluter fails to operate, it must fail safely (i.e., toxicant is not allowed to flow into the test chambers in the absence of dilution water).

76. The simplest method of operating a continuous-flow bioassay is to decant a preset dilution of toxicant from a constant-head vessel such as a Mariotte bottle. Flow rates are determined by the size of the delivery tube and the height of the Mariotte bottle above the opening of the delivery tube. Variations in flow rates may result from rapid changes in atmospheric pressure or temperature.

77. Coler, et al.<sup>60</sup> varied the flow of toxicant from a constant-head dosing device by changing the angle of the tubing leading from a constant-head container. Their apparatus was, however, somewhat improved over earlier similar attempts in that a needle was placed on the tubing leading from the constant-head device in order to make the flow rate more precise. A more stationary mechanism for effectively changing the height of the toxicant water column was designed by Drammo and Kohlberg.<sup>61</sup> In this apparatus the toxicant flows from the constant-head storage vessel into a manifold which has pipes threaded into its base. To alter the rate of flow, the investigator screws the pipe in or out, changing its effective length.

78. A constant-head device designed by Lichatowick et al.<sup>39</sup> was

found to be capable of dispensing accurate doses of more viscous fluids. The apparatus consisted of a constant-head plexiglass box with a piece of elbow tubing protruding from the side. When a greater flow was desired, the tip of the elbow tubing was bent down to a lower level. Lichatowich and his coworkers found that this device was capable of dosing all but the largest particles (less than one-half inch in diameter) and was fairly accurate.

79. Anderson<sup>62</sup> combined the flows from two constant-head vessels to deliver seawater and chlorinated seawater to the eggs of Pleuronectes platessa.

80. Abrams<sup>63</sup> utilized a hypodermic syringe to dispense toxicants. The plunger of the syringe was driven by the flow of the dilution water. Design modifications by Stark<sup>64</sup> enabled delivery volumes to be maintained at five-percent error over a 24-week period.

81. Multiple toxicant delivery systems most widely used are of two types: serial diluters and proportional diluters. The common feature of both diluter systems is their capability of providing multiple toxicant concentrations to separate exposure chambers. The method of toxicant dilution is implied by their respective names.

82. The diluter described by Burke and Ferguson<sup>65</sup> exemplifies the design and operation of serial diluters. Water and toxicant are mixed in a series of chambers connected by drip lines. Successive dilution of the toxicant occurs in the connecting chambers. Aliquots of water from the chambers are delivered to the test aquaria. The water flow in the system is regulated by adjustment of the drip lines. The system is self-sufficient in operation, requiring no outside regulatory devices. Frequent measurements of the dilution concentrations are recommended to ensure that flow rates have not been reduced due to residue accumulated in the drip lines.

83. Proportional diluters developed by Mount and Brungs<sup>66</sup> operate on a sequential filling and emptying of water chambers. The water chambers are calibrated to contain a measured amount of water. Separate water chambers are provided for toxicant and diluent waters. The water chambers empty by siphon tubes activated by a venturi. Diluent and

toxicant waters are mixed in the siphon tubes and delivered to the respective test tanks. The cyclic action of the diluter is regulated by a solenoid valve connected to the inflow dilution water. The system is subject to electrical power failure and an alternate emergency power source is recommended with this installation. The system is widely used in institutions because of the ease in calibration and low maintenance requirements.

84. On-site bioassay techniques. In recent years there has been a tendency for toxicologists to incorporate on-site bioassay studies in their analysis of environmental toxicity. On-site bioassays allow for the natural variability of the toxicant receiving water, which is not possible in a laboratory study. Laboratory and on-site studies complement each other in providing information on an environmental contaminant.

85. A simplistic form of in situ bioassay consists of maintaining test organisms in a holding box for observations. The holding box can be modified for specific research interests. A design for studying the effect of sediments upon benthic organisms was devised by Heitmuller, Del Wayne and Nimmo.<sup>67</sup> The structure consisted of a screen cage with a horizontal partition midway in the cage and an open bottom. To determine the effect of the sediment, organisms were placed below the partition. As a control, organisms were also placed in the top chamber removed from the sediment. Live cages made from braided nylon bags were used by Falk<sup>68</sup> to monitor the effects of an effluent upon pelagic organisms. In this design, light aeration was supplied to the test and control chambers by a small air pump.

86. Often benthic zoologists or phycologists allow naturally occurring fauna and flora to colonize on artificial substrate material. The number, size, or diversity of the populations which attach to the artificial substrates subjected to toxic chemicals are then compared to similar substrate in water above the effluent discharge.

87. The types of artificial substrates and the materials used to construct them are many and varied. The number, diversity, and per-day colonization rate of benthic macroinvertebrate populations taken by samplers made from 3M "conservation webbing" or from masonite plates



were compared by Dickson and Cairns.<sup>69</sup> Other macroinvertebrate substrates include hardware cloth baskets filled with stone (Wene and Eickliff<sup>70</sup>), hardware cloth and angle-iron cages (Henson<sup>71</sup>), and "Bar-B-Q" baskets filled with limestone (Mason et al.<sup>72</sup>). Kauss et al.<sup>73</sup> floated a polyethylene cylinder in shallow ocean water to test the toxicity of artificially produced oil spills on the various types of planktonic algae growing inside. Such conditions as algae density, temperature, dissolved oxygen, and pH of the water inside the cylinder were found to be very similar to the ocean conditions surrounding the cylinder. An artificial substrate for periphyton used by Dickman<sup>74</sup> consisted of agar-coated slides which could be floated in natural waters. Dickman<sup>74</sup> coated the slides with the bioassay treatment material, in this case germanium, and was thereby able to dispense with elaborate toxicant dosing mechanisms.

88. Another approach to conducting on-site biological monitoring is to actually bring a portable laboratory to the site to make closely controlled tests. The diluent waters are generally taken above the area where the industrial discharge is introduced. The toxicant can, likewise, be taken either directly from the industrial discharge or from a mixing zone within the receiving stream.

89. One of the most elaborate mobile bioassay laboratories was discussed by Zillich<sup>75</sup> in a review of on-site bioassays of combined chlorine residual. This apparatus--operated by the Michigan Water Resources Commission--is capable, through serial dilution, of running at many different dilutions (40 in the case of the Zillich study) with replication. The laboratory is also equipped to monitor the pH, dissolved oxygen, conductivity, and temperature of the effluents and of the polluted stream.

90. Exposure chamber modifications. Many modifications have been made on basic exposure tanks in an effort to more closely approximate the environmental conditions of organisms which have special requirements for survival. Planktonic organisms often fall out of suspension unless culture waters are circulated. Planktonic algae are often maintained in a flask or culture bottle with a magnetic stirrer inserted to

keep the medium in motion.<sup>76,77</sup> Other bioassay techniques for algae have been evaluated by Murray et al.<sup>78</sup>

91. Some animals require flowing water in order for their respiratory system to function properly. Benthic organisms requiring flowing water are generally exposed to toxicants with approximately the same type of equipment used in their culture. One such unit is an oval tank with the water constantly flowing in a circular direction.<sup>79,80</sup> Commercially prepared rectangular tanks which constantly force the water to the far side with a propeller on one end have recently become popular for tests on lotic benthos. The tank has a false floor through which the water can be drawn to be recirculated. The test organisms are generally situated on the top side of the false bottom. Both types of "artificial streams" have screens surrounding the water-impelling devices so that the test organisms will not be damaged.

92. Occasionally investigators have attempted to determine not only the effect of a toxicant on a single population of animals, but also the effect of biological influences such as parasitism and predation which would normally occur in that species' natural environment. One of the more closely controlled of such tests was devised by Goodyear.<sup>81</sup> Mosquitofish (Gambusia affinis) and largemouth bass (Micropterus salmoides) were placed in adjoining aquaria which were separated by a screened partition. The Gambusia could enter the largemouth bass tank where they would be subject to predation, but the largemouth bass were too large to enter the mosquitofish aquarium. Toxicant continually flowed into the bass aquarium and was allowed to pass into the Gambusia tank. Two types of controls were used for the test. One had a toxicant and no predation. The other contained no toxicant, but the predator was present. The relative number of Gambusia left in these variously treated aquaria after several days of exposure was felt to be a quantitative indication not only of the toxicant lethality but also of changes in predation due to the toxicant.

#### Sublethal tests

93. Environmental contaminants do not always elicit lethal effects when organisms are exposed to them. For this reason, researchers have

recently begun to investigate sublethal signs of stress in aquatic organisms. The numerous different types of sublethal parameters now monitored fall into the two major categories: either physiological or behavioral. Generally, larger organisms for which the physiological or behavioral indices are more easily monitored are used for these tests.

94. The amount of carbon dioxide evolved or the oxygen consumed by aquatic organisms has been monitored in much the same way by several toxicologists. An aquatic organism is placed in a sealed life-support system, the dissolved oxygen and carbon dioxide content of which is known. After a period of time, the concentrations of these two respiratory gases in the test water are again measured, and their depletions or increases are attributed to the metabolic functions of the experimental subject. Lee and Buzzell<sup>82</sup> measured the respiration rate of fish in a closed continuously flowing system with this technique. In an effort to measure the effect of various oxygen tensions on small benthic invertebrates, Gamble<sup>83</sup> bound the organisms with a hair inside a closed water trough. Securing the organisms in this manner obviated the need for other types of attachment substrate.

95. Monitoring of excretion ratio can be conducted with essentially the same support system as was used to measure respiratory gases. Lloyd and Orr,<sup>84</sup> for example, ran catheters from trout to an electrical contact. When the fish urinated, the electrical contact would close and a permanent recording device would be activated.

96. Equipment allowing the elicitation and measurement of behavioral responses has come into common usage only in recent years. Perhaps the most studied behavioral response is avoidance. Most avoidance tests have utilized large troughs which have inflow pipes for clean water and one or more discharge pipes. The amount of time that an aquatic organism spends near the inflow pipe carrying the toxicant is compared to the time that the organism spends in clean water areas of the trough. Most such troughs are large so as to allow the formation of distinct separation of clean and treated water. A much smaller chamber was designed by Jones<sup>85</sup> and has been used recently by Cook and Boyd<sup>86</sup> to study the avoidance reactions of the sand shrimp. The Jones<sup>85</sup>

avoidance apparatus consists of a glass cylinder secured in a ring clamp in the horizontal position. Toxicant enters one end of the cylinder at the same rate that clean water enters the other end. Both fluids continually drain through a hole in the center of the cylinder. Scherer and Nowak<sup>87</sup> described an electronic apparatus used to continuously record the movement of organisms during avoidance tests. The permanent chart recording of the movements of the test organism can be used to assess movement in avoiding the test material.

### PART III: TEST ORGANISMS

#### Criteria for Selection of Test Organisms

97. "The major requirements for conducting bioassays are an adequate supply of test fish and dilution water, and laboratory facilities for holding fish and conducting the bioassays," (Henderson and Tarzwell<sup>8</sup>). Though some authors recommend fish for use in bioassays, they also agree that the forage species such as invertebrates which support these fishes are of sufficient importance to warrant selection in bioassay research.<sup>89-91</sup> Whether predatory or forage species are chosen for testing, the criteria cited by the authors for selection of individual species or phylogenetic groups are often numerous; differ from author to author; and vary with the type of test, test conditions, facilities available, and test objectives. The criteria cited by most authors for selection of test species are listed below in no particular order of importance.

- a. Type of test.
- b. Economic importance.
- c. Ecological significance.
- d. Geographic distribution.
- e. Ease of capturing, handling, holding, and culturing.
- f. Availability and local abundance.
- g. Sensitivity to the toxicant.
- h. Consistency of response to toxicant.
- i. Reproductive success under assay conditions.

#### Type of test

98. The first criterion to be considered in the selection of a test species involves the consideration of the type of test to be performed (standardization, regional, or localized assessment, or behavioral response) and the substance being tested (turbidity, heavy metals, petroleum, pesticides). For example, in tests where the objective is to determine the lethality of a particular compound for establishing standardized concentration limits, only the very common laboratory species such as goldfish, guppy, or fathead minnow should be chosen. Mount<sup>90</sup>

recommended the goldfish as a possible aquatic equivalent to the laboratory white rat; although Katz<sup>91</sup> considered it "not highly regarded because of its tolerance." The EPA<sup>92</sup> in Duluth, Minnesota, has used the fathead minnow as a "white rat" in chronic tests and has developed a fathead minnow stock culture.

99. In tests that involve local toxicity or discharge problems from industrial effluents, pulp mills, dredging operations, or pesticide runoffs, those species which naturally occur in the body of water receiving the contaminant are often preferred. For example, bioassays involving contaminants discharged into cold northern streams should be conducted with cold-water organisms such as rainbow trout, walleye, northern pike, and cold-water invertebrates. Tests involving contaminants in temperate areas such as Chesapeake Bay should use endemic organisms exemplified by the striped bass, softshell clam, and blue crab. Becker et al.<sup>93</sup> recommended many suitable marine organisms according to eight geographic districts. Test data for laboratory species of goldfish and guppies may be of less value in a local toxicity problem than data collected with local or endemic species.

100. All species vary in their susceptibility to toxicants. Some are resistant to certain chemicals. For example, the quahog (Mercenaria mercenaria) is resistant to methoxychlor and malathion.<sup>94</sup> By careful examination of available literature, selection of resistant species can be avoided.

101. In short-term bioassays involving silt or turbidity, those species of snails, barnacles, and tubeworms which can seal themselves off from the surrounding water by means of an operculum or shell should be avoided. Numerous fishes and invertebrates appear to be suitable. The valuable oyster Crassostrea virginica was found to be very sensitive to turbidity and silt deposition and, therefore, may be suitable for use in turbidity bioassays. Certain other species, particularly those belonging to the families Libellulidae (skimmers), Chironomidae (midges), and the genera Simulium (black flies) and Chaoborus (phantom midge flies) tolerate high turbidity levels and would be less desirable selections for turbidity bioassays.

102. In avoidance tests, motile species which display an avoidance reaction should be selected. Many species of fishes have been used in avoidance tests, particularly species which are migratory and thus may be repelled from entering their natal streams by obnoxious or toxic substances.<sup>91</sup>

#### Economic importance

103. Many species of plants and animals are harvested in commercial or recreational activities due, in part, to their abundance, palatability, and availability. These harvests are of vast economic importance and need to be protected. Those species which satisfy this criterion of economic importance are often used in toxicity bioassays.<sup>89-91</sup>

104. The economically important freshwater species are almost entirely fishes, especially the Salmonids, Centrarchids, Cyprinids, and Catostomids. Various species of freshwater mussels and crayfish are of local economic importance in certain areas. In estuarine and marine waters, the spot, mullet, pinfish, striped bass, several salmonids, several decapod crustaceans, oysters, clams, and mussels are of economic value and have been used repeatedly in bioassays.<sup>93</sup>

#### Ecological significance

105. Those species which are of economic importance also play a major role in the ecosystems to which they belong due to their position as consumers and predators in food chains. Though most authors have only considered economically important animals, particularly fishes, for use in bioassays, some authors have recognized the significance of forage or food organisms in aquatic food webs and have selected them as test organisms in toxicity bioassays, since those primary food organisms which sustain the harvested crop must be protected.<sup>90</sup>

106. Diatoms have been selected as bioassay test organisms in marine and freshwater studies<sup>95-97</sup> because of their significance as food organisms and indicator species. The Florida Mangrove (Rhizophora mangle) is an important food source and nursery ground for many commercial fisheries, and has been used in bioassays.<sup>98</sup>

107. The saltwater shrimps Penaeus aztecus, P. duorarum, and Palaemonetes pugio are abundant and ecologically important along the

Gulf coast and have been used in bioassays.<sup>99,100</sup> Six species of freshwater crustaceans were selected for bioassays with herbicides in Missouri where they were considered to be important links in the food chains of fishes.<sup>101</sup> The ostracods are an important link in aquatic food chains, as they are widely distributed, omnivorous, scavenging crustaceans, and have been used in bioassays.<sup>102,103</sup> They occupy benthic and littoral niches in ponds, lakes, and slow moving streams where they filter-feed on bacteria, protozoans, algae and detritus.

108. Dragonfly nymphs have been subjected to DDT residue accumulation studies.<sup>104</sup> Many of the ecological and morphological characteristics of the insect suggest that they may be a major source of biological magnification in food chains. Dragonfly nymphs are described as a primary food source of carnivorous fishes and voracious predators of small insects. They pass a large amount of water over their gills, putting them in contact with pesticides in solution.

#### Geographic distribution

109. Data for organisms with widespread distribution are generally more valuable than those on species having restricted distributions. Most of the species recommended for use in bioassays by Doudoroff et al.,<sup>89</sup> Katz,<sup>91</sup> Henderson and Tarzwell,<sup>88</sup> Becker et al.,<sup>93</sup> Mount,<sup>90</sup> and Battelle-Columbus Laboratories<sup>105</sup> are widespread in distribution. Goldfish, bluegill, largemouth bass, fathead minnow, and rainbow trout have wide distribution throughout North America. They have also been selected often for use in bioassays.

#### Ease of capturing, handling, holding, and culturing

110. Keeping the test organisms healthy, unstressed, and disease-free is essential in conducting a bioassay. Those species which do not lend themselves to the procedures involved in capturing, handling, holding, and culturing should be avoided. Delicate fishes, and even some species which are quite hardy in nature, do not do well in captivity.<sup>89</sup>

111. The response of test organisms to toxicants varies with the relative well-being of the organism. Parasitism, injury, and other health problems may accentuate or mask the response of the organism to



the toxicant being studied, resulting in inconsistent or misleading data. Those chosen must also be capable of acclimating to a holding tank or aquarium and, if a chronic test is anticipated, must be culturable under controlled laboratory conditions.

112. Procedures for capturing and handling test fish for bioassays are summarized in several papers.<sup>50,88,89,91,106</sup> The authors of individual bioassay studies often include discussions of these methods in their text.

#### Availability and local abundance

113. Availability of an adequate supply of healthy fish of desirable uniform size may be the major factor in the selection of a test species.<sup>88</sup> Test species can be acquired from two major sources: collected from their natural habitat (wild-caught) or obtained from commercial bait shops, commercial hatcheries, government hatcheries, laboratory stock populations, or inbred, controlled pond populations.

114. Those species which are wild-caught must be available in large numbers year-round to ensure that an adequate number of healthy, uniformly sized individuals representing a natural male:female ratio are available from the population. There are species, readily obtainable in some seasons, that are impossible to obtain in others, or that may be available only in one stage of their life cycle. The seasonality of availability can be eliminated by maintenance of stock laboratory or pond populations.

#### Sensitivity to toxicants

115. The sensitivity of individual species to different chemicals varies greatly. Fishes that are tolerant of low oxygen conditions are not necessarily tolerant of toxic materials.<sup>88</sup> Some plants and animals are highly resistant to certain toxicants, while others may be extremely sensitive to the same toxicants. The sensitivity of some organisms changes with maturation from embryo to larva to adult and with variations in the physical-chemical conditions of the test water.

116. Those species which have median sensitivities to pollutants are most desirable for use as bioassay test organisms. Test species should not be resistant to the chemicals or conditions being tested.

Conversely, they should not be so sensitive that they die immediately upon contact with a very small amount of toxicant thus precluding a range of response being obtained.

117. The quahog (M. mercenaria) is resistant to methoxychlor and malathion<sup>94</sup> making it unsuitable for testing with these pesticides. The marine green alga Dunaliella tertiolecta and dinoflagellate Exuviella sp. are extremely resistant to lead and chromium.<sup>107</sup> Dunaliella tertiolecta was also most resistant to copper among nine species of phytoplankton,<sup>108</sup> and to urea herbicides among six species.<sup>109</sup> The marine clam Artica islandica is extremely resistant to hydrogen sulfide-rich solid waste eluates.<sup>110</sup> The periwinkle Littorina littorea was too resistant to oil to be used in bioassays; while the intertidal mollusc Patella vulgata was too sensitive.<sup>111</sup> Artemia salina larvae were resistant to copper and mercury in tests conducted in the United Kingdom.<sup>112</sup> The embryos of the sea urchin Strongylocentrotus purpuratus were selected for tests with oil products because of their noted sensitivity to pollutants but were found to be too sensitive to determine gradations of toxicity.<sup>113</sup>

#### Consistency of response to toxicity

118. Suitable test organisms should display consistent responses to toxic conditions to ensure uniformity in the data obtained. Those organisms which, due to morphologic or behavioral characteristics, respond variably among individual specimens should be avoided.

119. The degree of response consistency sometimes changes with variations in water conditions, age, sex, and life stage. The changes incurred by varying water conditions can be controlled and avoided; while those changes caused by biologic factors may not always be controllable. Sensitivity may also vary according to the environmental background of the specimens, i.e., the source from which the test organisms were acquired.

120. The greatest variations in toxicity response seem to occur among the arthropods. The arthropods not only differ in their responses with age and life stage (egg, larva, adult), but also differ during intermolt and molting stages. For example, several marine crustaceans

in Norway were found to be more resistant to surfactants during the in-termolt stages.<sup>114</sup> The lobster Homarus americanus was shown to be most susceptible to crude oil during molting.<sup>115</sup>

121. Among inbred strains of platyfish (Xiphophorus maculatus), the males and females show differences in response to zinc.<sup>116</sup> Among the marine annelids, Capitella capitata displays a differential sex susceptibility to detergents.<sup>117</sup>

122. Fish sometimes demonstrate inconsistent differences in responses with age. Endrin is more toxic to adult Gasterosteus aculeatus (threespine stickleback) than to eggs and larvae, and to adult rainbow trout, chinook salmon, coho salmon, bluegill, mosquitofish, and guppies than to their respective embryos.<sup>118</sup> The fry of bluegills, green sunfish, smallmouth bass, lake chubsucker, and stoneroller, are more sensitive to several herbicides than their fertilized eggs.<sup>119</sup> Large specimens of the smallmouth bass, largemouth bass, bluegill, and black crappie are less sensitive to DDT than small specimens.<sup>120</sup> Older chinook salmon are less sensitive to DDT than the young.<sup>121</sup>

#### Reproductive success under assay conditions

123. In chronic bioassays or those involving most of the species' life cycle, the ability of that organism to reproduce successfully in the laboratory is of importance. Suitable species are those that can be induced to spawn under laboratory conditions and demonstrate a relatively high survival of the fertilized embryos. It also is desirable that the larvae and immature stages develop to adults in satisfactory numbers.

124. The majority of the species selected by bioassay investigators are those which have previously been maintained under laboratory conditions following published cultural techniques of handling, maintenance, and care. Some species have been used more frequently than others; e.g., the bluegill (Lepomis macrochirus) appears in more published bioassay literature than does the orangespotted sunfish (Lepomis humilis). This does not necessarily imply that the bluegill is a better test species than the orangespotted sunfish. The dichotomy does serve to illustrate the widespread geographical distribution of the bluegill as

opposed to that of the orangespotted sunfish. Very few reports surveyed listed species which proved to be unsuitable for bioassays.

#### Species Most Commonly Selected

125. Table 1 lists the most commonly selected species from the reports surveyed which include 11 species of plants, 6 species of molluscs, 1 annelid, 13 crustaceans, 5 aquatic insects, and 31 fishes. All the species listed are of economic value or ecological significance, and most meet all nine selection criteria discussed above. The salmonid and centrarchid fishes comprise some of the most important game and commercial fishes in North America. The freshwater fishes as a group have been subjected to more bioassay tests with a greater variety of toxicants than any other group of organisms. All the decapod crustaceans (shrimps, crabs, lobsters, crayfish) and bivalve molluscs (clams, oysters, mussels) support multimillion dollar recreational and commercial industries in North America and abroad. The insects, annelids, and plants of Table 1 comprise important links in food chains since they are direct or indirect food sources for the economically important fishes, molluscs, and crustaceans.

126. A large number of marine and freshwater plants have been subjected to bioassays, but only 11 common species of planktonic diatoms and green algae have been used frequently (Table 1). The green alga Chlorella pyrenoidosa appeared most frequently in freshwater bioassays. The frequent appearance of the planktonic diatoms and green algae in bioassay experiments exemplifies their application as bioassay organisms. The small size, rapid reproductive rate, and extensive cultural techniques facilitate their use as test species. Patrick<sup>122</sup> discussed in detail the procedures for bioassays using diatoms.

127. Several species of marine molluscs have been selected often for use in bioassays. The mussel (Mytilus edulis), oyster (Crassostrea virginica and C. gigas), the quahog (Mercenaria mercenaria), and the soft-shell clam (Mya arenaria) have been used either as adults or as planktonic larvae (Table 1). Fewer freshwater molluscs have been used in bioassays. The snail Physa heterostropha was the most frequently

used freshwater mollusc encountered in the present survey.

128. A relatively small number of annelids have been reported for use in bioassays. The tubificid Limnodrilus hoffmeisteri is an important member of pond and stream benthic communities serving as a food source for fish populations. The frequency with which this organism appears in the literature indicates its general acceptance as a suitable test species. Several species of marine sandworm Nereis have been used in foreign and domestic research.

129. Among the freshwater fishes, five species, the bluegill (Lepomis macrochirus), rainbow trout (Salmo gairdneri), fathead minnow (Pimephales promelas), goldfish (Carassius auratus), and largemouth bass (Micropterus salmoides), appear to have been selected most often. These species are available throughout most of North America from bait shops, fish culturists, or hatcheries or are accessible from ponds, lakes, and streams. The above five species and many of the other species listed in Table 1 belong to the fish families Centrarchidae, Salmonidae, Cyprinidae, and Catostomidae. They were recommended for bioassay use by Doudoroff et al.<sup>89</sup> Katz<sup>91</sup> mentioned the bluegill, guppy, rainbow trout, fathead minnow, and goldfish as frequently used or standard fish. The above five species were among the 20 recommended for use in bioassay by Mount.<sup>90</sup>

130. The sheepshead minnow (Cyprinodon variegatus) and the mummichog (Fundulus heteroclitus), both common cyprinodonts, are among the most often selected marine/estuarine fishes. These two species are abundant in tidal ditches and estuarine marshes, and are easily maintained under laboratory conditions.

#### Species Infrequently Used

131. Table 2 illustrates the large number of species of plants and animals which have been used infrequently in bioassays. All species listed were either assayed in North America or are naturally available there. They appear to be suitable for use in bioassays. Most do not completely meet all nine selection criteria previously given. Some

have restricted geographic distribution or physiological and behavioral characteristics which present problems in assay conditions, and thus are slightly less desirable as test species. However, there are several notable exceptions to this rule which will be discussed below.

132. Numerous bacteria, yeasts, and marine fungi have been used in bioassays. There is a great amount of data in published reports and tests on isolating and culturing these organisms by microbiological techniques. However, most of this information deals with research efforts relative to disease transmission and human hygiene. Only 11 species were chosen as test organisms in the reports surveyed relative to dredging operations. The enteric bacterium Escherichia coli, is probably the most commonly selected form due to its reputation and importance as an indicator of water quality.

133. At least 70 species of freshwater and marine algae and 17 species of flowering plants have been used infrequently. As illustrated in Table 2, the vast majority of these species are planktonic, single-celled forms and few are macroscopic, benthic forms. Many of these planktonic organisms listed in the reports surveyed were not identified by species, indicating that the investigators experienced taxonomic problems with this group. That only a small number of benthic, macroscopic algae have been used is indicative of the problems inherent in laboratory culture of these plants. However, marine algae have been successfully cultured in several coastal laboratories and in the United Kingdom, and methods for culturing freshwater stream algae have been developed in several North American laboratories.

134. Corals, which are of great ecologic, geologic, and economic importance in tropical waters but which are relatively difficult to culture under laboratory conditions, did not appear as test organisms in the reports surveyed. Since these animals are very sensitive to siltation, they may be useful indicators of turbidity and siltation in the laboratory, provided methods for culturing them can be developed. Three Caribbean corals were maintained in the laboratory at the University of Georgia for a short time during feeding and respiration tests.<sup>123</sup>

135. Only two of the reports surveyed<sup>124,125</sup> include use of

platyhelminths in North America, both of which involved turbellarians. The turbellarian Polycelis nigra was selected for use in a bioassay with several toxicants in Wales.<sup>126</sup> Turbellarians such as the common Dugesia sp. are easily maintained and cultured under laboratory conditions.

136. The coot clam (Mulinia lateralis), which occurs in the coastal waters of the U. S., has proven to be an excellent laboratory animal and breeds successfully under laboratory conditions. Though this species has not yet been used extensively, its potential as a standard organism for bioassay and genetic research has been explored by the U. S. National Marine Fisheries Service in Connecticut.<sup>127</sup> It has been used successfully in temperature-tolerance tests in Chesapeake Bay.<sup>128,129</sup>

137. Many other species of molluscs not listed in either Tables 1 or 2 could be used in bioassays. For example, the freshwater mussels (Family Unionidae) selected in one study,<sup>124</sup> the marine shipworms (Family Terebinthidae), and the marine scallops (Family Pectinidae) used in another<sup>130</sup> have been proven usable under laboratory conditions.

138. In addition to the tubificid and the nereid listed in Table 1, 16 other species of annelids were selected infrequently by domestic investigators. These included a leech, two tubificids, and many errant and sedentary polychaetes. Both larvae and adults have been used. Most of the annelids selected were polychaetes, including several sandworms (Nereis sp.) and tubeworms (serpulids) which occur in estuarine and marine waters of the U. S. The polychaete Capitella capitata, used in a California study,<sup>131</sup> has been cultured through succeeding generations and used in bioassays in France.<sup>117</sup>

139. Various species of acorn barnacles have been cultured in North America, the United Kingdom, and Scandinavia. A great deal of information on the biology and ecology of barnacles is available, due to their economic importance as fouling pests. In North America Balanus balanoides has been bioassayed often (Table 1) while B. eburneus, B. cariosus, and B. improvisus have been used less frequently (Table 2). Both the larvae and adults of these species have been assayed. The

larvae are easily cultured and are often more sensitive to toxicants than the adults.

140. The most economically important group of crustaceans is the decapods. In addition to the shrimps, lobster, and crab listed in Table 1, there are many freshwater crayfish, marine crabs, fiddler crabs, hermit crabs, and shrimps which are suitable as bioassay organisms. These animals have been used in both the larval and adult forms and generally are adaptable to laboratory conditions. All life cycle stages of Cancer magister have been assayed with pesticides at Oregon State University.<sup>132</sup> Carcinus maenas, which occurs in New England, has presented feeding problems to investigators in England.<sup>133</sup>

#### Test Species Previously Used in Turbidity Bioassays

141. Various species representing the bacteria, algae, molluscs, annelids, crustaceans, and fishes have been used in bioassays involving many turbidity-producing substances. A list of species and the type of substance they were exposed to is presented in Table 3. Test organisms were subjected to sludge, stream sediment eluates, paper and pulp mill effluents, silt, kaolin, paper and wood fibers, harbor sediments, ferric hydroxide flakes, and other substances.

142. The salmonids (e.g., Oncorhynchus kisutch, Salmo gairdneri), walleye (Stizostedion y. vitreum), minnows (Pimephales promelas, Notropis spp.), oysters (Crassostrea gigas, C. virginica), and quahogs (Mercenaria mercenaria) were selected most often. The larvae, embryos, juveniles, and adults of these species have been subjected to turbidity bioassays in one or more studies. The majority of the species listed in Table 3 are either fishes or benthic molluscs. Among the fishes selected, the salmonids and walleye were often used to illustrate avoidance among important, mobile animals. Also, these species naturally occur in geographic areas where pulp mill effluents and other turbid materials are commonly discharged into streams, thereby making them ideal choices as test organisms. Among the molluscs selected, the oysters, quahogs, and mussels are particularly sensitive to siltation due to the



nature of their filter-feeding apparatus. Since these animals occur naturally in bays, estuaries, and nearshore coastal areas subject to dredging and disposal operations; are sensitive to siltation; and are known to accumulate and concentrate heavy metals, they also make ideal test organisms.

#### Test Species Which Present Special Problems

143. Among the reports surveyed, numerous authors noted some of the difficulties or drawbacks in using certain species as test organisms (Table 4). Among these difficulties are problems which resulted in high mortalities, cannibalism, territoriality and aggression, susceptibility to parasitism, feeding problems, handling problems, and other special physiological and behavioral difficulties. Some of these species are also listed in Table 1, as "suitable (commonly used)," indicating that though these organisms have been selected often by a variety of researchers, they occasionally require special procedures to successfully implement a laboratory bioassay.

144. Large kelps such as Laminaria saccharina, which can attain a total length of 1 meter or more, can grow to be too large for use in laboratory conditions.<sup>134</sup>

145. Most benthic bivalve molluscs are filter-feeders and normally pump a large amount of water through their filtering apparatus to procure suspended food particles. In controlled laboratory conditions, a large amount of artificial or natural seawater supplied with suspended food must be provided to keep these animals alive for extended period of time. Many authors discussed mortality and feeding difficulties with Mytilus edulis.

146. The sensitivity of arthropods often changes from one life cycle stage to another. Most species are very sensitive during molting or ecdysis, prompting some authors to advise running bioassays on arthropods for all phases of life cycle or the most sensitive stages.<sup>135</sup> The major problem with some crustaceans is cannibalism, which makes necessary the providing of individual holding containers for each animal.

The time or point of death is often difficult to accurately determine among the small crustaceans, such as the amphipods and ostracods.

147. Among the problems encountered with the fishes, all three species of Alosa were susceptible to mortality during handling. Several fish species were susceptible to impaired health due to parasitism, including the goldfish, bluegill, mosquitofish, largemouth bass, and brook trout. Some authors experienced unexplained mortalities among test and control fishes. Some species tended to jump out of their test containers.

#### Test Species Unsuitable for Bioassays

148. Among the reports surveyed, only three species were considered by the authors to be unsatisfactory for use in bioassays. In a study with silt in South Dakota, tests with the drum (Aplodinotus grunniens) and the sauger (Stizostedion canadense) were unsuccessful because they did not adapt to aquarium conditions. The white crappie (Pomoxis annularis) was referred to as unsuitable under laboratory conditions.<sup>136</sup> However, since the crappie had been used successfully in pond studies in Illinois<sup>137</sup> and short-term aquarium studies in Iowa,<sup>138</sup> it must be considered as, perhaps, marginally suitable.

149. Studies with the marine lichen (Lichina pygmaea) in the United Kingdom<sup>139</sup> showed that such organisms, as a group, are not suitable for laboratory research, unless freshly collected specimens are used quickly for short-term studies.

150. The freshwater water strider (Gerris remigis) was found to be unsatisfactory in a study involving mirex in Mississippi.<sup>140</sup> Experimentation with this insect was terminated due to high mortalities in the control group.

#### Recommendations for Capturing, Handling and Maintaining Wild-Caught Organisms

151. The recommended methods for capturing, handling, transporting, holding, and maintaining numerous species of plants and animals for

use in bioassays are summarized and discussed in many published papers and texts. For example, the use of various capturing and holding techniques is included in Standard Methods<sup>50</sup> and discussed by several other authors.<sup>88,89,141,142,143</sup> Capturing and culturing techniques are also available for various freshwater and marine invertebrates, fish and algae.<sup>144,145</sup> The methods recommended by EPA for collecting bioassay test specimens and organisms for other laboratory analyses are described by Weber.<sup>146</sup>

152. Several physiological changes can result from handling stress<sup>147</sup> and subsequently influence the results of bioassay analyses. Among the reports surveyed, many authors mentioned problems with fishes, insects, crustaceans, and planktonic larvae being sensitive to handling, sometimes requiring termination of the tests.

153. Most of the reports surveyed which described research involving wild-caught organisms included a brief discussion of capture techniques and the equipment and methods used in holding and maintaining the organisms. The reports dealing with each species selected in the research surveyed are listed in Tables 1, 2, and 3. It is recommended that the appropriate reports be examined prior to a bioassay project to determine the acceptable and successful methods to be used with each species.

#### Recommendations for Culture of Test Organisms and Stock Populations

154. Methods and apparatus for culturing some species of edible plants and animals have been available for hundreds of years, while the techniques necessary to culture others have been developed only recently. When the purpose of culturing is to obtain organisms for use in bioassays, cultured stock populations provide the advantage of ensuring that the test organisms are free of pollutant contamination and are descendants of the same genetic stock. However, it should be cautioned that laboratory cultured organisms may have a different sensitivity than wild-caught or endemic species.

155. The majority of the freshwater and marine phytoplankton

utilized in bioassays were obtained from laboratory stock cultures. Among the reported bioassays involving use of freshwater species, many authors<sup>148,149</sup> obtained specimens from the Indiana University stock culture populations. Kauss et al.<sup>151</sup> used Bold's<sup>150</sup> basal medium in maintenance of their stock cultures. Most cultures were axenic and unialgal, though some investigators reported problems with bacterial contamination.

156. Bold<sup>150</sup> described in detail the procedures and apparatus for cultivation of algae and discussed special methods suitable for each algal group. The cultivation of marine and freshwater phytoplankton populations, particularly as food sources of cultured animals, has been described.<sup>145,152-158</sup> The green alga Selenastrum capricornutum which, along with Anabena flos-aquae and Microcystis aeruginosa, is recommended for use in bioassays, has been laboratory cultured according to the methods of the "Provisional Algal Assay Procedure" (PAAP).<sup>159</sup>

157. The methods of cultivating river periphyton communities have been developed and described.<sup>160,161</sup> The apparatus for culturing thermophilic stream algae under the conditions of natural stream flow have been developed by Sperling and Grunewald.<sup>162</sup>

158. Methods used in culturing and maintaining continuous supplies of marine diatom, dinoflagellate, and green phytoplankters for use in bioassays and similar experiments are available.<sup>156,163-165</sup>

159. Several macroscopic benthic algae have been laboratory-cultured for bioassays, mostly in the United Kingdom. Methods and apparatus for rearing marine benthic algae were summarized and described by Kornmann.<sup>166</sup> The red alga Callithamnion hookeri, a common British intertidal species, requires no medium enrichment, is easily cultured in small flasks, is morphologically simple, and grows rapidly.<sup>167</sup> Zoospores of Laminaria saccharina, a brown alga, were acquired from laboratory-reared adult plants and successfully tested with sewage, silt, and other pollutants.<sup>168</sup> This species and other members of the genus Laminaria are abundant subtidally in North America and are easily cultured, but present problems if they become too large for available facilities.<sup>134</sup>

### Coelenterates

160. Many hydrozoans have been maintained in laboratories on a diet of brine shrimp (Artemia salina).<sup>169</sup> Among the species utilized, some have been wild-caught, while others have been cultured in the laboratory. The highly sensitive marine hydroid Eirene varidule was cultivated easily in Germany and proved to be suitable for marine bioassays.<sup>170</sup> The "sea nettle" (Chrysaora quinquecirrha) of Chesapeake Bay has been laboratory-cultured through the strobilae, ephyrae, and adult stages.<sup>171</sup> Adult medusae of Aurelia aurita developed from ephyrae within four months under controlled conditions in a chemically defined salt solution.<sup>172</sup> The methods used for raising Phialidium gregarium through its complete life cycle have been described.<sup>173</sup>

### Bryozoans

161. Although few bryozoans were used in the bioassay reports surveyed, these animals are amenable to handling and culturing. Jebram<sup>174</sup> discussed the use of glass slides and coverslips in the cultivation of colonies of marine and estuarine species. The estuarine species Conopeum tenuissimum was also collected as larvae on glass slides and reared in a laboratory seawater system.<sup>175</sup> Freshwater ectoprocts were grown in aquaria in uncovered inverted petri dishes where they fed on natural suspended organic matter present in the water.<sup>176</sup> Several species of bryozoans have been raised on diets of phytoplankton.<sup>169</sup>

### Rotifers

162. No rotifers were used as bioassay organisms in the reports surveyed. However, rotifers can be cultured in the laboratory. Species of Asplanchna were raised on Paramecium, which were fed cultures of bacteria.<sup>177</sup> Gilbert<sup>178</sup> described methods for initiating and maintaining cultures of Brachionus calyciflorus.

### Annelids

163. A variety of oligochaetes and polychaetes have been used as bioassay organisms (Tables 1 and 2). Some of these species were acquired from laboratory stock cultures. The polychaete Capitella capitata was cultured through its entire life cycle, feeding upon dried Ulva every 5 days.<sup>117</sup> This species and other polychaetes have also been

raised on dried, resoaked Enteromorpha.<sup>179</sup> Hauenschild<sup>169</sup> lists many marine polychaetes as being laboratory-cultured. The relative rates of growth of annelid larvae fed various prepared and natural food materials were determined by Howie.<sup>180</sup> None of the authors who performed bioassays involving freshwater tubificids used specimens obtained from culture stocks.

#### Molluscs

164. Loosanoff and Davis<sup>181</sup> reviewed and described the recommended apparatus, conditions, and procedures for rearing eggs, larvae, and adults of many marine bivalve molluscs. Among the species they listed as having been successfully reared from fertilization to metamorphosis, the mussels (Mytilus edulis and Modiolus demissus), scallop (Aequipecten irradians), oysters (Ostrea edulis, Crassostrea virginica, and C. gigas), quahog, (Mercenaria mercenaria), surf clam (Spisula solidissima), and soft shell clam (Mya arenaria) have been used in bioassays in foreign and domestic laboratories. The rearing of marine bivalve larvae requires a large supply of unpolluted water, use of nontoxic pipes, filtration of the seawater, the addition of phytoplankton cultures as food, and apparatus to regulate water temperature. Bivalves can be conditioned to spawn out-of-season by manipulation of water temperature. Most cultured larvae were raised in natural seawater, though some authors<sup>182,183</sup> reported success with synthetic seawater. Most authors reported feeding the larvae protozoans and/or single-cell phytoplankton.

165. Woelke<sup>184</sup> proposed standard receiving water tests and criterion based on the use of 48-hour old Pacific oyster (Crassostrea gigas) embryos and described the equipment and facilities necessary for conducting such bioassays. Other reports which discuss rearing, supplemental feeding, bacterial flora, seawater sterilization, and production of clutch-free spat relative to laboratory analyses of oyster larvae are available.<sup>185-187</sup> The effects of microorganisms, organic substances, and algae cultured on treated sewage effluent on the feeding activities and growth rates of adult oysters have been described.<sup>188-190</sup>

166. The small coot clam, Mulinia lateralis, which was used occasionally in bioassays (see Table 2), has a short generation period,

a high reproductive rate, reasonable longevity, is easily cultured, and requires little space for rearing. This bivalve can be raised to metamorphosis in containers as small as one liter.<sup>127</sup> The soft-shell clam (Mya arenaria), often used as a bioassay organism, has been cultured and induced to spawn in prepared seawater.<sup>191</sup> Cyclic temperature fluctuations were used to induce ripe M. arenaria to spawn, and laboratory tests were undertaken to determine the basic environmental requirements of eggs and larvae of this species.<sup>192</sup> Stimuli used to induce Mytilus californianus to spawn were described by Young<sup>193</sup> and Carriker and Van Zandt.<sup>194</sup>

167. None of the reports involving bioassays with freshwater molluscs mentioned use of cultured specimens. Several freshwater river mussels, however, have been propagated successfully.<sup>195</sup>

#### Arthropods

168. Numerous crustaceans and some insects have been laboratory-reared for use in bioassays and other research. The methods for laboratory culture have been developed for some crustaceans because of their commercial value; while methods for rearing smaller crustaceans and insects for bioassay use were devised due to these animals' ecological significance, ease of handling, short life cycle, and ubiquity.

169. Several cladocerans, especially Daphnia magna, have been used extensively in bioassays because of their suitability as laboratory organisms.<sup>196</sup> The optimal photoperiod, temperature and feeding necessary to culture various species of Daphnia have been documented.<sup>197-199</sup> Field-collected Daphnia should be reared through at least two generations before commencing chronic toxicity tests.<sup>200</sup>

170. The brine shrimp (Artemia salina), often selected for use as a bioassay organism, are readily available as eggs from pet shops, easy to handle, usable in in-plant effluent surveys, and easily cultured. They can be sustained on daily rations of cultured yeasts. There is evidence that evaporation of the culture medium and replenishment with distilled water eliminated the repressive influence of accumulated secretions and excretions.<sup>201</sup> Several culture devices have been developed which facilitate continuous circulation and aeration of the culture

medium.<sup>202</sup> Currently, however, brine shrimp are not favored by EPA due to their general resistance and the fact that they are not typical or representative marine organisms.

### Fishes

171. Many of the fishes often used in bioassays are acquired from cultured populations. These fishes have been reared in various aquaria, tanks, troughs, and ponds from fertilized eggs obtained from captured and cultured female adults. Several texts and publications contain general recommendations and methods for rearing fishes.<sup>143,144,203-206</sup>

172. Among the marine species Cyprinodon variegatus, Morone saxatilis, Clupea harengus, and several European species, including the plaice (Pleuronectes platessa), have been cultured from eggs for use in bioassays.<sup>207,208</sup> Northern anchovy larvae (Eugraulis mordax) have been successfully reared in laboratory containers where they were fed dino-flagellates, gastropod veligers, and brine shrimp nauplii.<sup>209</sup> Another oceanic species, the pilchard (Sardinia pilchardus) has been raised from larvae in circular tanks where they were fed natural offshore live plankton.<sup>210</sup> Several species of flatfish not used in the bioassays surveyed have been successfully reared.<sup>211</sup> Shelbourne<sup>212</sup> reviewed American and European methods of propagating marine fishes, experimental data, and recommended methods of rearing pelagic and demersal species.

173. Most of the North American salmonids, particularly the commercially valuable anadromous species, have been reared from eggs in large numbers in hatchery operations. Among the numerous bioassays conducted with salmonids, the majority of the specimens used were obtained as eggs or fry from cultured populations. Mason and Fessler<sup>213</sup> discussed use of a simple apparatus for incubating salmonid eggs at controlled levels of temperature, water flow, and oxygen. Culturing methods, spawning characteristics, and early life history of the commonly selected brook trout (Salvelinus fontinalis) are described.<sup>214-216</sup> The recommended bioassay procedures for the brook trout include information on all aspects of culturing this species.<sup>217</sup> The incubation period for trout eggs can be shortened by heating the water flowing through the incubators.<sup>218</sup> Poon and Johnson<sup>219</sup> recommended procedures for transporting



salmonid eggs, important in bioassays where experimentation is carried out at a location distant from the hatchery where eggs are acquired. A compact recirculation system can be used to rear trout with low mortality and satisfactory growth rates.<sup>220</sup> Burrows and Combs<sup>221</sup> reviewed the use of controlled environments for propagation of salmon, and Burrows and Chenoweth<sup>222</sup> described rectangular circulating ponds as having higher operational efficiency than other types of ponds for rearing salmon. Brown trout (Salmo trutta) have been raised in glass aquaria during growth rate research.<sup>223</sup> Nutritional data for raising fingerling salmon are provided by Fowler et al.<sup>224</sup> Most salmonids raised for use in bioassays were fed prepared dry foods and chows, which has been found to contain DDT, chopped liver, or moist pellets.<sup>225</sup> Thus, DDT and residues have been detected in various salmonids acquired from hatcheries and commercial suppliers, resulting in 30 to 90 percent mortalities during the days following the swim-up stage.

174. Cultivated aquarium goldfish (Carassius auratus) are quite hardy, do well in aquaria, tanks, bowls, and ponds at room temperature; and breed successfully at one year of age.<sup>226</sup> Another aquarium species, the flagfish (Jordanella floridae), is readily available through aquarium shops, easily maintained, and reared. "Spawning mops" made of nylon yarn can be used to collect the fish's eggs.<sup>227</sup> Adult flagfish over 2.5 cm in length should be fed frozen adult brine shrimp ad libitum twice daily, supplemented by one daily feeding with a high-quality, fine-granule dry trout food. Only F<sub>1</sub> or later generations should be used in bioassay testing.<sup>228</sup> The inbred platyfish (Xiphophorus maculatus) has been suggested as an animal that can be used for studies of aquatic toxicants, and has been recommended as a bioassay animal.<sup>229</sup>

175. The fathead minnow has been cultured often for the purpose of providing a continuous supply of eggs or fish of known age, and raised under identical conditions for use in bioassays. Procedures for establishing and operating a fathead minnow stock culture are recommended by EPA.<sup>92</sup>

176. Guppies (Lebistes reticulatus) can be maintained and cultured easily in the laboratory where the females produce broods about

once per month. The adults are fed prepared dry foods, brine shrimp, or other common aquarium fish foods.<sup>229,230</sup>

177. Although the warmouth (Lepomis gulosus) was not used in the reports surveyed, it is suitable as it will nest and spawn in laboratory aquaria.<sup>231</sup>

178. The bluegill (Lepomis macrochirus) can be raised in the laboratory for use in bioassays. Procedures for maintaining adults, gathering eggs, incubation, and rearing of young are described by the EPA.<sup>232</sup>

179. Most of the fishes listed in Table 1 and some of those listed in Table 2 have been cultured. Among those species cultured for use in bioassays, besides those discussed immediately above, are the walleye (Stizostedion v. vitreum), sucker (Catostomus commersoni), lake chubsucker (Erimyzon sucetta), stoneroller (Campostoma anomalum), largemouth bass (Micropterus salmoides), smallmouth bass (M. dolomieu), green sunfish (Lepomis cyanellus), redear sunfish (L. microlophus), pumpkinseed (L. gibbosus), channel catfish (Ictalurus punctatus), black catfish (I. melas), yellow perch (Perca flavescens), carp (Cyprinus carpio), striped bass (Morone saxatilis), northern pike (Esox lucius), golden shiner (Notemigonus crysoleucas), and bluntnose shiner (Notropis heterolepis). Yellow perch often do not accept trout food pellets but accept mixtures of trout food and ground beef liver.<sup>233</sup> The white sucker has spawned and been reared in the laboratory being fed brine shrimp when young and trout food when older.<sup>234</sup> Apparatus and methods for rearing carp and northern pike were discussed by Blaylock and Griffith<sup>235</sup> and Adelman and Smith,<sup>236</sup> respectively.

180. In bioassays involving use of marine organisms, artificial rather than natural seawater may be used to minimize biological effects and to provide a reproducible solution of known composition. Several commercial salt preparations are available and usually result in successful bioassays. Kester et al.<sup>237</sup> provided a method for preparing artificial seawater. The design and operation of open seawater systems using natural seawater have been reviewed and described.<sup>238-240</sup> A closed recirculated seawater system was described by Parisot.<sup>241</sup>

181. Apparatus for rearing aquatic plants and animals is

described by Sudia<sup>242</sup> and Whitford and Dillard.<sup>243</sup> This equipment provides for controlling water flow, photoperiod, and other critical parameters.

#### Wild-Caught Versus Stock-Cultured Test Organisms

182. As discussed above, many test species have been acquired as wild specimens with nets, seines, traps, etc. and have also been cultured in the laboratory. Other species have only been obtained from wild populations and a few (mostly exotic tropical fishes) have been acquired only as inbred cultured organisms. It is possible that wild-caught and cultured specimens of one species may differ in toxicity responses due to their environmental background or genetic makeup. In bioassays of zinc and copper with the Atlantic salmon (Salmo salar) collected from a river and from a federal hatchery, no differences in avoidance response were apparent in fish from the two sources.<sup>244</sup> However, Irwin<sup>245</sup> observed that a species population in a continually polluted stream will be composed of fish with higher resistance than the original population or one collected from an unpolluted source. Some species acquire resistance to pollutants when living in affected streams.<sup>246</sup> All specimens used in a single test should be acquired from the same source.

183. Since samples of a single species of fish from two or more sources may differ greatly in their response to a given chemical, Lennon<sup>247</sup> strongly urged that fish toxicologists follow the lead of mammalian toxicologists and pharmacologists and develop standard reference strains of fish for reproducible and comparable bioassays. Many species, particularly freshwater fishes, have been repeatedly cultured in inbred stocks in attempts to eliminate from laboratory experiments the natural biological variability among members of a population and to insure that the test organisms have spent their lives in the same controlled, reproducible medium. Acclimation or immunity to toxicants is also prohibited via the culturing method. However, there is the hazard that the cultured organisms will acclimate to unnatural

laboratory conditions or conditions which do not undergo natural diurnal, monthly, or seasonal cycles, or that the inbred population will be selective in succeeding generations for genetic factors not normally selected for; thus leading to a population gene pool unlike that of a natural population. Therefore, toxicity response may differ between wild-caught and cultured specimens, and data collected in laboratory tests with laboratory-reared animals may not be applicable to natural populations living in streams, ponds, and oceans.

#### Recommended Test Species

184. Plant and animal species often recommended and chosen for use in bioassays are those which meet most or all of the nine criteria listed earlier. The selection criteria may also include uniformity of size, small or practical size, and availability in a healthy disease-free condition. Several recent publications have included lists of recommended test organisms, particularly fishes, in discussions of bioassay methods.

185. Although any fish species which suits the purpose of the investigation may be used, Doudoroff et al.,<sup>89</sup> recommended selection of fish species belonging to any of the following widely distributed and important families, unless there is sufficient reason for making a different choice:

Centrarchidae (sunfishes, basses, crappies)

Salmonidae (trouts, charrs, salmons)

Cyprinidae (true minnows), exclusive of carp and goldfish

Catostomidae (suckers)

186. Henderson and Tarzwell<sup>88</sup> discussed fish species which have been used successfully in bioassays. The authors feel that the fish should be readily available, adaptable to laboratory conditions, and should be of small size. They describe the fathead minnow as an excellent choice based on their selection criteria. Other fish species are also described which have been used for bioassays.

187. Henderson and Pickering<sup>248</sup> suggest the use of the guppy

(Lebestes reticulatus), mosquitofish (Gambusia affinis), goldfish (Carassis auratus), fathead minnow (Pimephales promelas), and bluegill (Lepomis macrochirus) in warm-water bioassay studies. While other species of sunfish, bass, and minnows may also be used, they caution the use of certain minnows (Notropis) as they are difficult to maintain in the laboratory. Although cold-water species are generally not suitable for warm-water studies because of their temperature requirements, trout can be used if water temperature is carefully controlled.<sup>248</sup>

188. Irwin<sup>245</sup> described the techniques for capture, transfer, transport, and feeding of fish as being critical to the success of a bioassay, and discussed several problems encountered in using various species. The white bass (Morone chrysops), for example, went into tetany when placed in shallow water, and the gizzard shad (Dorosoma cepedianum) leaped out of the tanks if not handled properly. He listed several criteria for selecting a test fish: availability in large numbers, proper (small) size, health, adaptability to laboratory life, metabolic rate, acceptance of food, and no tendency to jump out of test containers.<sup>245</sup> He also presented a ranking of 57 species of freshwater fishes as to their suitability as test animals in oil refinery waste bioassays (see Table 5).

189. Other researchers have proposed lists of suitable test species.<sup>90,142,249</sup> A comparison of these lists are presented in Table 6 along with a list of commonly selected species compiled from a recent survey of bioassay research.<sup>105</sup> The list of commonly selected species in the Battelle survey is nearly identical to that compiled in the present study (Table 1).

190. Mount<sup>90</sup> compiled two lists of fishes for use in bioassay research based on the recommendations of the staff at the EPA Laboratory, Duluth, Minnesota, and recommendations of U. S. authorities (see Table 6). Those species included in the second list (for selected pollutants) were described as having limited distribution or having special requirements but also suitable for study in regard to certain kinds of pollution. The goldfish was selected as the species most meeting the description of the piscine equivalent of the laboratory white rat.

191. Among numerous groups and species studied, Tarzwell<sup>250</sup> selected four species for use in bioassays: the fathead minnow (Pimephales promelas) for freshwater tests; the mummichog (Fundulus heteroclitus), eastern oyster larvae (Crassostrea virginica), and brine shrimp larvae (Artemia salina) for marine tests.

192. Finally, Becker et al.<sup>93</sup> recommended use of many marine vertebrates, invertebrates, and plants in oil and dispersant bioassays in eight geographic areas of the United States. Many of the species selected were recommended for more than one geographic region and many are of some economic value or local ecologic significance. Other selection criteria used were vulnerability to marine pollutants such as oil slicks in a critical life stage, availability and ease of collection, ease of rearing and maintaining in the laboratory, occurrence in intertidal or estuarine areas, and existing knowledge on ecological requirements. The species included in the lists of key reference biota discussed above are compiled in Table 7.

193. Based on the information gathered in this survey on test species suitability and on the recommendations and conclusions of reports discussed above, the authors recommended selection of test species from the lists presented in Tables 1 and 7. Selection of these species is contingent upon local availability.

194. In addition to those species listed in Table 1, various other organisms appear to be suitable or have been recommended by authoritative toxicologists. These include the algae Anabaena flos-aquae, Microcystis aeruginosa, and Selenastrum capricornutum; the molluscs Mulinia lateralis, and Rangia cuneata; the crustaceans Acartia tonsa, Asellus militarum, Cancer magister, Carcinus maenas, Gammarus oceanicus, Orconectes nais, Procambarus clarkii; the insects Classenia sabulosa, Ephemera simulans; the fishes Clupea harengus, Jordanella floridae, Xiphophorus maculatus. Numerous other plants and animals have been used successfully in laboratory and field bioassays and are suitable for use, but due to their limited distribution or other factors have not been used extensively. If the investigators are familiar with these less frequently used species and these species have proven to be

suitable for bioassay use, then they should be selected if the more commonly used species are not available or if the use of the commonly used species will not provide the data needed in the research being conducted. That is, any species which suits the purpose of the anticipated research should be used; however, those species which have been used often before, and for which there are previously acquired data, should be given first consideration.

#### PART IV: DREDGED MATERIAL BIOASSAY DEVELOPMENT

195. Prior to about 1970, the only Federal regulatory control of dredging was under the River and Harbor Act of 1899, which required that a permit be obtained from the Corps of Engineers for virtually all work in navigable waters which might affect navigation. In the 1960's public awareness and concern over environmental problems increased, with the greatest initial concern over dredged material disposal in the Great Lakes region. In the late 1960's at the request of the Federal Water Quality Administration (FWQA), the Corps of Engineers Buffalo District initiated studies on the chemical characteristics of selected Great Lakes harbors. The earliest guidelines proposed for the regulation of dredging and dredged material disposal based on results of the Great Lakes Survey, were promulgated in 1971 by the EPA. That year the EPA and the Corps of Engineers issued Engineering Circular 1165-2-97 based on EPA's memo which stated that under Section 10 of the 1899 Refuse Act the dredged material disposal "bulk sediment" chemical criteria formulated by the EPA, commonly called "the Jensen Criteria," should be applicable to sediments dredged from all U. S. waters.

196. Regulation of dredged material disposal under the 1899 Refuse Act came to an end in 1972. That year Congress passed the Marine Protection, Research and Sanctuaries Act of 1972 (P.L. 93-532), and the Federal Water Pollution Control Act Amendments of 1972 (P.L. 92-500). It was then mandated that the Corps of Engineers, using criteria and guidelines developed by EPA in consultation and conjunction, respectively, with the Corps, use Section 103, P.L. 92-532 to regulate dredged material disposal in marine waters, and Section 404, P.L. 92-500 to regulate dredged material and fill material disposal in freshwater and fill material disposal in marine waters.

197. Testing of dredged material prior to disposal is required both by P.L. 92-500 and 92-532 to evaluate the potential environmental impact of disposal operations. As presently constituted, the guidelines for implementing Section 404 of the Federal Water Pollution Control Act Amendments of 1972 (P.L. 92-500) call for evaluation of the physical



and chemical-biological interactive effects of dredged material disposal, and a site selection evaluation. Tests which may be conducted to evaluate chemical-biological interactive effects include the elutriate test, water column and benthic bioassays.

198. The criteria for ocean disposal of dredged material under Public Law 92-532 take into account provisions of the Convention on the Prevention of Marine Pollution by Dumping of Wastes and Other Matter. The Convention, of which the U. S. is a signatory party, was held in November 1972 and became binding international law on 31 July 1975. The Convention bans the ocean dumping of materials containing other than traces of a long list of compounds. Compounds on the prohibited list are considered to be present in trace quantities when the dumping of dredged sediments containing these contaminants will not cause significant undesirable effects.

199. The potential for undesirable impacts of dredging and dredged material disposal are assessed under the ocean dumping criteria by means of liquid, suspended particulate, and solid phase bioassays. The bioassays required by the criteria make possible the direct assessment of potential environmental effects on aquatic organisms during and after the dredged material disposal operations. This direct determination of the biological effects is much more accurate than attempting to infer biological effects from chemical constituent concentrations in water or sediment. Bioassays that reflect conditions at the disposal site provide a solid technical basis upon which regulatory decisions can be defensibly implemented.

200. It is officially recognized that the present state-of-the-art does not allow the development of "final" criteria, test procedures, and similar decision-making guidance for either Section 103 of Public Law 92-532 or Section 404 of Public Law 92-500. Provisions are therefore present in both laws whereby periodic review and revision or updating are possible as more is learned through both research and experience. Research is continuing to close knowledge gaps and develop better disposal criteria and guidelines. Environmental research relative to dredging and disposal is being closely coordinated between the Corps and EPA

under the auspices of the EPA/CE Technical Committee on Criteria for Dredged and Fill Material. This joint technical committee was formed to focus attention on research coordination and needs and is staffed with senior level scientists from both agencies with broad knowledge and responsibilities for research activities. The Committee has published a technical procedures implementation manual for the ocean dumping criteria,<sup>831</sup> and will prepare a similar manual for implementation of the final regulatory guidelines on Section 404 of P.L. 92-500.

201. Bioassay studies conducted with dredged material have been limited in number and scope. Two general types of dredged material bioassays are of interest; those addressing water column effects and those concerned with effects upon benthic organisms.

202. Workers conducting water column bioassays have investigated a limited number of organisms and methods of bioassay water and sediment preparation. Emerson<sup>832</sup> used benthic polychaetes and sediment extracts of varying sediment/water ratios. Hoss et al.<sup>833</sup> used sediment extracts made from seawater and marine sediments to determine the effects of soluble compounds released from the sediments on larval fish. Both found that the sediment/water ratio used in preparing the extract was important, and that extracts of some sediments were toxic to some organisms after prolonged exposure to high concentrations. Both liquid and suspended particulate phases of Bailey Creek, Virginia, sediments were toxic to Daphnia at 100 percent concentration; that is, when no allowance was made for the mixing and dilution required by the guidelines.<sup>834</sup> The soluble and particulate phases of Perth Amboy and Bay Ridge, New York, sediments also showed some toxicity to the estuarine copepod Acartia tonsa, grass shrimp (Palaemonetes sp.) larvae, and adult opossum shrimp (Mysidopsis bahia) when tested at 100 percent concentration.<sup>834</sup> No toxicity was observed when the bioassays were conducted with elutriate preparations mixed with disposal site water at concentrations representative of observed mixing conditions, as is required by the criteria and guidelines. Thus it can be seen that the current bioassay techniques tend to overestimate sediment pollution potential.

203. The short-term impact of dredging and disposal operations

upon aquatic test organisms has generally been negligible. Shuba et al.<sup>835</sup> investigated the effect of the liquid phase of dredged material upon various aquatic organisms. An inhibitory effect on algal growth was found only in the liquid phase of sediments from Mobile Bay, Alabama. The inhibition was noted under "worst case" conditions when the elutriate was used full strength with no dilution by disposal site water and is indicative of conditions in a barge, hopper, or pipeline rather than the water column immediately after disposal. Algae showed promise as a group of organisms for use in regulatory testing. Bacteria and protozoans were found unacceptable for use in dredged material bioassays because of poor response to test solutions. Development of an algal bioassay was continued by Shuba et al.<sup>834</sup> and refined to the point where it could be incorporated into dredged material evaluation procedures.<sup>831, 836</sup>

204. Lee et al.<sup>837</sup> conducted static standard elutriate bioassays and found the grass shrimp (P. pugio) sensitive to 96-hr continuous exposure to high concentrations of elutriates prepared from some sediments.

205. Lee et al.<sup>838</sup> reported that laboratory bioassays with unfiltered elutriates (suspended particulate phase) have shown very little toxicity to aquatic organisms during worst case simulations of disposal in open waters. Daphnia or grass shrimp (Palaemonetes) survival for 96 hours in the laboratory without significant mortality was observed in the equivalent of a settled discharge from a dredging pump or hopper and barge bin.<sup>838</sup>

206. Research results have generally shown that the soluble and particulate phases of dredged materials exhibit little or no toxicity with minimal or very conservative mixing and initial dilution. It is highly unlikely that the toxicity which occurred with prolonged exposure to high concentrations would be observed in the field. Such conditions occur only inside the dredge pipe or hopper and barge bin. The intermittent nature of discrete dumping operations and the relatively rapid dispersion<sup>838,839,840</sup> of released contaminants makes the likelihood of significant acute toxicity to aquatic organisms extremely remote. A suspended particulate phase bioassay suitable for a large variety of

aquatic organisms is presented in detail in the Implementation Manual for Section 103 of Public Law 92-532.<sup>831</sup> Algal bioassays are not recommended for use in the suspended particulate phase because of interferences and predation on the test species by indigenous protozoans in the dredged material being tested and suspended particulates in the test media. Zooplankton were recommended as a test species for suspended particulate bioassays in place of algae.<sup>831</sup>

207. The impact of aquatic disposal of dredged materials upon benthic and epibenthic organisms may potentially be more pronounced than water-column impacts. These organisms are in close contact with deposited dredged materials for long periods of time in contrast to the short exposure of water column organisms.

208. Benthic bioassays with dredged material have been even more limited than water column bioassays with dredged materials. Gannon and Beeton<sup>841</sup> conducted benthic bioassays and sediment selectivity tests with dredged material. Their results, however, are open to question since oxygen depletion in the overlying water caused by high oxygen demand from some sediments may have caused the death of test organisms, rather than any toxic materials that may have been present in the sediments.

209. Bioassay results suggest that dredged sediments can, under some circumstances, exert adverse effects upon benthic and epibenthic organisms that survive burial during the disposal operation or recolonize the site after disposal operations cease. Benthic bioassays showed some degree of toxicity to freshwater grass shrimp (P. kadiakensis) following 6 days of exposure to Bailey Creek, Virginia, sediments.<sup>834</sup> Benthic bioassays also showed that sediments from Bay Ridge Channel in New York City were toxic to M. bahia (oppossum shrimp); sediments from Perth Amboy Channel in New York City were toxic to grass shrimp larvae; and Vicksburg, Mississippi, sediments subject to sewage and chemical plant contamination were highly toxic to adult grass shrimp.<sup>834</sup>

210. Lee et al.<sup>838</sup> evaluated the accumulation of chlorinated hydrocarbon pesticides and PCB's within aquatic organisms residing in the area of several aquatic disposal sites. Even in sediment containing

very high concentrations of these compounds, none of the organisms in the monitored sites exhibited elevated body burdens. Shuba et al.<sup>834</sup> did note uptake and release of kepone by the Asiatic clam Corbicula fluminea. The work of Shuba et al.<sup>834</sup> in conjunction with similar work by Swartz et al.<sup>842</sup> of the US EPA has led to the development of practical benthic bioassay procedures which have been suggested for use in implementation of dredged material disposal criteria.<sup>831</sup> The benthic bioassay procedures are designed to approximate conditions found within or at disposal site boundaries and can determine whether a biological effect is likely. The benthic bioassay procedure measures the overall chemical impact of the dredged material, and cannot be used to determine the constituent causing any biological effect observed.<sup>831</sup> The benthic bioassay procedure can be used to determine the sensitivity of a wide variety of benthic organisms.<sup>834,842</sup>

211. Benthic bioassay procedures, due to their recent development are relatively basic and much work remains before more sophisticated tests can be developed. At present there is no quantitative method for estimating the meaning of a difference between exposed and control test animals and how that difference might be assumed to predict the occurrence of ecologically adverse impacts in the field. Before this need can be addressed, additional study is needed to determine the consistency and variability of current and projected benthic bioassay procedures.

212. Relatively simple and implementable liquid, suspended particulate, and solid phase bioassays have been developed for assessing the impact of dredging and disposal operations upon aquatic organisms. The bioassay results must be interpreted in relation to the dilution that occurs under field conditions. Detailed bioassay interpretation procedures are given in the appropriate dredged material disposal criteria implementation manuals.<sup>831,836</sup>

## REFERENCES

1. Boyd, M. B., et al., "Disposal of Dredge Spoil--Problem Identification and Assessment and Research Program Development," U. S. Army Engineer Waterways Experiment Station, 1972, Technical Report H-72-8.
2. Goldstein, Avram, Aronow, Lewis, and Kalman, Sumner, "Principles of Drug Action," The Basic Pharmacology, Harper and Row, New York, 1969, 884 p.
3. Bliss, C. I. and Cattell, McKeen, "Biological Assay," Annual Review of Physiology, Vol 5, 1943, pp 479-573.
4. Bliss, C. I., "Some Principles of Bioassay," American Scientist, Vol 45, 1957, pp 449-466.
5. Clark, A. J., General Pharmacology, Handbuch der Experimentellen Pharmakologie, Verlag von Julius Springer, Berlin, 1937.
6. Gaddum, J. H., "Bioassay and Mathematics," Pharmacological Review, Vol 5, 1953, pp 156-173.
7. Gaddum, J. H., "The Estimation of the Safe Dose," British Journal of Pharmacology, Vol 11, 1956, pp 156-173.
8. Litchfield, J. T. and Wilcoxon, F., "A Simplified Method of Evaluating Dose-Effect Experiments," Journal of Pharmaceudics and Experimental Therapeudics, Vol 96, 1949, pp 99-113.
9. Litchfield, J. T., "A Method for Rapid Graphic Solution of Time-Percent Effect Curves," Journal of Pharmaceudics and Experimental Therapeudics, Vol 96, 1949, pp 99-113.
10. Burdick, G. E., "A Graphical Method for Deriving Threshold Values of Toxicity and Equation of the Toxicity Curve," New York Fish and Game, Vol 4, 1957, pp 102-108.
11. Harris, E. K., "Confidence Limits for the LD<sub>50</sub> Using the Moving Average-Angel Method," Biometrics, Vol 15, 1959, pp 424-432.
12. Sprague, J. B., "Measurements of Pollutant Toxicity to Fish, I. Bioassay Methods for Acute Toxicity," Water Research, Vol 3, 1969, pp 793-821.
13. Erichsen-Jones, J. R., Fish and River Pollution, Butterworths and Company, Ltd., London, 1964, 203 p.
14. Katz, M., "Toxicity Bioassay Techniques Using Aquatic Organisms," Water and Water Pollution Handbook, Vol 2, 1971, pp 763-800.
15. Waldichuk, M., "Effects of Pollutants on Marine Organisms: Improving Methodology of Evaluation--A Review of the Literature," Journal of the Water Pollution Control Federation, Vol 41, 1969, pp 1586-1601.

16. Lloyd, R., "Problems in Determining Water Quality Criteria for Freshwater Fishes," Proc. Royal Society of London, Series B., Vol 180, 1972, pp 439-658.
17. Weiss, C. M., "Use of Fish to Detect Organic Insecticides in Water," Journal of the Water Pollution Control Federation, Vol 37, 1965, pp 647-658.
18. Weiss, C. M., "Use of Fish to Detect Organic Insecticides in Water," Proc. 19th Industrial Waste Conf., Engineering Extension Series #117, Purdue University, 1967, pp 112-125.
19. Muirhead-Thomson, R. C., Pesticides and Freshwater Fauna, Academic Press, New York, 1971, 248 p.
20. Skidmore, J. F., "Toxicity of Zinc Compounds to Aquatic Animals, with Special Reference to Fish," The Quarterly Review of Biology, Vol 39, No. 3, 1964, pp 227-248.
21. Doudoroff, P. and Katz, M., "Critical Review of Literature on the Toxicity of Industrial Wastes and Their Components to Fish, II. The Metals, as Salts," Sewage and Industrial Wastes, Vol 25, 1953, pp 802-839.
22. Ball, I. R., "The Relative Susceptibilities of Some Species of Fresh-Water Fish to Poisons, II. Zinc," Water Research, Vol 1, 1967, pp 777-783.
23. Becker, C. D. and Thatcher, T. O. for USAEC, "Toxicity of Power Plant Chemicals to Aquatic Life," Battelle Pacific Northwest Laboratories, Richland, Washington, 1973.
24. Doudoroff, P. and Katz, M., "Critical Review of Literature on the Toxicity of Industrial Wastes and Their Components to Fish, I. Alkalies, Acids, and Inorganic Gases," Sewage and Industrial Works, Vol 22, 1950, pp 1432-1458.
25. Ball, I. R., "The Relative Susceptibilities of Some Species of Fresh-Water Fish to Poisons, I. Ammonia," Water Research, Vol 1, 1967, pp 767-775.
26. Lee, G. F., "Chemical Aspects of Bioassay Techniques for Establishing Water Quality Criteria," Water Research, Vol 7, 1973, pp 1525-1546.
27. European Inland Fisheries Advisory Commission, Working Party on Water Quality Criteria for European Freshwater Fish, "Water Quality Criteria for European Freshwater Fish, Report on Ammonia and Inland Fisheries," Water Research, Vol 7, 1973, pp 1011-1022.
28. Huet, M., "Water Quality Criteria for Fish Life," Biological Problems in Water Pollution, 3rd Seminar, 1962, USHEW Publ. 999-WP-25.
29. Federal Water Pollution Control Administration, "Water Quality Criteria Report of the National Technical Advisory Committee to the Secretary of the Interior," GPO, 1968, Washington, D.C.

30. Aquatic Life Advisory Committee of the Ohio River Valley Water Sanitation Commission, "Stream Pollution. Aquatic Life Water Quality Criteria," Second Progress Report, Sewage and Industrial Wastes, Vol 28, 1956, pp 678-690.
31. European Inland Fisheries Advisory Commission Working Party on Water Quality Criteria for European Freshwater Fish, "Water Quality Criteria for European Freshwater Fish--Extreme pH Values and Inland Fisheries," Water Research, Vol 3, 1969, pp 593-611.
32. Katz, M. and Woelke, C. E., "Water Quality Requirements of Estuarine Organisms," ASTM Special Publication, Vol 416, 1967, pp 90-99.
33. Alderdice, D. F., "The Detection and Measurement of Water Pollution--Biological Assays," 1967, Canadian Fisheries Report No. 9.
34. Sprague, J. B., "Measurement of Pollutant Toxicity to Fish, II. Utilizing and Applying Bioassay Results," Water Research, Vol 4, 1970, pp 3-32.
35. Sprague, J. B., "Measurement of Pollutant Toxicity to Fish, II. Sublethal Effects and "Safe" Concentrations," Water Research, Vol 5, 1971, pp 245-266.
36. Warren, C. E. and Doudoroff, P., "The Development of Methods for Using Bioassays in the Control of Pulp Mill Waste Disposal," TAPPI, Vol 48, No. 8, 1958, pp 211A-216A.
37. Betts, J. L., Beak, T. W. and Wilson, G. G., "A Procedure for Small-Scale Laboratory Bioassays," Journal of the Water Pollution Control Federation, Vol 39, 1967m pp 89-96.
38. LaRoche, G. Eisler, R. and Tarzwell, C. M., "Bioassay Procedures for Oil and Oil Dispersant Toxicity Evaluation," Journal of the Water Pollution Control Federation, Vol 42, No. 11, 1970, pp 1982-1989.
39. Lichatowich, J. A., O'Keefe, P. W., Strand, J. A. and Templeton, W. L., "Development of Methodology and Apparatus for the Bioassay of Oil," Proc. of the Joint Conf. (in) Prevention and Control of Oil Spills, March 13-15, 1973, Washington, D.C.
40. Alderdice, D. F., "Some Effects of Simultaneous Variation in Salinity, Temperature and Dissolved Oxygen to the Resistance of Young Coho Salmon to a Toxic Substance," Journal of the Fisheries Research Board of Canada, Vol 20, No. 2, 1963, pp 525-543.
41. Lloyd, R., "Factors That Affect the Tolerance of Fish to Heavy Metal Poisoning," (in) Biological Problems in Water Pollution, 3rd Seminar, 1962, USHEW Publication 999-WP-25.
42. Durham, W. F., "The Interaction of Pesticides with Other Factors," Residue Reviews, Vol 18, 1967.
43. Cherkin, A. and Catchpool, J. F., "Temperature Dependence of Anesthesia in Goldfish," Science, Vol 144, 1964, pp 1460-1461.



44. Cairus, J., Jr., "The Effects of Increasing Temperatures Upon Aquatic Organisms," Proc. 10th Industrial Waste Conf., Engineering Extension Series #89, Purdue University, 1956.
45. Cairns, J., Jr. and Scheirer, A., "The Effects of Periodic Low Oxygen Upon the Toxicity of Various Chemicals to Aquatic Organisms," Proc. 12th Industrial Waste Conf., Engineering Extension Series #97, Purdue University, 1957, pp 165-176.
46. Lloyd, R. and Jordon, D. H. M., "Some Factors Affecting the Resistance of Rainbow Trout (Salmo gairdneri Richardson) to Acid Waters," International Journal of Air and Water Pollution, Vol 8, 1964, pp 393-403.
47. Roberts, R. F. and Allen, H. E., "The Control of pH and Total Alkalinity or Total Carbonate in Aquatic Bioassays," Transactions of the American Fisheries Society, Vol 101, No. 4, 1974, pp 752-756.
48. Cairns, J., Jr. and Dickson, K. L., eds., 1973, "Biological Methods for the Assessment of Water Quality," ASTM Spec. Publ. 528, American Society for Testing and Materials, Philadelphia.
49. Weber, C., ed., "Biological Field and Laboratory Methods for Measuring the Quality of Surface Waters and Effluents," 1973, Environmental Monitoring Series EPA-670/4-73-001, July 1973, Cincinnati, Ohio.
50. Standard Methods for the Examination of Water and Wastewater, 13th ed., 1971, Prepared and published jointly by American Public Health Association, American Water Works Association, and Water Pollution Control Federation, New York, N.Y.
51. Lennon, R. E. and Walker, C. R., "Investigations in Fish Control 1. Laboratories and Methods for Screening Fish--Control Chemicals," Bureau of Sport Fisheries and Wildlife, Circular 185, June 1964.
52. Anon., "Summary of Marine Waste Disposal Research Program in California," State Water Pollution Control Board, Publ. #22, 1964.
53. Biological Water Quality Committee, "ORSANCO 24-Hour Bioassay," Ohio River Valley Water Sanitation Commission, January 1974.
54. Anon., "Recommended Bioassay Procedure for Brook Trout, Salvelinus fontinalis (Mitchill), Partial Chronic Tests, April, 1971 (Revised January, 1972), National Water Quality Laboratory, Duluth, Minnesota.
55. Anon., "Recommended Bioassay Procedure for Fathead Minnow Pimephales promelas (Rafinesque), Chronic Tests, April, 1971 (Revised January, 1972), National Water Quality Laboratory, Duluth, Minnesota.
56. Belding, D. L., "Toxicity Experiments with Fish in Reference to Trade Waste Pollution," Transactions of the American Fisheries Society, Vol 57, 1927, pp 100-119.

57. Hubble, D. R. and Reiff, B., "Reproduction of Guppies (Lebistes reticulatus) after a Single Exposure to Dieldrin--a 12 Months' Study," Bull. of Environmental Contamination & Toxicology, Vol 2, 1967, pp 57-63.
58. Davis, H. C. and Hidu, H., "Effects of Turbidity-Producing Substances in Sea Water on Eggs and Larvae of Three Genera of Bivalve Mollusks," The Veliger, Vol 11, 1969, pp 316-323.
59. Davis, H. C., "Effects of Turbidity-Producing Materials in Sea Water on Eggs and Larvae of the Clam (Venus (Mercenaria) mercenaria)," Biological Bull., Vol 118, No. 1, 1960, pp 98-54.
60. Coler, C. T., Gunner, A. K. and Gunner, E. J., "An Improved Bottom-Water Sampler," Journal of the Marine Biological Association, U. K., Vol 53, 1973, pp 741-744.
61. Drammo, F. R. and Kohlberg, J. R., "Controlled Aqueous Environments for Bioassay," Laboratory Practice, Vol 6(8), 1957, pp 456-457.
62. Anderson, R., "Effects of Low Concentrations of Free Chlorine on Eggs and Larvae of Plaice, Pleuronectes platessa L.," Marine Pollution and Sea Life, FAO, Fishing News, Ltd., 1972, pp 312-315.
63. Abrams, F. S. H., "An Automatic Dosage Apparatus," Laboratory Practice, Vol 9, 1960, pp 796-797.
64. Stark, G. T. C., "An Automatic Dosing Apparatus Made with Standard Laboratory Ware," Laboratory Practice, Vol 16, 1967, pp 594-595.
65. Burke, W. D. and Ferguson, D. E., "A Simplified Flow-Through Apparatus for Maintaining Fixed Concentration of Toxicants in Water," Transactions of the American Fisheries Society, Vol 97, 1968, pp 498-501.
66. Mount, D. L. and Brungs, W., "A Simplified Dosing Apparatus for Fish Toxicology Studies," Water Research, Vol 1, 1967, pp 21-29.
67. Heitmuller, B. W., Del Wayne, M. and Nimmo, F. H., "Quantitative Aspects of Filter Feeding in Invertebrates," Biological Review, Vol 30, 1955, pp 391-393.
68. Falk, R., "Aquatic Organisms as an Aid in Salvaging Waste Disposal Problems," Sewage and Industrial Waste, Vol 25(2), 1953, pp 210-213.
69. Dickson, K. L. and Cairns, J., "The Relationship of Fresh-Water Macroinvertebrate Communities Collected by Floating Artificial Substrates to the MacArthur-Wilson Equilibrium Model," American Midland Naturalist, Vol 81, 1972, pp 68-75.
70. Wene, G. and Eickliff, E. L., "Modification of the Stream Bottom and Its Effect on the Insect Fauna," Can. Entomologist, Vol 72, 1940, pp 131-135.
71. Henson, E. G., Jr., "A Cage Sampler for Collecting Aquatic Fauna," Turtlox News, Vol 43, 1965, pp 298-299.

72. Mason, W. T., Anderson, J. B. and Morrison, G. E., "Limestone Filled, Artificial Substrate Sampler-Float Unit for Collecting Macroinvertebrates in Large Streams," Progressive Fish Culturist, Vol 29, 1967, 74 p.
73. Kauss, P., Hutchinson, T. C., Soto, C., Hellebust, J. and Griffiths, M., "The Toxicity of Crude Oil and Its Components to Freshwater Algae," Proceedings, Joint Conference on Regulation and Control of Oil Spills, Washington, D.C., March 13-15, 1973.
74. Dickman, M., "A Quantitative Method for Assessing the Toxic Effects of Some Water Soluble Substances, Based on Changes in Periphyton Community Structure," Water Research, Vol 3, 1969, pp 963-972.
75. Zillich, J. A., "Toxicity of Combined Chlorine Residuals to Freshwater Fish," Journal of the Water Pollution Control Federation, Vol 44, No. 2, 1972, pp 212-220.
76. Dunstan, W. M. and Menzel, D. W., "Continuous Cultures of Natural Populations of Phytoplankton in Dilute, Treated Sewage Effluent," Limnol. and Oceanog., Vol 16, No. 4-6, 1971, pp 623-632.
77. Savage, H. P. and Hanes, N. B., "Toxicity of Seawater to Coliform Bacteria," Journal of the Water Pollution Control Federation, Vol 43, No. 5, 1971, pp 854-861.
78. Murray, S., Scherfig, J. and Dixon, P. S., "Evaluation of Algal Assay Procedures--PAAP Batch Test," Journal of the Water Pollution Control Federation, Vol 43, No. 10, 1971, pp 1991-2003.
79. Surber, E. W. and Thatcher, T. O., "Laboratory Studies of the Effects of Alkyl Benzene Sulfonate (ABS) on Aquatic Invertebrates," Transactions of the American Fisheries Society, Vol 92, No. 2, 1963, pp 152-160.
80. Knight, A. W. and Swallow, W. T., "A Flowing Reaction Vessel," Hydrobiologica, Vol 35, No. 1, 1970, pp 1-6.
81. Goodyear, C. P., "A Simple Technique for Detecting Effects of Toxicants or Other Stresses on Predator-Prey Interaction," Transactions of the American Fisheries Society, Vol 101, No. 2, 1972, pp 367-370.
82. Lee, E. L. and Buzzell, J. C., "Measurement of Pesticide Toxicity by Fish Respiration Rate," Proc. 24th Industrial Waste Conf., Engineering Extension Series #135, Purdue University, 1969, pp 595-609.
83. Gamble, J. C., "Effect of Low Dissolved Oxygen Concentrations on the Ventilation Rhythm of Three Tubicolous Crustaceans, with Species Reference to the Phenomenon of Intermittent Ventilation," Marine Biol., Vol 6, 1970, pp 121-127.
84. Lloyd, R. and Orr, L. D., "The Diuretic Response by Rainbow Trout to Sublethal Concentrations of Ammonia," Water Research, Vol 3, 1969, pp 335-344.

85. Jones, J. R. E., "The Reactions of Fish to Water of Low Oxygen Concentration," J. Exp. Biol., Vol 29, 1952, pp 403-415.
86. Cook, R. M. and Boyd, S. R., "Acute Toxicity of Heavy Metals to Some Marine Larvae," Marine Pollution Bulletin, Vol 3, 1972, pp 190-192.
87. Scherer and Nowak, M. D., "The pH Tolerance of the Common Blue-gill," Not. Nat., 256, 1964, 13 p.
88. Henderson, G. and Tarzwell, C., "Bioassays for Control of Industrial Effluents," Proc. 12th Industrial Waste Conf., Engineering Extension Series #94, Purdue University, 1957, pp 123-144.
89. Doudoroff, P. (chairman), et al., "Bioassay Methods for the Evaluation of Acute Toxicity of Industrial Wastes to Fish," Sewage and Industrial Wastes, Vol 23, 1951, pp 1380-1397.
90. Mount, D., "Test Animals for Water Quality," Amer. Fish. Soc. Newsletter, Vol 12, No. 54, 1968, pp 3,6.
91. Katz, M., "Toxicity Bioassay Techniques Using Aquatic Organisms," Water and Water Pollution Handbook, Vol 2, 1971, pp 763-800.
92. EPA, "Tentative Plans for the Design and Operation of a Fathead Minnow Stock Culture Unit," U. S. EPA Duluth, Minn., Manuscript, 1971, 6 p.
93. Becker, C. D., Lichatowich, J. A., Schneider, M. J. and Strand, J. A., "Regional Survey of Marine Biota for Bioassay Standardization of Oil and Oil Dispersant Chemicals," Amer. Petrol. Inst. Publ. No. 4167, 1973.
94. Eisler, R. and Weinstein, M. P., "Changes in Metal Composition of the Quahog Clam, Mercenaria mercenaria, After Exposure to Insecticides," Chesapeake Science, Vol 8, No. 4, 1967, pp 253-258.
95. Erickson, S. J., Maloney, T. E. and Gentile, J. H., "Effect of Nitritotriacetic Acid on the Growth and Metabolism of Estuarine Phytoplankton," Journal of the Water Pollution Control Federation, Vol 42, No. 8, 1970, pp R329-R335.
96. Erickson, S. J., Lackie, N. and Maloney, T. E., "A Screening Technique for Estimating Copper Toxicity to Estuarine Phytoplankton," Journal of the Water Pollution Control Federation, Vol 42, No. 8, 1970, pp R270-R279.
97. Larson, F. C. and Staub, R. J., "Effects of Industrial Effluents on Primary Phytoplankton Indicators," Report No. 26, 1972, Water Resources Research Center.
98. Walsh, G., Barrett, R., Cook, G. and Hollister, T., "Effects of Herbicides on Seedlings of the Red Mangrove, Rhizophora mangle L.," BioScience, Vol 23, No. 6, 1973, pp 361-364.
99. Steed, D. L. and Copeland, B. J., "Metabolic Responses of Some Estuarine Organisms to an Industrial Effluent," Contributions in Marine Science, Vol 12, 1967, pp 143-159.

100. Hansen, D. J., Schimmel, S. C. and Keltner, J. M., "Avoidance of Pesticides by Grass Shrimp (Palaemonetes pugio)," Bull. of Envir. Contam. & Toxicology, Vol 9, No. 3, 1973, pp 129-133.
101. Sanders, H. O., "Toxicities of Some Herbicides to Six Species of Freshwater Crustaceans," Journal of the Water Pollution Control Federation, Vol 42, No. 8, 1970, pp 1544-1550.
102. Kawatski, J. A. and Schmulbach, J. C., "Toxicities of Aldrin and Dieldrin to the Freshwater Ostracod, Chlamydotheca arcuata," J. of Econ. Entomol., Vol 64, No. 5, 1971, pp 182-185.
103. Kawatski, J. A. and Schmulbach, J. C., "Accumulation of Insecticide in Freshwater Ostracods Exposed Continuously to Sublethal Concentrations of Aldrin or Dieldrin," Trans. Amer. Fish. Soc., Vol 100, No. 3, 1971, pp 565-567.
104. Wilkes, F. G. and Weiss, C. M., "The Accumulation of DDT by the Dragonfly Nymph, Tetragoneuria," Trans. Amer. Fish. Soc., Vol 100, No. 2, 1971, pp 222-236.
105. Battelle-Columbus Laboratories, Water Quality Criteria Data Book Vol 3. Effects of Chemicals on Aquatic Life, EPA Water Pollution Control Research Series, 1971, 100 p.
106. Howard, T. E. and Walden, C. C., "Basic Bioassay Techniques," Pulp and Paper Mfg. Canada, Vol 73, No. 10, 1972, pp 85-89.
107. Salazar, M. H., Yamamoto, S., Shipman, W. H. and Zirino, A. R., "Informal Report on Lead and Chromium Pollution Study," (unpublished), Reg. 2501-145.
108. Mandelli, E. R., "The Inhibitory Effects of Copper on Marine Phytoplankton," Contributions in Marine Science, Vol 14, 1969, pp 47-57.
109. Walsh, G. E. and Grow, T. E., "Depression of Carbohydrate in Marine Algae by Urea Herbicides," Weed Sci., Vol 19, No. 5, 1971, pp 568-570.
110. Pratt, S. D., Saila, J. B., Gaines, A. G., Jr., and Kront, J. E., "Biological Effects of Ocean Disposal of Solid Waste," Marine Technical Report Series #9, 1972, Marine Experiment Station, Univ. of Rhode Island, pp 41-45.
111. Ottway, S., "The Comparative Toxicities of Crude Oils," The Ecological Effects of Oil Pollution on Littoral Communities, 1970, p.
112. Corner, E. D. S. and Sparrow, B. W., "The Modes of Action of Toxic Agents, I. Observations on the Poisoning of Certain Crustaceans by Copper and Mercury," Journ. Marine Biol. Assoc. UK, Vol 35, 1956, pp 531-548.
113. Allen, H., "Effects of Petroleum Fractions on the Early Development of a Sea Urchin," Mar. Poll. Bull., Vol 2, No. 9, 1971, pp 138-140.
114. Swedmark, M., Braaten, B., Emanuelsson, E. and Granmo, R., "Biological Effects of Surface Active Agents on Marine Animals," Marine Biology, Vol 9, 1971, pp 183-201.

115. Wells, P. G., "Influence of Venezuelan Crude Oil on Lobster Larvae," Mar. Poll. Bull., Vol 3, No. 7, 1972, pp 105-106.
116. Rachlin, J. W. and Perlmutter, A., "Response of an Inbred Strain of Platyfish and the Fathead Minnow to Zinc," Prog. Fish. Culturist, Vol 30, No. 4, 1968, pp 203-207.
117. Bellan, G., Reish, D. J. and Foret, J. P., "The Sublethal Effects of a Detergent on the Reproduction Development and Settlement in the Polychaetous Annelid Capitella capitata," Marine Biology, Vol 14, 1972, pp 183-188.
118. Katz, M. and Chadwick, G. G., "Toxicity of Endrin to Some Pacific Northwest Fishes," Trans. Amer. Fish. Soc., Vol 90, No. 4, 1961, pp 394-397.
119. Hiltibran, R. C., "Effects of Some Herbicides on Fertilized Fish Eggs and Fry," Trans. Amer. Fish. Soc., Vol 96, No. 4, 1967, pp 414-416.
120. Hoffman, C. H. and Surber, E. W., "Effects of Feeding DDT-Sprayed Insects to Freshwater Fish," Special Scientific Report - Fisheries #3, 1948.
121. Buhler, D. R. and Shanks, W. E., "Influence of Body Weight on Chronic Oral DDT Toxicity in Coho Salmon," J. Fish. Res. Bd. Canada, Vol 27, No. 2, 1970, pp 347-358.
122. Patrick, R., "Diatoms as Bioassay Organisms, In: Bioassay Techniques and Environmental Chemistry," Ann Arbor Science Publication, Inc., 1973.
123. Coles, S. L., "Quantitative Estimates of Feeding and Respiration for Three Scleractinian Corals," Limnol. & Oceanog., Vol 14, No. 6, 1969, pp 949-953.
124. Smith, A. J., "The Effect of the Lamprey Larvicide, 3-trifluoromethyl-4-Nitrophenol, on Selected Aquatic Invertebrates," Trans. Amer. Fish. Soc., Vol 96, No. 4, 1967, pp 410-413.
125. Dowden, F. F. and Bennett, H. J., "Toxicity of Selected Chemicals to Certain Animals," J. Water Poll. Contr. Fed., Vol 37, No. 9, 1965, pp 1308-1316.
126. Erichsen-Jones, J. R., "A Further Study of the Relation Between Toxicity and Solution Pressure, with Polycelis nigra as test Animal," J. Exp. Biol., Vol 17, 1940, pp 408-415.
127. Calabrese, A., "Mulinia lateralis: Molluscan Fruit Fly?," Proc. National Shellfisheries Assoc., Vol 59, 1970, pp 65-66.
128. Kennedy, V. S. and Mihursky, J. A., "Upper Temperature Tolerances of Some Estuarine Bivalves," Chesapeake Science, Vol 12, No. 4, 1971, pp 193-204.
129. Kennedy, V. S. and Mihursky, J. A., "Effects of Temperature on the Respiratory Metabolism of Three Chesapeake Bay Bivalves," Chesapeake Science, Vol 13, No. 1, 1972, pp 1-22.

130. Price, T. J., "Accumulation of Radionuclides and the Effects of Radiation on Molluscs," 3rd Seminar, 1972.
131. Reish, D. J. and Hetherington, W. M., III, "The Effects of Hyper- and Hypo-Chlorinities on Members of the Wood-boring Genus Limnoria," Marine Biol., Vol 2, 1969, pp 137-139.
132. Buchanan, D. V., Millemann, R. E. and Stewart, N. E., "Effects of the Insecticide Sevin on Various Stages of the Dungeness Crab, Cancer Magister," J. Fish. Res. Bd. Canada, Vol 27, No. 1, 1970, pp 93-104.
133. Perkins, E. J., "The Toxicity of Oil Emulsifiers to Some Inshore Fauna in the Biological Effects of Oil Pollution on Littoral Communities," The Biological Effects of Oil Pollution on Littoral Communities, 1968, pp 81-90.
134. Burrows, E. M., "Assessment of Pollution Effects by the Use of Algae," Proc. Roy. Soc. Lond. B., Vol 177, 1971, pp 295-306.
135. Anderson, B. G., "The Apparent Thresholds of Toxicity to Daphnia magna for Chlorides of Various Metals When Added to Lake Erie Water," Trans. Amer. Fish. Soc., Vol 78, 1948, pp 96-113.
136. Heimstra, N. W. and Damkot, D. K., "Some Effects of Silt Turbidity on Behavior of Juvenile Largemouth Bass and Green Sunfish," Technical Paper #20, Jan 1963, Bureau of Sport Fisheries, and Wildlife, U. S. Department of the Interior, Washington, D.C.
137. Lewis, W. M., "Isobornyl Thiocynoacetate as a Fish Drugging Agent and Selective Toxin," Prog. Fish. Cult., Vol 30, No. 1, 1968, pp 29-31.
138. Wiebe, A. H., "Notes on the Exposure of Young Fish to Varying Concentrations of Arsenic," Trans. Amer. Fish. Soc., Vol 60, 1930, pp 270-278.
139. Brown, D. H., "The Effect of Kuwait Crude Oil and a Solvent Emulsifier on the Metabolism of the Marine Lichen Lichina pygmaea," Mar. Biol., Vol 12, 1972, pp 309-315.
140. Naqui, S. M. and de la Cruz, A. A., "Mirex Incorporation into the Environment: Toxicity in Selected Freshwater Organisms," Bull. Envir. Contam. Toxicol., Vol 10, No. 5, 1973, pp 305-308.
141. Dobie, J., Meehan, O. L., Snieszko, S. F. and Washbur, G. N., "Raising Bait Fishes," USDI, Fish & Wildlife Serv. Cir., Vol 35, 1957, 123 p.
142. Hunn, J. B., Schoettger, R. A. and Whealdon, E. W., "Observations on the Handling and Maintenance of Bioassay Fish," Prog. Fish. Cult., Vol 30, No. 3, 1968, pp 164-167.
143. Randall, D. J. and Hoar, W. S., "Special Techniques, Maintenance of Fish," Fish Physiology, Environmental Relations and Behavior, Vol 6, W. S. Hoar and D. J. Randall, Editors, Academic Press, 1971, 559 p.

144. Spotte, S. H., "Fish and Invertebrate Culture," Wiley Interscience Publ., 1969, 145 p.
145. Fogg, G. E., "Algal Cultures and Phytoplankton Ecology," Univ. of Wisconsin Press, 1963, 126 p.
146. Weber, C. I., "Biological Field and Laboratory Methods for Measuring the Quality of Surface Waters and Effluents," EPA Environmental Monitoring Series, 1973, 67014-73-001.
147. Wedemeyer, G., "Some Physiological Consequences of Handling Stress in the Juvenile Coho Salmon (Oncochychus kisutch) and Steelhead Trout (Salmo gairdneri)," J. Fish. Res. Bd. Can., Vol 29, No. 12, 1972, pp 1780-1783.
148. Coleman, R. D., Coleman, R. L. and Rice, E. L., "Zinc and Cobalt Bioconcentration and Toxicity in Selected Algal Species," Bot. Gaz., Vol 132, No. 2, 1971, pp 102-109.
149. Keenan, J. D., "Response of Anabaena to pH Carbon, and Phosphorus," J. Envir. Eng. Div., Proc. ASCE, EE5, 1973, pp 607-620.
150. Bold, H. C., "The Cultivation of Algae," The Botanical Review, Vol VIII, No. 2, 1961, pp 69-138.
151. Kauss, P., Hutchinson, T. C., Soto, C., Hellebust, J. and Griffiths, M., "The Toxicity of Crude Oil and its Components to Freshwater Algae," Proceedings, Joint Conference on Prevention and Control of Oil Spills, Washington, D.C., March 13-15, 1973.
152. Conover, R. J., "Cultivation of Plankton Populations," Helgolauder Wiss Meertesunters, Vol 21, 1970, pp 401-444.
153. Davis, H. C. and Ukeles, R., "Mass Culture of Phytoplankton as Foods are Metazoans," Science, Vol 134, 1961, pp 562-564.
154. Ukeles, R., "A Simple Method for the Mass Culture of Marine Algae," Limnol. & Oceanog., Vol 10, No. 3, 1965, pp 492-495.
155. Loosanoff, V. L. and Engle, J. B., "Use of Complete Fertilizers in Cultivation of Microorganisms," Science, Vol 95, 1942, pp 487-488.
156. Ketchum, B. H. and Redfield, A. C., "A Method for Maintaining a Continuous Supply of Marine Diatoms by Culture," Biol. Bull., Vol 75, 1938, pp 165-169.
157. Drebes, G., "Conservation and Culture in the Laboratory, Planktonic Algae," Research Methods in Marine Biology, Curt Schlieper, Editor, Univ. of Washington Press, 1972, 356 p.
158. Sievers, A. M., "Comparative Toxicity of Gonyaulax moniclata and Gymnodinium breve to Annelids, Crustaceans, Molluscs and a Fish," J. Protozool., Vol 16, 1969, pp 401-404.
159. Murray, S., Scherfig, J. and Dixon, P. S., "Evaluation of Algal Assay Procedures--PAAP Batch Test," Journ. Wat. Poll. Contr. Fed., Vol 43, No. 10, 1971, pp 1991-2003.



160. Dickman, M., "A Quantitative Method for Assessing the Toxic Effects of Some Water Soluble Substances, Based on Changes in Periphyton Community Structure," Water Res., Vol 3, 1963, pp 963-972.
161. Phinney, H. K. and McIntire, C. D., "Effect of Temperature on Metabolism of Periphyton Communities Developed in Laboratory Streams," Limno. and Oceanogr., Vol 10, No. 3, 1965, pp 341-344.
162. Sperling, J. A. and Grumewald, R., "Batch Culturing of Thermophilic Benthic Algae and Phosphorus Uptake in a Laboratory Stream Model," Limnol. & Oceanog., Vol 14, No. 6, 1969, pp 944-949.
163. Natarajan, K. V., "Toxicity of Ammonia to Marine Diatoms," J. Wat. Poll. Contr. Fed., Vol 42, No. 5, 1970, pp 184-190.
164. Prakash, A. and Rashid, M. A., "Influence of Humic Substances on the Growth of Marine Phytoplankton: Dinoflagellates," Limnol. and Oceanog., Vol 13, No. 4, 1968, pp 598-606.
165. Pratt, D. M., "Competition Between Skeletonerna costatum and Olithodiscus luteus in Narragansett Bay and in Culture," Limnol. and Oceanog., Vol 11, No. 4, 1966, pp 447-455.
166. Kornmann, P., "Benthonic Algae, Conservation and Culture in the Laboratory," Research Methods in Marine Biology, Carl Schlieper, Editor, U. Washington Press. 1972, 356 p.
167. Edwards, P., "Cultured Red Alga to Measure Pollution," Mar. Poll. Bull., Vol 3, No. 12, 1972, pp 184-187.
168. Pybus, C., "Effects of Anionic Detergent on the Growth of Laminaria," Mar. Poll. Bull., Vol 4, No. 5, 1973, pp 73-77.
169. Hauenschild, C., "Invertebrates Conservation and Culture in the Laboratory," Research Methods in Marine Biology, Carl Schlieper, Editor, U. Washington Press, 1972, 356 p.
170. Karbe, L., "Marine Hydroids as Test Organisms for Assessing the Toxicity of Water Pollutants. The Effect of Heavy Metals on Colonies of Eirene viridula," Marine Biology, Vol 12, 1972, pp 316-328.
171. Calder, D. R., "Development of the Sea Nettle Chrysaora quinquecirrha," Chesa. Sci., Vol 13, No. 1, 1972, pp 40-44.
172. Sparrgenbert, D. B., "Cultivation of the Life Stages of Aurelia aurita Under Controlled Conditions," J. Exp. Zool., Vol 159, No. 3, 1965, pp 303-318.
173. Roosen-Runge, E. C., "Life Cycle of the Hydromedusae Phialidium gregarium in the Laboratory," Biol. Bull., Vol 139, 1970, pp 203-221.
174. Jebram, D., "A Cultivation Method for Saltwater Biyozoa and an Example for Experimental Biology," Proc. 1st Intl. Bryozool. Assoc. Conf. Milan, Aug 12-16, 1968, pp 119-128.
175. Dudley, J. E., "Observations on the Reproduction Early Larval Development, and Colony Astogery of Conoperium tenuissimum," Chesa. Sci., Vol 14, No. 4, 1973, pp 270-278.

176. Wood, T. S., "Laboratory Culture of Fresh Water Ectoprocta," Trans. Amer. Micros. Soc., Vol 90, No. 1, 1971, pp 92-94.
177. Aloia, R. C. and Moretti, R. L., "Sterile Culture Technique for Species of the Rotifer Asplanadina," Trans. Amer. Micros. Soc., Vol 92, No. 3, 1973, pp 364-371.
178. Gilbert, J. J., "Monoxenic Cultivation of the Rotifer Brachionus calyciflorus in a Defined Medium," Oecologia (Berl.), Vol 4, 1970, pp 89-101.
179. Reish, D. J., "The Effects of Varying Concentrations of Nutrients, Chlorinity, and Dissolved Oxygen on Polychaetous Annelids," Water Res., Vol 4, 1970, pp 721-735.
180. Howie, D. I. D., "Dried Organic Substances as Food for Larval Annelids," Nature, Vol 181, 1958, pp 1486-1497.
181. Loosanoff, V. L. and Davis, H. C., "Rearing of Bivalve Mollusks," Advances In Marine Biology, Vol 1, 1963, 136 p.
182. Renzoni, A., "Influence of Crude Oil, Derivatives and Dispersants on Larvae," Mar. Poll. Bull., Vol 4, No. 1, 1973, pp 9-13.
183. Calabrese, A., Collier, R. S., Nelson, D. A. and MacInnes, J. R., "The Toxicity of Heavy Metals to Embryos of the American Oyster Crassostrea virginica," Mar. Biol., Vol 18, 1973, pp 162-166.
184. Woelke, C. E., "Development of a Receiving Water Quality Bioassay Criterion Based on the 48-Hour Pacific Oyster (Crassostrea gigas) Embryo," Washington Dept. of Fisheries Tech., Rpt. 9, 1972, 93 p.
185. Helm, M. M., Hollared, D. L. and Stephenson, R. R., "The Effect of Supplementary Algal Feeding of a Hatchery Breeding Stock of Ostrea Edulis on Larval Vigour," J. Marine Biol. Assoc. of United Kingdom, Vol 53, 1973, pp 673-684.
186. Dupuy, J. L. and Rivkin, S., "The Development of Laboratory Technique for the Production of the Oyster, Crassostrea virginica," Chesapeake Science, Vol 13, No. 1, 1972, pp 45-52.
187. Millar, R. H. and Scott, J. N., "Bacteria-Free Culture of Oyster Larvae," Nature, Vol 216, 1967, pp 1139-1140.
188. Collier, A., Ray, S. N., Magnitzky, A. W. and Bell, J. O., "Effect of Dissolved Organic Substances on Oysters," U. S. Fish & Wildlife Service Fishery Bulletin, Vol 84, 1953, pp 167-185.
189. Loosanoff, V. L. and Engle, J. B., "Effect of Different Concentrations of Micro-Organisms on the Feeding of Oysters (c. virginica)," U. S. Fish & Wildlife Service Fishery Bulletin, Vol 23, 1947, pp 31-57.
190. Tenore, K. R. and Dunstan, W. M., "Growth Comparisons of Oysters, Mussels and Scallops Cultivated on Algae Grown with Artificial Medium and Treated Sewage Effluent," Chesa. Sci., Vol 14, No. 1, 1973, pp 64-66.

191. Zuraw, E. A., Leone, D. E. and Grision, C. A., "Ecology of Molluscs and the Culture of Mya arenaria," General Dynamics, Electric Boat Division Informal Report U413, 1967, pp 206-299.
192. Stickney, A. P., "Salinity, Temperature, and Food Requirements of Soft-Shell Clam Larvae in Laboratory Culture," Ecology, Vol 45, 1964, pp 283-291.
193. Young, R. T., "Stimulation of Spawning in the Mussel (Mytilus californianus)," Ecology, Vol 26, No. 1, 1945, pp 58-69.
194. Carriker, M. R. and Van Zandt, D., "Activity of the Marine Gastropod Urosalpinx cinerea in the Absence of Hibernation," Chesapeake Science, Vol 14, No. 4, 1973, pp 285-288.
195. Jones, R. O., "Propagation of Fresh-Water Mussels," Progressive Fish Culturist, Vol 12, 1950, pp 13-25.
196. Dewey, J. E. and Parker, B. L., "Mass Rearing of Daphnia magna for Insecticide Bioassay," J. Econ. Entomology, Vol 6, 1957, pp 821-825.
197. Taub, F. B. and Dollar, A. M., "The Nutritional Inadequacy of Chlorella and Chlamydomonas as Food for Daphnia pulex," Limnology & Oceanography, Vol 13, No. 4, 1968, pp 607-617.
198. McMahon, J. W., "Some Physical Factors Influencing the Feeding Behavior of Daphnia magna Straus.," Canadian J. Zoology, Vol 43, No. 4, 1965, pp 603-611.
199. Schindler, D. W., "Feeding, Assimilation and Respiration Rates of Daphnia magna Under Various Environmental Conditions and Their Relation to Production Estimates," J. Animal Ecology, Vol 37, No. 2, 1968, pp 369-385.
200. Biesinger, K. E., "Unpublished Manuscript. Recommended Bioassay Procedure for Daphnia magna Chronic Tests in a Standing System," EPA Water Quality Lab., Duluth, 3 p.
201. Grosch, D. S., "Reproduction Tests: The Toxicity for Artemia of Derivatives from Nonpersistent Pesticides," Biol. Bull., Vol 145, 1973, pp 340-351.
202. Sargeloos, P. and Persoane, G., "Three Simple Culture Devices for Aquatic Invertebrates and Fish Larvae with Continuous Recirculation of the Medium," Marine Biology, Vol 15, 1972, pp 251-254.
203. Jones, J. C., "A Study of Possible Modifications of the WHO Method for Testing DDT-Resistance in Mosquito Larvae with Special Reference to Anopheles quadrimaculatur (Say)," Bull. WHO, Vol 36, No. 2, 1957, pp 353-356.
204. Maki, A. W., Stewart, K. W. and Silvey, J. K. G., "The Effects of Dibrom on Respiratory Activity of the Stonefly, Hydroperla crosbyi, Hellgrammite, Corydalis cornutus and the Golden Shiner, Notemigonus crysoleucas," Trans. Amer. Fish. Soc., Vol 102, No. 4, 1973, pp 806-815.

205. Hicks, D. B. and DeWitt, J. S., "Effects of Dissolved Oxygen on Kraft Pulp Mill Effluent Toxicity," Water Res., Vol 5, No. 9, 1971, pp 693-701.
206. MacPhee, C. and Ruelle, R., "A Chemical Selectivity Lethal to Squawfish (Ptychocheilus oregonensis and P. umpqua)," Trans. Amer. Fish. Soc., Vol 98, No. 4, 1963, pp 676-684.
207. Wilson, K. W., "Toxicity of Oil-Spill Dispersants to Embryos and Larvae of Some Marine Fish," Marine Pollution and Sea Life, FAO, Fishing News, Ltd., 1972, pp 318-322.
208. Rosenthal, H. and Stelzer, R., "Effects of 2,4- and 2,5-Dinitrophenol on the Embryological Development of the Herring Clupea harengus," Mar. Biol., Vol 5, 1970, pp 325-336.
209. Lasker, R., Freder, H. M., Theilacker, G. H. and May, R. C., "Feeding, Growth, and Survival of Eurgraulis mordax Larvae Reared in the Laboratory," Marine Biology, Vol 5, 1970, pp 345-353.
210. Blaxter, J. H. S., "Experimental Rearing of Pilchard Larvae, Sardina pilchardus," J. Marine Biological Assoc. U. K., Vol 49, 1969, pp 557-575.
211. Porter, P. E., "Flatfish Culture," Research in Fisheries, Contr. No. 375, College of Fisheries, U. of Washington, 1973, 100 p.
212. Shelbourne, J. E., "The Artificial Propagation of Marine Fish," Advances in Marine Biology, Vol 2, 1964, pp 1-83.
213. Mason, J. C. and Fessler, J. L., "A Simple Apparatus for the Incubation of Almonid Embryos at Controlled Levels of Temperature, Water Flow, and Dissolved Oxygen," Progressive Fish Culturist, Vol 28, No. 3, 1966, pp 171-174.
214. Allison, L. N., "Delay of Spawning in Eastern Brook Trout by Means of Artificially Prolonged Light Intervals," Progressive Fish Culturist, Vol 13, 1951, pp 111-116.
215. Hokanson, K. E. F., McCormick, J. H., Jones, B. R. and Tucker, J. H., "Thermal Requirements for Maturation, Spawning, and Embryo Survival of the Brook Trout, Salvelinus fontinalis," J. Fisheries Res. Bd. Can., Vol 30, 1973, pp 975-984.
216. Hale, J. G. and Hilden, D. A., "Spawning and Some Aspects of Early Life History of Brook Trout, Salvelinus fontinalis (Mitchell), in the Laboratory," Trans. Amer. Fisheries Society, Vol 98, No. 3, 1969, pp 473-477.
217. EPA, 1972, "Recommended Bioassay Procedure for Brook Trout," Salvelinus fontinalis (Mitchell) Partial Chronic Tests," U. S. EPA, Duluth, Minn. Manuscript, April, 1971. Revised, January, 1972, 12 p.
218. McCouley, R. W. and Trimbom, F., "Incubating Rainbow Trout Eggs in Heated, Recirculated Water," Progressive Fish Culturist, Vol 30, 1968, 64 p.

219. Poon, D. C. and Johnson, A. K., "The Effect of Delayed Fertilization on Transported Salmon Eggs," Progressive Fish Culturist, Vol 32, 1970, pp 81-84.
220. Scott, K. R. and Gillespie, D. C., "A Compact Recirculation Unit for the Rearing and Maintenance of Fish," Jo. Fisheries Res. Bd. Can., Vol 29, 1972, pp 1071-1074.
221. Burrows, R. E. and Combs, B. D., "Controlled Environments for Salmon Propagation," Progressive Fish Culturist, Vol 30, 1968, pp 123-136.
222. Burrows, R. E. and Chenoweth, H. W., "The Rectangular Circulating Rearing Pond," Progressive Fish Culturist, Vol 32, 1970, pp 67-80.
223. Brown, M. E., "The Growth of Brown Trout (Salmo trutta Linn.) IV. The Effect of Food and Temperature on the Survival and Growth of Fish," Jo. Experimental Biology, Vol 28, 1951, pp 473-491.
224. Fowler, L. G., McCormick, J. H., Jr., and Thomas, A. E., "Studies of Caloric and Vitamin Levels of Salmon Diets," U. S. Bureau of Sport Fisheries of Wildlife Technical Papers, No. 6, 1966, 14 p.
225. Cuerrier, J. P., Keith, J. A., and Stone, E., "Problems with DOT in Fish Culture Operations," Naturaliste Canadienne, Vol 94, 1967, pp 315-320.
226. Wallach, D., "Management and Medical Care of Goldfish," J.A.Y.M.A., Vol 159(5), 1971, pp 583-595.
227. Fostes, N. R., Cairns, J., Jr., and Kaesler, R. L., "The Flagfish, Jordanella floridae, as a Laboratory Animal for Behavioral Studies," 1969.
228. EPA, Manuscript, "Recommended Bioassay Procedure for Jordanella floridae (Goode and Bean) Chronic Tests," EPA, Duluth, Minn. Manuscript, 9 p.
229. Crandall, C. A. and Goodnight, C. J., "The Effects of Sublethal Concentrations of Several Toxicants to the Common Guppy, Lebistes reticulatus," Trans. Amer. Micros. Soc., Vol 82, 1963, pp 59-73.
230. Crandall, C. A. and Goodnight, C. J., "Effects of Sublethal Concentrations of Several Toxicants on Growth of the Common Guppy, Lebistes reticulatus," Limnol. and Oceanogr., Vol 7, 1962, pp 233-239.
231. Larimore, R. W., "Ecological Life History of the Warmouth (Centrarchidae)," Bulletin Illinois Natural History Survey, Vol 27, Article No. 1, 1957.
232. EPA, 1972, "Recommended Bioassay Procedure for Bluegill Lepomis macrochirus (Rafinesque) Partial Chronic Tests," U. S. EPA, Duluth, Minn. Manuscript, 11 p.
233. Benoit, D., "Nutrition Study for Maintaining Yellow Perch Fingerlings in the Laboratory," Progressive Fish Culturist, Vol 30, 1968, 234 p.

234. Hale, J. G., "White Sucker Spawning and Culture of the Young in the Laboratory," Progressive Fish Culturist, Vol 32, 1970, 169 p.
235. Blaylock, B. G. and Griffith, H. A., "A Laboratory Technique for Spawning Carp," Progressive Fish Culturist, Vol 33, 1971, pp 48-50.
236. Adelman, I. R. and Smith, L. L., "Effect of Oxygen on Growth and Food Conversion Efficiency of Northern Pike," Progressive Fish Culturist, Vol 32, 1970, pp 93-96.
237. Kester, D. R., Duedall, I. W., Connors, D. H., and Pytkowicz, R. M., "Preparation of Artificial Seawater," Limnol. and Oceanog., Vol 12(1), 1967, pp 176-179.
238. White, D. B., Stickney, R. R., Miller, D., and Knight, L. H., "Seawater System for Aquaculture of Estuarine Organisms at the Skidaway Institute of Oceanography," Georgia Marine Science Center, U. Georgia, Tech. Report Series No. 73-1-18, 1973.
239. Wood, L., "A Controlled Conditions System (CCS) for Continuously Flowing Seawater," Limnol. and Oceanog., Vol 10(3), 1965, pp 475-477.
240. Hettlen, W. F., Jr., Lichtenheld, R. W., and Gordy, H. R., "Open Seawater System with Controlled Temperature and Salinity," Progressive Fish Culturist, Vol 33, 1971, pp 3-11.
241. Parisot, T. J., "A Closed Recirculated Seawater System," Progressive Fish Culturist, Vol 29(3), 1967, pp 133-139.
242. Sudia, W. D., "A Device for Raising Animals Requiring a Flowing Water Environment," Ohio Jo. Science, Vol 51(4), 1951, pp 197-202.
243. Whitford, L. A. and Dillard, G. E., "An Artificial Stream Apparatus for the Study of Lotic Organisms," Limnol. and Oceanog., Vol 9(4), 1964, pp 598-600.
244. Sprague, J. B., "Avoidance of Copper-Zinc Solutions by Young Salmon in the Laboratory," Journ. Water Poll. Contr. Fed., Vol 36, No. 8, 1964, pp 930-1004.
245. Irwin, W. H., "Fifty-Seven Species of Fish in Oil-Refinery Waste Bioassay," Trans. North American Wildlife Conf., Vol 30, 1965, pp 89-99.
246. Cope, O. B., "Contamination of the Freshwater Ecosystem by Pesticides," J. Applied Ecology 3, (Supplement), 1966, pp 33-44.
247. Lennon, R. E., "Selected Strains of Fish as Bioassay Animals," Progressive Fish Culturist, Vol 29, 1967, pp 129-132.
248. Henderson, C. and Pickering, Q. H., "Use of Fish in the Detection of Contaminants in Water Supplies," Jo. American Water Works Assoc., Vol 55(6), 1963, pp 715-720.
249. Lennon, R. E. and Walker, C. R., "Investigations in Fish Control. Laboratories and Methods for Screening Fish-Control Chemicals," Bureau Sport Fisheries and Wildlife, Circular 185, 1964, 15 p.

250. Tarzwell, C. M., "Measurements of Pollution Effects on Living Organisms. Bioassays to Determine Allowable Waste Concentrations in the Aquatic Environment," Proc. Royal Society London B., Vol 177, 1971, pp 279-285.
251. Anonymous, "Report to the Congress - Observations on Dredging Activities and Problems," 1972, U. S. General Accounting Office.
252. Sherk, A. J., Jr., "Current Status of the Knowledge of the Biological Effects of Suspended and Deposited Sediments in Chesapeake Bay," Chesapeake Science, Vol 13 (Supplement), 1972, S137-S144.
253. Anonymous, "Environmental Investigations of Dredging Activity in Mobile Bay, Alabama," Final Report of the Technical Committee for Analysis of Mobile Bay Dredging, U. S. Army Corps of Engineers, Mobile District, Alabama, 1973, 53 p.
254. Saila, S. B., Pratt, S. D., and Polgor, T. T., "Dredge Spoil Disposal in Rhode Island Sound," Marine Technical Report No. 2, University of Rhode Island, 1972, 48 p.
255. Anonymous, "Research to Determine the Environmental Response to the Deposition of Spoil on Salt Marshes using Diked and Undiked Techniques--Second Annual Progress Report," Skidaway Institute of Oceanography, Georgia, 1973, 189 p.
256. Anonymous, "Base Line Environmental Investigation Dredging Activities, Mobile Bay, Alabama, a Progress Report of the Technical Committee." Prepared for the Conferees of the January 27-28, 1970 Conference, In the Matter of Pollution of the Navigable Waters of Mobile Bay and Its Tributaries, 1972, 72 p.
257. Price, K. C., "Environmental Impact of Dead Reef Shell Dredging in Mobile Bay, Alabama," Gulf South Research Institute, La., 1972, 72 p.
258. Mackin, J. G., "Canal Dredging and Silting in Louisiana Bays," Institute of Marine Science (Texas), Vol 7, 1961, pp 262-314.
259. Flemer, D. A., et al., "Biological Effects of Spoil Disposal in Chesapeake Bay," Journal of the Sanitary Engineering Division, Proceedings of the American Society of Civil Engineering, 1968.
260. Pratt, S. D., et al., "Biological Effects of Ocean Disposal of Solid Waste," Marine Experiment Station. Marine Technical Report Series No. 9, University of Rhode Island, 1973, 53 p.
261. Pierce, H. D., "Inland Lake Dredging Evaluation," Technical Bulletin No. 46, Department of Natural Resources, Madison, Wisc., 1970.
262. Hollis, E. H., et al., "A Literature Review of the Effects of Turbidity and Siltation on Aquatic Life," (Mimeograph), 1964.
263. Sherk, J. A., O'Connor, J. M., and Neumann, D. A., "Effects of Suspended and Deposited Sediments on Estuarine Organisms-Phase II," Chesapeake Biological Laboratory, Reference No. 72-9E, University of Maryland, 1972.

264. Cairns, J., Jr., "Suspended Solids Standards for the Protection of Aquatic Organisms," Purdue University Engineering Bulletin, No. 129(1), 1968, pp 16-27.
265. Rogers, B. A., "Tolerance Levels of Four Species of Estuarine Fishes to Suspended Mineral Solids," M. S. Thesis, University of Rhode Island, Kingston, R. I., 1969.
266. Alabaster, J. S., "Suspended Solids and Fish," Proceedings of the Royal Society of London, Vol 180, 1972, pp 395-406.
267. Murphy, G. I., "Effect of Mixing Depth and Turbidity on the Productivity of Fresh-Water Impoundments," Transactions of the American Fisheries Society, Vol 91, 1962, pp 69-76.
268. Jones, D. and Wills, M. S., "The Attenuation of Light in Sea and Estuarine Waters in Relation to the Concentration of Suspended Solid Matter," Journal of the Marine Biological Association, U. K., Vol 35, 1965, pp 431-444.
269. Schinck, H., Jr., and Davis, A., "A Turbidity Survey of Narragansett Bay," Ocean Engineering, Vol 2, 1963, pp 169-178.
270. Stickney, R. R. and Miller, D., "Chemical and Biological Survey of the Savannah River Adjacent to Elba Island," Technical Report Series No. 73-3, Georgia Marine Science Center, Ga., 1973, 68 p.
271. Biggs, R. B. 1968. "Environmental Effects of Overboard Spoil Disposal," Journal of the Sanitary Engineering Division, Proceedings of the American Society of Civil Engineering. 94:(SA3):447-478.
272. Cronin, L. E. et al. 1970. "Gross Physical and Biological Effects of Overboard Spoil Disposal in Upper Chesapeake Bay." NRI Special Report No. 3, National Resources Institute, University of Maryland. 66 p.
273. May, E. B. 1973. "Environmental Effects of Hydraulic Dredging in Estuaries." Alabama Marine Resources Laboratory, Alabama. pp
274. Gilderhus, P. A., "Some Effects of Sublethal Concentrations of Sodium-Arsenite on Bluegills and the Aquatic Environment," Trans. Amer. Fish. Soc., Vol 95, No. 3, 1966, pp 289-296.
275. Weiss, C. M., "Physiological Effect of Organic Phosphorus Insecticides on Several Species of Fish," Trans. Amer. Fish. Soc., Vol 90, No. 2, 1961, pp 143-152.
276. LeGore, R. S. and DesVoigne, D. M., "Absence of Acute Effects on Threespine Sticklebacks (Gasterosteus aculeatus) and Coho Salmon (Oncorhynchus kistutch) Exposed to Resuspended Harbor Sediment Contaminants," J. Fish. Res. Bd. Can., Vol 39, No. 8, 1973, pp 1240-1242.
277. Gannon, J. E. and A. N. Beeton, 1969. "Studies on the Effects of Dredged Materials from Selected Great Lakes Harbors on Plankton and Benthos," Special Report No. 8, Center for Great Lakes Studies, University of Wisconsin, Milwaukee. 82 p.



278. Cherry, D. S., K. L. Dickson, and J. Cairns, Jr., 1974. "The Use of a Mobile Laboratory to Study Temperature Response of Fish." Purdue University of Engineering Bulletin.
279. Gannon, J. E., and A. M. Beeton. 1971. Procedures for determining the effects of dredged sediments on biota-benthos viability and sediment selectivity tests. Journal of Water Pollution Control Federation. Vol. 43 (pt. 1), pp. 392-398.
280. Smith, L. L., and D. M. Oseid. 1972. The effects of hydrogen sulfide on fish eggs and fry. Water Research 6:711-720.
281. Hokenson, K. E. F., and L. L. Smith, Jr. 1971. Some factors influencing toxicity of linear alkylate sulfonate to the bluegill. Trans. Amer. Fish. Soc. 100(1):1-12.
282. Sykora, J. L., E. J. Smith, and M. Synak. 1972. Effects of lime neutralized iron hydroxide suspensions on juvenile brook trout (Salvelinus fontinalis, Mitchill). Water Research 6:933-950.
283. Bryan, G. W., and L. G. Hummerstone. 1971. Adaptations of the polychaete Nereis diversicolor to estuarine sediments containing high concentrations of heavy metals. I. General Observations and adaptation to copper. J. Mar. Biol. Assoc., U. K. 51:845-863.
284. Irwin, J. C., and L. H. Karstace. 1972. The toxicity for ducks of disintegrated lead shot in a simulated marsh environment. J. Wildlife Diseases 8:149-154.
285. Vaccaro, R. F., G. D. Grice, G. T. Rowe, and P. H. Wiele. 1972. Acid-iron waste disposal and the summer distribution of standing crops in the New York Bight. Water Research 6(3):231-256.
286. Eisler, R., G. R. Gardner, R. J. Henneky, G. LaRode, D. F. Walsh, and P. P. Yevich. 1973. Acute toxicology of sodium nitrilotriacetic acid (NTA) and NTA containing detergents to marine organisms. Water Resources 6:1009-1027.
287. Kawatski, J. A., and J. C. Schmulback. 1971. Toxicities of aldrin and Dieldrin to the fresh-water ostracod Chlamydotheca arcata. J. Econ. Entomol. 64(5):182-185.
288. Wilbur, R. L., and E. W. Whitney. 1973. Toxicity of the herbicide Kuron (Silvex) to bluegill eggs and fry. Trans. American. Fish. Soc. 102(3):630-633.
289. Wildish, D. J. 1972. Acute toxicity of polyoxyethylene esters and polyoxyethylene ethers to S. salar and G. oceanicus. Wat. Res. 6:759-762.
290. Earnest, R. D., and P. E. Benville, Jr. 1972. Acute toxicity of four organochlorine insecticides to two species of surf perch. Calif. Fish and Game 58(2):127-132.
291. Eisler, R., and G. La Roche. 1972. Elemental composition of the estuarine teleost Fundulus heteroclitus (L). J. Exp. Mar. Biol. 9:29-42.

292. Gillespie, D. C. 1972. Mobilization of mercury from sediments into guppies (Poecilia reticulata). Fisheries Research Board of Canada Journal. 29:1035-1041.
293. Gillespie, D. C., and D. P. Scott. 1971. Mobilization of mercuric sulphide from sediment into fish under aerobic conditions. Journal of the Fisheries Research Board of Canada. 28(11):1807-1808.
294. Rose, C. D. 1973. Mortality of market sized oysters (Crassostrea virginica) in the vicinity of a dredging operation. Chesapeake Sciences 14(2):135-138.
295. Thomas, P. M., and R. O. Legault. 1967. The effects of industrial wastes from Charmin Paper Products Company on fish of the Cheboggan River drainage system. Water Resources 1:217-229.
296. Ferguson, D. E., and C. R. Bingham. 1966. Endrin resistance in the yellow bullhead (Ictalurus natalis). Trans. Amer. Fish. Soc. 95(3):325-326.
297. Sanders, H. O., and O. B. Cope. 1966. Toxicities of several pesticides to two species of cladocerans. Trans. Amer. Fish. Soc. 95(2):165-169.
298. Mount, D. I., and C. E. Stephan. 1967. A method for establishing acceptable toxicant limits for fish--malathion and butoxyethanol ester of 2,4-D. Trans. Amer. Fish. Soc. 96(4):185-193.
299. Hiltibran, R. C. 1967. Effects of some herbicides on fertilized fish, eggs, and fry. Trans. Amer. Fish. Soc. 96(4):414-416.
300. Bonn, E. W., and B. J. Folis. 1967. Effects of hydrogen sulfide on channel catfish (Ictalurus punctatus). Trans. Amer. Fish. Soc. 96(1):31-36.
301. Stewart, N. E., R. E. Millerman, and W. P. Breese. 1967. Acute toxicity of the insecticide Sevin and its hydrolitic product 1-naphthol to some marine organisms. Trans. Amer. Fish. Soc. 96(1):25-30.
302. Smith, A. J. 1967. The effects of the lamphrey larvicide, 3-trifluoromethyl-4-nitrophenol, on selected aquatic invertebrates. Trans. Amer. Fish. Soc. 96(4):410-413.
303. Herr, F., E. Greselin, and C. Chappel. 1967. Toxicology studies of antimycin, a fish eradicant. Trans. Amer. Fish. Soc. 96(3):320-326.
304. Marking, L. L., and V. K. Dawson. 1972. The half life of biological activity of antimycin determined by fish bioassay. Trans. Amer. Fish. Soc. 101(1):100-105.
305. Wade, R. A. 1969. Ecology of juvenile tarpon and effects of dieldrin on two associated species. U. S. Department of Interior, Bureau of Sport Fisheries and Wildlife. Tech. Paper #41.

306. Kramer, R. H., and L. L. Smith. 1965. Effects of suspended wood fiber on brown and rainbow trout eggs and alevins. Trans. Amer. Fish. Soc. 94(3):252-258.
307. Smith, L. L., and R. H. Kramer. 1965. Survival of walleye fingerlings in conifer ground wood fibers. Trans. Amer. Fish. Soc. 94(4):402-404.
308. Carlson, C. A. 1966. Effects of three organophosphorus insecticides on immature Hexagenia and Hydropsyche of the upper Mississippi River. Trans. Amer. Fish. Soc. 95(1):1-5.
309. Doudoroff, P., G. Leduc, and C. R. Schneider. 1966. Acute toxicity to fish of solutions containing complex metal cyanides in relation to concentrations of molecular hydroganic acids. Trans. Amer. Fish. Soc. 95(1):6-22.
310. Smith, L. L., R. H. Kramer, and D. M. Oseid. 1966. Long term effects of conifer groundwood paper fiber on walleye. Trans. Amer. Fish. Soc. 95(1):60-70.
311. Croker, R. A., and A. J. Wilson. 1965. Kinetics and effects of DDT in a tidal marsh ditch. Trans. Amer. Fish. Soc. 94(2):152-159.
312. Andrews, A. K., C. C. Van Valin, and B. E. Stebbings. 1966. Some effects of heptachlor on bluegills (Lepomis macrochirus). Trans. Amer. Fish. Soc. 95(3):297-309.
313. Gilderhus, P. A. 1966. Some effects of sublethal concentrations of sodium arsenate on bluegills and the aquatic environment. Trans. Amer. Fish. Soc. 95(3):289-296.
314. Hollano, H. T., D. L. Coppage, and P. A. Butler. 1966. Increased sensitivity to pesticides in sheepshead minnows. Trans. Amer. Fish. Soc. 95(3):289-296.
315. Eisler, R., and P. H. Edmunds. 1966. Effects of endrin on blood and tissues chemistry of a marine fish. Trans. Amer. Fish. Soc. 95(2):153-159.
316. Hicks, D. B., and J. W. DeWitt. 1971. Effects of dissolved oxygen on Kraft pulp mill effluent toxicity. Water. Res. 5(9):693-701.
317. MacLeod, J. C., and L. L. Smith, Jr. 1966. Effect of pulp wood fiber on oxygen consumption and swimming endurance of the fathead minnow (Pimephales promelas). Trans. Amer. Fish. Soc. 95(1):71-84.
318. Freeman, R. A., and W. H. Everhart. 1971. Toxicity of aluminum hydroxide complexes in neutral and basic media to rainbow trout. Trans. Amer. Fish. Soc. 100(4):644-658.
319. Lincer, J. L., J. M. Solon, and J. H. Nair III. 1970. DDT and endrin fish toxicity under static versus dynamic bioassay conditions. Trans. Amer. Fish. Soc. 100(1):13-19.

320. Cope, O. B., E. M. Wood, and G. H. Waller. 1970. Some chronic effects of 2,4-D on the bluegill (Lepomis macrochirus). Trans. Amer. Fish. Soc. 99(1):1-12.
321. Macek, K. J., and W. A. McAllister. 1970. Insecticide susceptibility of some common fish family representatives. Trans. Amer. Fish. Soc. 99(1):20-27.
322. Lane, C. E., and R. J. Livingston. 1970. Some acute and chronic effects of dieldrin on the sailfin molly (Poecilia latipinna). Trans. Amer. Fish. Soc. 99(3):489-495.
323. Sanders, H. O. 1969. Toxicity of pesticides to the crustacean Gammarus lacustris. U. S. Dept. of Interior, Bureau of Sport Fisheries & Wildlife, Tech. Paper #25.
324. Meyer, F. P. 1965. The experimental use of Guthion as a selective fish eradicator. Trans. Amer. Fish. Soc. 94(3):203-209.
325. Ferguson, D. E., J. L. Ludke, and G. C. Murphy. 1966. Dynamics of endrin uptake and release by resistant and susceptible strains of the mosquitofish. Trans. Amer. Fish. Soc. 95(4):335-344.
326. Neuhold, J. M., and W. F. Sigler. 1960. Effects of sodium fluoride on carp and rainbow trout. Trans. Amer. Fish. Soc. 89(4):358-370.
327. Gaufin, A. R., L. Jensen, and T. Nelson. 1961. Bioassays determine pesticide toxicity to aquatic invertebrates. Water and Sewage Works: Sept. 1961.
328. Van Valin, C. C., A. K. Andrews, and L. L. Eller. 1968. Some effects of Mirex on two warm-water fishes. Trans. Amer. Fish. Soc. 97(2):185-196.
329. McKim, J. M., G. M. Christensen, and E. P. Hunt. 1970. Changes in the blood of brook trout (Salvelinus fontinalis) after short-term and long-term exposure to copper. J. Fish. Res. Bd. Canada 27(10):1883-1889.
330. Eisler, R., and D. G. Duell. 1965. Acute toxicity of soaps to estuarine fishes. Prog. Fish-Cult. 27(1):45-48.
331. Pickering, Q. H., and W. N. Vigor. 1965. The acute toxicity of zinc to eggs and fry of the fathead minnow. Prog. Fish. Cult. 27(3):153-157.
332. Kramer, R. H., and L. L. Smith. 1966. Survival of walleye eggs in suspended wood fibers. Prog. Fish-Cult. 28(2):79-82.
333. Eisler, R. 1966. Effects of apholate, an insect sterilant, on an estuarine fish, shrimp, and gastropod. Prog. Fish-Cult. 28(3):154-158.
334. Helms, D. R. 1967. Use of formalin for selective control of tadpoles in the presence of fish. Prog. Fish-Cult. 29(1):43-47.

335. Adelman, I. R., and L. L. Smith. 1972. Toxicity of hydrogen sulfide to goldfish (Carassius auratus) as influenced by temperature, oxygen, and bioassay techniques. J. Fish. Res. Bd. Canada 29(9): 1309-1317.
336. Cairns, J. Jr., T. K. Bahns, D. T. Burton, K. L. Dickson, R. E. Sparks, and W. T. Waller. 1972. The effects of pH, solubility, and temperature upon the acute toxicity of zinc to the bluegill sunfish (Lepomis macrochirus Raf.). Trans. Kansas Acad. Sci. 74(1):81-92.
337. Matton, P., and Q. N. Lattam. 1969. Effect of the organophosphate Dylox on rainbow trout larvae. J. Fish. Res. Bd. Canada 26(8): 2193-2200.
338. Dorfman, D., and W. R. Whitworth. 1969. Effects of fluctuations of lead, temperature, and dissolved oxygen on the growth of brook trout. J. Fish. Res. Bd. Canada 26(9):2495-2501.
339. Culley, D. D., and D. E. Ferguson. 1968. Patterns of insecticide resistance in the mosquitofish, Gambusia affinis. J. Fish. Res. Bd. Canada 26(9):2395-2401.
340. Buchanan, D. V., R. E. Milleman, and N. E. Stewart. 1970. Effects of the insecticide Sevin on various stages of the dungeness crab, Cancer magister. J. Fish. Res. Bd. Canada 27(1):93-104.
341. Zitko, V., D. E. Aiken, S. N. Tibbo, K. W. T. Besch, and J. M. Anderson. 1970. Toxicity of yellow phosphorus to herring (Clupea harengus), Atlantic salmon (Salmo salar), lobster (Homarus americanus), and beach flea (Gammarus oceanicus). J. Fish. Res. Bd. Canada. 27(1):21-29.
342. Marking, L. L. 1970. Juglone (5-hydroxy-1,4-naphthoquinone) as a fish toxicant. Trans. Amer. Fish. Soc. 99(3):510-514.
343. Naqui, S. M., and Denzel E. Ferguson. 1970. Levels of pesticide resistance in the fresh water shrimp, Palaemonetes kadiakeesis. Trans. Amer. Fish. Soc. 99(4):696-699.
344. Butler, J. A., R. E. Milleman, and N. E. Stewart. 1968. Effects of the insecticide Sevin on survival and growth of the cockle clam, Clinocardium nuttalli. J. Fish. Res. Bd. Canada 25(8): 1621-1635.
345. Macek, K. J. 1968. Reproduction in brook Trout (Salvelinus fontinalis), fed sublethal concentrations of DDT. J. Fish. Res. Bd. Canada 25(9):1787-1796.
346. Adelman, I. R., and L. L. Smith. 1970. Effect on hydrogen sulfide on Northern Pride eggs and sac fry. Trans. Amer. Fish. Soc. 99(3):501-509.
347. Fletcher, G. L., R. J. Hoyle, and D. A. Horne. 1970. Yellow phosphorus pollution: it's toxicity to seawater maintained brook trout (Salvelinus fontinalis), and smelt (Osmerus mordax). J. Fish. Res. Bd. Canada 27:1379-1384.

348. Arthur, J. W., and E. N. Leonard. 1970. Effects of copper on Gammarus pseudolimnaeus, Physa integra, and Campeloma decisum in soft water. J. Fish. Res. Bd. Canada 27:1277-1283.
349. Smith, L. L., Jr., and D. M. Oseid. 1970. Toxic effects of hydrogen sulfide to juvenile fish and fish eggs. 25th Industrial Waste Conference, May 5-7, 1970, Engineering Extension Series No. 137, Purdue Univ.
350. Biesinger, K. E., and G. M. Christenson. 1972. Effects of various metals on survival, growth, and reproduction of Daphnia magna. J. Fish. Res. Bd. of Canada 29(12):1691-1700.
351. Katz, M., and G. C. Chadwick. 1961. Toxicity of endrin to some Pacific Northwest fishes. Trans. Amer. Fish. Soc. 90(4):394-397.
352. Applegate, V. C., and E. L. King, Jr. 1962. Comparative toxicity of 3-trifluoromethyl-4-nitrophenol (TFM) to larval lampreys and eleven species of fish. Trans. Amer. Fish. Soc. 91(4):342-345.
353. Gilderhus, P. A. 1967. Effects of diquat on bluegills and their food organisms. Prog. Fish. Cult. 29(2):67-74.
354. Pickering, Q. H., and A. E. Lemke. 1962. The toxicity of organo-phosphorus insecticides to different species of warm-water fish. Trans. Amer. Fish. Soc. 91(2):175-184.
355. Lowe, J. I. 1964. Chronic exposure of spot (Leiostomus xanthurus), to sublethal concentrations of toxaphene in sea water. Trans. Amer. Fish. Soc. 93(4):396-399.
356. Katz, M. 1961. Acute toxicity of some organic insecticides to three species of salmonids and to the three-spine stickleback. Trans. Amer. Fish. Soc. 90(3):264-268.
357. Surber, E. W., and Q. H. Pickering. 1962. Acute toxicity of endothal, diquat, hyamine, dalapon and silvex to fish. Prog. Fish. Cult. 24(4):164-171.
358. Weiss, C. M. 1961. Physiological effect of organic phosphorus insecticides on several species of fish. Trans. Amer. Fish. Soc. 90(2):143-152.
359. Colby, P. J., and L. L. Smith. 1967. Survival of walleye eggs and fry on paper fiber sludge deposits in Rainy River, Minnesota. Trans. Amer. Fish. Soc. 96(3):278-296.
360. Muncy, R. J., and A. D. Oliver, Jr. 1963. Toxicity of ten insecticides to the red crawfish Procambarus clarki (Girard). Trans. Amer. Fish. Soc. 92(4):428-431.
361. Sreenivasan, A., and M. V. Natarajan. 1962. Use of endrin in fishery management. Prog. Fish-Cult. 24(4):181.
362. Lemke, A. E., and D. I. Mount. 1963. Some effects of alkyl benzene sulfonate on the bluegill, Lepomis macrochirus. Trans. Amer. Fish. Soc. 92(4):372-378.

363. Workman, G. W., and J. M. Neuhold. 1963. Lethal concentrations of toxophene for goldfish, mosquitofish, and rainbow trout with notes on detoxification. *Prog. Fish-Cult.* 25(1):23-30.
364. Surber, E. W., and T. O. Thatcher. 1963. Laboratory studies of the effects of alkyl benzene sulfonate (ABS) on aquatic invertebrates. *Trans. Amer. Fish. Soc.* 92(2):152-160.
365. Oseid, D., and L. L. Smith, Jr. 1972. Swimming endurance and resistance to copper and melathion of bluegills treated by long-term exposure to sublethal levels of hydrogen sulfide. *Trans. Amer. Fish. Soc.* 101(4):620-625.
366. Mount, D. I. 1964. An autopsy technique for zinc-caused fish mortality. *Trans. Amer. Fish. Soc.* 93(2):174-182.
367. Ludke, J. L., D. E. Ferguson, and W. D. Burke. 1968. Some endrin relationships in resistant and susceptible populations of golden shiners (Notemigonus chryssoleucas). *Trans. Amer. Fish. Soc.* 97(3):260-263.
368. Coppage, D. L. 1972. Organophosphate pesticides: specific level of brain AChE inhibition related to death in sheepshead minnows. *Trans. Amer. Fish. Soc.* 101(3):534-536.
369. Macek, K. J., D. F. Walsh, J. W. Hogan, and D. D. Holz. 1972. Toxicity of the insecticide Darsban to fish and aquatic invertebrates in ponds. *Trans. American Fish. Society.* 101(3):420-427.
370. Burton, D. T., E. L. Morgan, and J. Cairns, Jr. 1972. Mortality curves of bluegills (Lepomis macrochirus Rafinesque) simultaneously exposed to temperature and zinc stress. *Trans. American Fish Society.* 101(3):435-441.
371. Anderson, J. M. 1968. Effect of sublethal DDT on the lateral line of brook trout, Salvelinus fontinalis. *J. Fish. Res. Bd. Canada* 25(12):2677-2882.
372. MacPhee, C., and R. Ruelle. 1969. A chemical selectively lethal to squawfish (Phychocheilus oregonensis and P. umpqua). *Trans. American Fish. Soc.* 98(4):676-684.
373. Amend, D. F., W. T. Yasutake, and R. Morgan. 1969. Some factors influencing susceptibility of rainbow trout to the acute toxicity of an ethyl mercury phosphate formulation (Timsan). *Trans. Amer. Fish. Soc.* 98(3):419-425.
374. Hatfield, C. T., and P. H. Johansen. 1972. Effects of four insecticides on the ability of Atlantic salmon parr (Salmo salar) to learn and retain a simple conditioned response. *J. Fish. Res. Bd. Can* 29(3):315-321.
375. Mulla, M. S., J. St. Amant, and L. D. Anderson. 1967. Evaluation of organic pesticides for possible use as fish toxicants. *Prog. Fish-Cult.* 29(1):36-42.

376. Carlson, A. R. 1971. Effects of long-term exposure to Carbaryl (Sevin) on survival, growth, and reproduction of the fathead minnow (Pimephales promelas). J. Fish. Res. Bd. Canada 29:583-587.
377. Burton, D. T., A. H. Jones, and J. Cairns, Jr. 1972. Acute zinc toxicity to rainbow trout (Salmo gairdneri): confirmation of the hypothesis that death is related to tissue hypoxia. J. Fish. Res. Bd. Canada 29(10):1463-1466.
378. Hatfield, C. T., and J. M. Anderson. 1972. Effects of two insecticides on the vulnerability of Atlantic salmon (Salmo salar) parr to brook trout (Salvelinus fontinalis) predation. J. Fish. Res. Bd. Canada 29(1):27-29.
379. Arthur, J. W., and J. G. Eaton. 1971. Chloramine toxicity to the amphipid Gammarus pseudolimnaeus and the fathead minnow, Pimephales promelas. J. Fish. Res. Bd. Canada 28(12):1841-1845.
380. Gilderhus, P. A. 1972. Exposure times necessary for antimycin and rotenone to eliminate certain fresh water fish. Jour. Fish. Res. Bd. Canada 29(2):199-202.
381. Eisler, R. 1971. Cadmium poisoning in Fundulus heteroclitus (Pisces: Cyprinodontidae) and other marine organisms. J. Fish. Res. Bd. Canada 28(9):1225-1234.
382. Webb, P. W., and J. R. Brett. 1972. The effects of sublethal concentrations of whole bleached Kraft mill effluent on the growth and food conversion of underyearling sockeye salmon (Oncorhynchus nerka). J. Fish. Res. Bd. Canada 29(11):1555-1563.
383. Wilson, D. C., and C. E. Bond. 1969. The effects of the herbicides Diquat and Dichlobenil (Casoron) on pond invertebrates. Part I: Acute toxicity. Trans. American Fish. Society. 98(3): 438-443.
384. Jensen, L. D., and A. R. Guafin. 1964. Long-term effects of organic insecticides on two species of stonefly naiads. Trans. Amer. Fish Society. 93(4):357-363.
385. Jensen, L. D., and A. R. Guafin. 1964. Effects of ten organic insecticides on two species of stonefly naiads. Trans. Amer. Fish. Society. 93(1):27-34.
386. Patrick, R., J. Cairns, Jr., and A. Scheier. 1968. The relative sensitivity of diatoms, snails, and fish to twenty common constituents of industrial wastes. Prog. Fish. Cult. 30(3):137-140.
387. Bond, C. E., J. D. Fortune, Jr., and F. Young. 1965. Results of preliminary bioassays with Karosal-SL and Dicamba. Prog. Fish-Cult. 27(1):49-51.
388. Cairns, J., Jr., and A. Scheier. 1968. A comparison of the toxicity of some common industrial waste components tested individually and in combination. Prog. Fish-Cult. 30(1):3-8.



389. Howland, R. M. 1969. Interaction of antimycin A and rotenone in fish bioassays. Prog. Fish-Cult. 31(1):33-34.
390. Heinle, D. R., and R. P. Morgan III. 1972. Bioassay for chronic effects of water from Baltimore Harbor. Final Report to the Maryland Environmental Service. Ref. #72-15.
391. Garrison, R. L. 1968. The toxicity of pro-noxfish to salmonid eggs and fry. Prog. Fish-Cult. 30(1):35-38.
392. Lewis, W. M. 1968. Isobornyl thiocynoacetate as a fish drugging agent and selective toxin. Prog. Fish-Cult. 30(1):29-31.
393. Courtwright, R. C., and C. E. Bond. 1969. Potential toxicity of Kraft mill effluent after oceanic discharge. Prog. Fish-Cult. 31(4):207-212.
394. Ray, J., and V. Stevens. 1970. Using Baytex to control crayfish in ponds. Prog. Fish-Cult. 32(1):58-60.
395. Lloyd, R. 1960. The toxicity of zinc sulphate to rainbow trout. Ann. Appl. Biol. 48(1):84-94.
396. Jensen, L. D., and A. R. Guafin. 1966. Acute and long-term effects of organic insecticides on two species of stonefly naiads. J. Water Poll. Contr. Fed. 38(8):1273-1286.
397. Fletcher, G. L., and R. J. Hoyle. 1972. Acute toxicity of yellow phosphorus to Atlantic cod (Gadus morhua) and Atlantic salmon (Salmo salar) smolts. J. Fish. Res. Bd. Canada 29(9):1295-1301.
398. Marking, L. L. 1969. Toxicity of rhodamine B and fluorescein sodium to fish and their compatibility with antimycin A. Prog. Fish-Cult. 31(3):139-142.
399. LeGore, R. S., and D. M. Des Voigne. 1973. Absence of acute effects on threespine sticklebacks (Gasterosteus aculeatus) and coho salmon (Oncorhynchus kisutch) exposed to resuspended harbor sediment contaminants. J. Fish. Res. Bd. Canada 30(8):1240-1242.
400. Peterson, R. H. 1973. Temperature selection of Atlantic Salmon (Salmo salar) and brook trout (Salvelinus fontinalis) as influenced by various chlorinated hydrocarbons. J. Fish. Res. Bd. Canada 30(8):1091-1097.
401. Smith, E. J., J. L. Sykora, and M. A. Shapiro. 1973. Effect of lime neutralized iron hydroxide suspensions on survival, growth, and reproduction of the fathead minnow (Pimephales promelas). J. Fish. Res. Bd. Canada 30(8):1147-1153.
402. Olson, L. E., and L. L. Marking. 1973. Toxicity of TPM (Lampricide) to six early life stages of rainbow trout. (Salmo gairdneri). J. Fish. Res. Bd. Canada 30(8):1047-1052.
403. Rachlin, J. W., and A. Perlmutter. 1968. Response of an inbred strain of platyfish and the fathead minnow to zinc. Prog. Fish-Cult. 30(4):203-207.

404. Lloyd, R. 1961. The toxicity of mixtures of zinc and copper sulphates to the rainbow trout. (Salmo gairdneri Richardson). Ann. Appl. Biol. 49:535-536.
405. Gannon, J. E., and A. M. Beeton. 1971. Procedures for determining the effects of dredged sediments on biota-benthos viability and sediment selectivity tests. J. Water Pollution Contr. Fed. 43:392-398.
406. Wilson, R. C. H. 1972. Prediction of copper toxicity in receiving waters. J. Fish. Res. Bd. Canada 29(10):1500-1502.
407. Davy, F. B., and H. Kleerekoper, and H. Gensler. 1972. Effects of exposure to sublethal DDT on the locomotor behavior of the goldfish (Carrassius auratus). J. Fish. Res. Bd. Canada 29(9):1333-1336.
408. Sparks, R. E., W. T. Waller, and J. Cairns, Jr. 1972. Effects of shelters on the resistance of dominant and submissive bluegills (Lepomis macrochirus) to a lethal concentration of zinc. J. Fish. Res. Bd. Canada 29(9):1356-1358.
409. Anderson, J. M., and H. B. Prins. 1970. Effects of sublethal DDT on a simple reflex in brook trout. J. Fish. Res. Bd. Canada 27(2):331-334.
410. Lewis, A. G., P. Whitfield, and A. Ramnarine. 1973. The reduction of copper toxicity in a marine copepod by sediment extract. Limnol. and Oceanogr. 18(2):324-326.
411. Smith, L. L., Jr., and D. M. Oseid. 1973. Effect of hydrogen sulfide on development and survival of eight fresh-water fish species. International Symposium on the Early Life History of Fish, Oban, Scotland. May 17-23, 1973.
412. O'Rear, C. W., Jr. 1972. The toxicity of zinc and copper to striped bass eggs and fry with the methods for providing confidence limits. 26th Meeting of the Southern Association of Game and Fish Commissioners, Knoxville, Tenn., Oct. 22-25, 1972.
413. Coleman, R. D., R. L. Coleman, and E. L. Rice. 1971. Zinc and cobalt bioconcentration and toxicity in selected algal species. Bot. Gaz. 32(2):102-109.
414. Courtwright, R. C., W. P. Brease, and H. Krueger. 1971. Formulation of a synthetic seawater for bioassays with Mytilus edulis embryos. Water Res. 5:877-888.
415. Smith, L. L., Jr., and R. H. Kramer. 1963. Survival of walleye eggs in relation to wood fibers and Sphaerotilus natans in the Rainy River, Minnesota. Trans. Amer. Fish. Soc. 92(3):220-234.
416. Emery, R. M. 1970. The comparative acute toxicity of cresol to two benthic crustaceans. Water. Res. 4:485-491.
417. Reish, D. J. 1970. The effects of varying concentrations of nutrients, chlorinity, and dissolved oxygen on polychaetous annelids. Water Res. 4:721-735.

418. Dickman, M. 1969. A quantitative method for assaying the toxic effects of some water soluble substances based on changes in periphyton community structure. *Water Res.* 3:963-972.
419. Solon, J. M., J. L. Lincer, and J. H. Nair III. 1969. The effect of sublethal concentrations of LAS on the acute toxicity of various insecticides to the fathead minnow (Pimephales promelas Rafinesque). *Water Res.* 3:767-775.
420. Bender, M. E. 1969. The toxicity of the hydrolysis and breakdown products of malathion to the fathead minnow (Pimephales promelas Rafinesque). *Water Res.* 3:571-582.
421. Walden, C. C., T. E. Howard, and G. C. Frond. 1970. A quantitative assay of the minimum concentration of Kraft mill effluents which affect respiration. *Water. Res.* 4:61-68.
422. Lloyd, R., and L. D. Orr. 1969. The diuretic response by rainbow trout to sublethal concentrations of ammonia. *Water Res.* 3:335-344.
423. Ball, I. R. 1967. The toxicity of cadmium to rainbow trout. (Salmo gairdneri Richardson). *Water Res.* 1:805-806.
424. Ball, I. R. 1967. The relative susceptibilities of some species of fresh-water fishes to poisons II Zinc. *Water Res.* 1:777-783.
425. Ball, I. R. 1967. The relative susceptibilities of some species of fresh-water fishes to poisons - I. Ammonia. *Water Res.* 1:767-775.
426. Schaumburg, G. D., T. E. Howard, and C. C. Walden. 1967. A method to evaluate the effects of water pollutants on fish respiration. *Water Res.* 1:731-737.
427. Brown, V. M., D. G. Shurben, and J. K. Fawell. 1967. The acute toxicity of phenol to rainbow trout in saline waters. *Water Res.* 1:683-685.
428. Brown, V. M., D. H. M. Jordan, and B. A. Tiller. 1967. The effect of temperature on the acute toxicity of phenol to rainbow trout in hard water. *Water Res.* 1:587-594.
429. Parker, C. 1966. Influence of water hardness on the phytotoxicity of Paraquat, *Nature.* 212:1465-1466.
430. Sprague, J. B. 1968. Promising anti-pollutant: chelating agent NTA protects fish from copper and zinc. *Nature* 220:1345-1346.
431. Alabaster, J. S., and F. S. H. Abram. 1965. Development and use of a direct method of evaluating toxicity to fish. In Advances in Water Pollution Research, Proc. 2nd Int'l Conf. Tokyo, 1964, v.i. Pergamon Press, Oxford, p. 41-54.
432. Arthur, J. W. 1970. Chronic effects of linear alkylate sulfonate detergent on Gammarus pseudolimnaeus, Campeloma decisum, and Physa integra. *Water Res.* 4:251-257.

433. Eaton, J. G. 1970. Chronic malathion toxicity to the bluegill (Lepomis macrochirus Rafinesque). Water. Res. 4:673-684.
434. Isom, B. G. 1960. Toxicity of elemental phosphorus. J. Water. Poll. Contr. Fed. 32(12):1312-1316.
435. Cairns, J., Jr., and A. Scheier. 1957. The effects of temperature and hardness of water upon the toxicity of zinc to the common bluegill (Lepomis macrochirus Raf.). Not. Nat. 299, 12 p.
436. Cairns, J., Jr., and A. Scheier. 1964. The effect upon the pumpkinseed sunfish Lepomis gibbosus (Linn.) of chronic exposure to lethal and sublethal concentrations of dieldrin. Not. Nat. 370, 10 p.
437. Cairns, J., Jr., and A. Scheier. 1963. Environmental effects upon ozanide toxicity to fish. Not. Nat. 361, 11 p.
438. Cairns, J., Jr., and A. Scheier. 1962. The effects of temperature and water hardness upon the toxicity of naphthenic acids to the common bluegill Lepomis macrochirus Raf. and the pond snail Physa heterostrophia Say. Not. Nat. 353, 12 p.
439. Alderdice, D. F., and J. R. Brett. 1957. Some effects of Kraft mill effluent on young Pacific salmon. J. Fish. Res. Bd. Canada 14(5):783-795.
440. Dowden, B. F., and H. J. Bennett. 1965. Toxicity of selected chemicals to certain animals. J. Water Poll. Contr. Fed. 37(9):1308-1316.
441. Alderdice, D. F. 1963. Some effects of simultaneous variation in salinity, temperature, and dissolved oxygen on the resistance of young coho salmon to a toxic substance. J. Fish. Res. Bd. Canada 20(2):525-550.
442. Hidu, H. 1965. Effects of synthetic surfactants on the larvae of clams (M. mercenaria) and oysters (Crassostrea virginica). J. Water Poll. Contr. Fed. 37(2):262-270.
443. Pickering, Q. H., and C. Henderson. 1966. Acute toxicity of some important petrochemicals to fish. J. Water Poll. Contr. Fed. 38(9):1419-1429.
444. Summerfelt, R. C., and W. M. Lewis. 1967. Repulsion of green sunfish by certain chemicals. J. Water Poll. Contr. Fed. 39(12):2030-2038.
445. Chadwick, G. C., and U. Kiigemazi. 1968. Toxicity evaluation of a technique for introducing dieldrin into water. J. Water Poll. Contr. Fed 40(1):76-82.
446. Carlson, G. F., Jr., F. E. Woodard, D. F. Wentworth, and O. J. Sproul. 1968. Virus inactivation on clay particles in natural waters. J. Water Poll. Contr. Fed. 40:76-82.

447. Anderson, D. A. 1965. Growth responses of certain bacteria to ABS and other surfactants. Proc. 19th Ind. Waste Conf. 1964, Purdue University Engineering Extension Series No. 117, p. 592-601.
448. Pickering, Q. H., and C. Henderson. 1965. The acute toxicity of some heavy metals to different species of warm water fishes. Proc. 19th Ind. Waste Conf., 1964 Purdue Univ. Engineering Extension Series No. 117, p. 578-591.
449. Dugan, P. R. 1967. Influence of chronic exposure to anionic detergents on toxicity of pesticides to goldfish. J. Water Poll. Contr. Fed. 39(1):63-71.
450. Hanes, N. B., and R. Frangala. 1967. Effect of seawater concentration on survival of indicator bacteria. J. Water Poll. Contr. Fed. 39(1):97-104.
451. Henderson, C., Q. H. Pickering, and J. M. Cohen. 1959. The toxicity of synthetic detergents and soaps to fish. Sew. Ind. W. 31(3):295-306.
452. Weiss, C. M. 1959. Response of fish to sublethal exposures of organic phosphorus insecticides. Sew. Ind. Wastes 31(5):580-593.
453. Weiss, C. M., and J. L. Botts. 1957. Factors affecting the response of fish to toxic materials. Sew. Ind. W. 29(7):810-818.
454. Malaney, G. W., W. D. Sheets, and R. Quillen. 1959. Toxic effects of metallic ions on sewage microorganisms. Sew. Ind. Wastes 31(11):1309-1315.
455. Eisler, R. 1965. Some effects of a synthetic detergent on estuarine fishes. Trans. Amer. Fish. Soc. 94:26-31.
456. Foster, N. R., A. Scheier, and J. Cairns, Jr. 1966. Effects of ABS on feeding behavior of flagfish, Jordanella floridae. Trans. Amer. Fish. Soc. 95:109-110.
457. Trama, F. B., and R. J. Benoit. 1960. Toxicity of hexavalent chromium to goldfish. J. Water Poll. Contr. Fed. 32(8):868-877.
458. Hood, D. W., T. W. Duke, and B. Stevenson. 1960. Measurement of toxicity of organic wastes to marine organisms. J. Water. Poll. Contr. Fed. 23(9):982-993.
459. Ferguson, D. E., and C. R. Bingham. 1966. The effects of combinations of insecticides on susceptible and resistant mosquitofish. Bull. Envir. Cont. Tox. 1:97-103.
460. Hicks, C. E., and J. M. Neuhold. 1966. Alkyl benzene sulfonate effects on stream algae communities. Bull. Envir. Contr. Toxicol. 1:225-236.
461. Hazel, C. R., and S. J. Meith. 1970. Bioassay of king salmon eggs and sac fry in copper solutions. Calif. Fish. Game 56(2):121-124.

462. McKim, J. M., and D. A. Benoit. 1971. Effects of long-term exposures to copper on survival, growth, and reproduction of brook trout (Salvelinus fontinalis). J. Fish. Res. Bd. Canada 28(5):655-662.
463. Johnson, B. T., C. R. Saunders, and H. O. Sanders. 1971. Biological magnification and degradation of DDT and aldrin by freshwater invertebrates. J. Fish. Res. Bd. Canada 28(5):655-662.
464. Lowe, J. I. 1967. Effects of prolonged exposure to Sevin on an estuarine fish, Leiostomus xanthurus Lacepede. Bull. Envir. Contam. Toxicol. 2:147-155.
465. Jones, J., R. Erichsen. 1938. The relative toxicity of salts of lead, zinc, and copper to the stickleback (Gasterosteus aculeatus L.) and the effect of calcium on the toxicity of lead and zinc salts. J. Expt. Biol. 15:394-407.
466. Dawson, A. B. 1935. The hemopoietic response in the catfish Ameiurus nebulosus, to chronic lead poisoning. Biol. Bull. 68: 335-346.
467. Lloyd, R. 1961. The effect of dissolved oxygen concentrations on the toxicity of several poisons to rainbow trout (Salmo gairdneri Richardson). J. Expt. Biol. 38:447-455.
468. Lowe, J. I. 1965. Chronic exposure of blue crabs, Callinectes sapidus, to sublethal concentrations of DDT. 46:899-900.
469. Tagatz, M. E. 1961. Reduced oxygen tolerance and toxicity of petroleum to juvenile American shad. Ches. Sci. 2(1-2):65-71.
470. Rawls, C. K. 1965. Field tests of herbicide toxicity to certain estuarine animals. Ches. Sci. 6(3):150-161.
471. Eisler, R., and M. P. Weinstein. 1967. Changes in metal composition of the Quahog clam, Mercenaria mercenaria after exposure to insecticides. Ches. Sci. 8(4):253-258.
472. Dahlberg, M. D. 1971. Toxicity of acrolein to barnacles (Balanus eburneus). Ches. Sci. 12(4):282-284.
473. Karlander, E. P., and R. W. Krauss. 1972. Absorption and toxicity of beryllium and lithium in Chlorella vannielli Shihira and Krauss. Ches. Sci. 13(4):245-253.
474. Morgan, R. P. II, R. F. Fleming, V. J. Rasin, Jr., and D. R. Heinle. 1973. Sublethal effects of Baltimore harbor water on the white perch, Morone americana, and the hogchoker, Trinectes maculatus. Ches. Sci. 14(1):17-27.
475. Eisler, R. 1967. Acute toxicity of zinc to the killfish, Fundulus heteroclitus. Ches. Sci. 8(4):262-264.
476. Mitrovic, V. V., V. M. Brown, D. G. Shurben, and M. H. Berryman. 1968. Some pathological effects of sub-acute and acute poisoning of rainbow trout by phenol in hard water. Water Res. 2:249-254.

477. Brown, V. M., V. V. Mitrovic, and G. T. C. Stark. 1968. Effects of chronic exposure to zinc on toxicity of a mixture of detergent and zinc. *Water Res.* 2:255-263.
478. Sprague, J. B. 1968. Avoidance reactions of rainbow trout to zinc sulphate solutions. *Water Res.* 2:367-372.
479. Sprague, J. B. 1968. Avoidance reactions of salmonid fish to representative pollutants. *Water Res.* 2:23-24.
480. Huang, J., and E. F. Glogna. 1968. Effect of organic compounds on photosynthetic oxygenation - I. Chlorophyll destruction and suppression of photosynthetic oxygen production. *Water Res.* 2:347-366.
481. Pickering, Q. H. 1968. Some effects of dissolved oxygen concentrations upon the toxicity of zinc to the bluegill, Lepomis macrochirus, Raf. *Water Res.* 2:187-194.
482. Sprague, J. B., and D. W. McLeese. 1969. Toxicity of Kraft pulp mill effluent for larval and adult lobsters and juvenile salmon. *Water Res.* 2:753-760.
483. Sprague, J. B., and D. W. McLeese. 1968. Different toxic mechanisms in Kraft pulp mill effluent for two aquatic animals. *Water Res.* 2:761-765.
484. Graham, R. J., and T. C. Dorris. 1968. Long-term toxicity bioassay of oil refinery effluents. *Water Res.* 2:643-663.
485. Tracy, H. B., R. A. Lee, C. E. Woelke, and G. Sanborn. 1969. Relative toxicities and oil dispersing evaluations of eleven oil-dispersing products. *J. Water Poll. Contr. Fed.* 41(12):2062-2069.
486. Warnick, S. L., and H. L. Bell. 1969. The acute toxicity of some heavy metals to different species of aquatic insects. *J. Water Poll. Contr. Fed.* 41(2):280-284.
487. Davis, H. C. 1961. Effects of some pesticides on eggs and larvae of oysters (C. virginica) and clams (M. mercenaria). *Comm. Fish. Rev.* 23(12):8-23.
488. Gaufin, A. R. 1961. Bioassays to determine the toxicity of pesticides to aquatic invertebrates. *Proc. 15th Ind. Waste Conf., Purdue Univ., 1960, Eng. Ext. Ser. #106:94-98.*
489. Henderson, C., Q. H. Pickering, and A. E. Lemke. 1961. The effect of some organic cyanides (nitriles) on fish. *Proc. 15th Ind. Waste Conf., Purdue Univ., 1960, Eng. Ext. Ser. #106:120-130.*
490. Lammering, M. W., and N. C. Burbank. 1961. The toxicity of phenol, o-chlorophenol, and o-nitrophenol to bluegill sunfish. *Proc. 15th Ind. Waste Conf., Purdue Univ., 1960, Eng. Ext. Ser. #106:541-555.*

491. Gardner, G. R., and P. P. Yevich. 1969. Toxicological effects of cadmium on Fundulus heteroclitus under various oxygen, pH, salinity, and temperature changes. Amer. Zool. 9:1096.
492. Grosch, D. S. 1967. Poisoning with DDT: Effect on reproductive performance of Artemia. Sci. 155:592-593.
493. Maloney, T. E., and E. L. Robinson. 1961. Growth and respiration of a green alga in spent sulfite liquor. J. Tech. Assn. Pulp Paper Ind. 44:137-141.
494. Fromm, P. O., and R. H. Schittman. 1958. Toxic action of hexavalent chromium on largemouth bass. J. Wildl. Manage. 22:40-44.
495. Belding, D. L. 1929. The respiratory movements of fish as an indicator of a toxic environment. Trans. Amer. Fish. Soc. 59: 238-245.
496. Herbert, D. W. M. 1962. The toxicity to rainbow trout of spent still liquors from the distillation of coal. Ann. Appl. Biol. 50:755-777.
497. Henderson, C., Q. H. Pickering, and C. M. Tarzwell. 1959. Relative toxicity of ten chlorinated hydrocarbons to four species of fish. Trans. Amer. Fish. Soc. 88:23-32.
498. Davidson, G. 1963. DDT resistance and dieldrin resistance in Anopheles quadrimaculatus. Bull. WHOI 29:177-184.
499. Daugherty, F. M., Jr. 1951. A proposed toxicity test for industrial wastes to be discharged to marine waters. Sewage Ind. Wastes. 23:1029-1031.
500. Freeman, L. 1953. Toxicity thresholds of certain sodium sulfonates for Daphnia magna Straus. Sewage Ind. Wastes. 25:1331-1335.
501. Freeman, L., and I. Fowler. 1953. Toxicity of combinations of certain inorganic compounds to Daphnia magna Straus. Sewage Ind. Wastes. 25:1191-1195.
502. Doudoroff, P., M. Katz, and C. M. Tarzwell. 1953. Toxicity of some organic insecticides to fish. Sewage Ind. Wastes. 25:840-844.
503. Wiebe, A. H. 1930. Notes on the exposure of young fish to varying concentrations of arsenic. Trans. Amer. Fish. Soc. 60:270-278.
504. Guthrie, J. E., and O. E. Acres. 1970. Toxicity to fish of two organic reactor coolants. Bull. Envir. Contam. Toxicol. 5:145-151.
505. Moss, S. A., and W. N. McFarland. 1970. The influence of dissolved oxygen and carbon dioxide on fish schooling behavior. Marine Biology 5:100-107.



506. Wildish, D. J. 1970. The toxicity of polychlorinated biphenyls (PCB) in seawater to Gammarus oceanicus. Bull. Envir. Contam. Toxicol. 5:202-204.
507. Baptist, J. P. 1966. Uptake of mixed fission products by marine fishes. Trans. Amer. Fish. Soc. 95(2):145-152.
508. Jones, J. C. 1957. A study of possible modifications of the WHO method for testing DDD resistance in mosquito larvae with special reference to Anopheles quadrimaculatus. (Say.). Bull. WHO 36(2): 353-356.
509. O'Hara, J. 1971. Alterations in oxygen consumption by bluegills exposed to sublethal treatment with copper. Water Res. 5:321-327.
510. Mount, D. I. 1962. Chronic effects of endrin on bluntnose minnows and guppies. Bureau Sport. Fish. Wildlife Res. Rept. No. 58.
511. Allison, D., B. J. Kallman., O. B. Cope, and C. Van Valin. 1964. Some chronic effects of DDT on cutthroat trout. Bureau of Sport Fisheries and Wildlife Res. Rpt. 64.
512. Pratt, S. D., S. B. Saila, A. G. Gaines, Jr., and J. E. Kront. 1972. "Bioassays" in: Biological Effects of Ocean Disposal of Solid Wastes. Marine Experiment Station. U. Rhode Island Marine Technical Report Series #9.
513. Murray, S., J. Scherfig, and P. S. Dixon. 1971. Evaluation of algal assay procedures - PAAP Batch test. J. Wat. Poll. Contr. Fed. 43(10):1991-2003.
514. Carter, J. W., and I. L. Cameron. 1973. Toxicity bioassay of heavy metals in water using Tetrahymena pyriformis. Wat. Res. 7:951-961.
515. Jernelev, A., R. Rosenberg, and S. Jensen. 1972. Biological effects and physical properties in the marine environment of aliphatic chlorinated by-products from vinyl chloride production. Wat. Res. 6:1181-1191.
516. Davis, H. C. 1960. Effects of turbidity-producing materials in sea water on eggs and larvae of the clam (Venus [Mercenaria] mercenaria). Biol. Bull. 118(1):48-54.
517. Weiss, C. M., and J. H. Gakstatter. 1964. Detection of pesticides in water by biochemical assay. J. Wat. Poll. Contr. Fed. 36(2):240-253.
518. Sparks, R. E., J. Cairns, Jr., and A. G. Heath. 1972. The use of bluegill breathing rates to detect zinc. Wat. Res. 6:895-911.
519. Kontogiannis, J. E., and C. J. Barnett. 1973. The effect of oil pollution on survival of the tidepool copepod Tigriopus californicus. Envir. Pollut. 4:69-79.
520. Zitko, V. 1970. Polychlorinated biphenyls (PCB) solubilized in water by anionic surfactants for studies of toxicity to aquatic animals. Bull. Envir. Contam. Toxicol. 5(3):279-285.

521. Davis, H. C., and H. Hidu. 1969. Effects of pesticides on embryonic development of clams and oysters and on survival and growth of the larvae. *Fishery Bull.* 67(2):393-404.
522. Murphy, P. G. 1970. Effects of salinity on uptake of DDT, DDE, and DDD by fish. *Bull. Envir. Contam. Toxicol.* 5(5):404-407.
523. Rachlin, J. W., and A. Perlmutter. 1969. Response of rainbow trout cells in culture to selected concentrations of zinc sulfate. *Prog. Fish-Cult.* 31(2):94-98.
524. Morgan, J. R. 1972. Effects of Aroclor 1242 (a polychlorinated biphenyl) and DDT on cultures of an alga, protozoan, daphnid, ostracod and guppy. *Bull. Envir. Contam. Toxicol.* 8(3):129-137.
525. Rehwoldt, R., L. W. Menapace, B. Nerrie, and D. Alessandrello. 1972. The effect of increased temperature upon the acute toxicity of some heavy metal ions. *Bull. Envir. Contam. Toxicol.* 8(2): 91-96.
526. Albaugh, D. W. 1972. Insecticide tolerances of two crayfish populations (Procambarus acutus) in south central Texas. *Bull. Envir. Contam. Toxicol.* 8(6):334-338.
527. Hansen, D. J., E. Matthews, S. L. Nall, and D. P. Dumas. 1972. Avoidance of pesticides by untrained mosquitofish (Gambusia affinis). *Bull. Envir. Contam. Toxicol.* 8(1):46-57.
528. Skidmore, J. F., and P. W. A. Tovell. 1972. Toxic effects of zinc sulphate on the gills of rainbow trout. *Wat. Res.* 6:217-230.
529. Waller, W. T., and J. Cairns. 1972. The use of fish movement patterns to monitor zinc in water. *Water Research* 6:257-269.
530. Smith, L. L., and D. M. Oseid. 1972. Effects of hydrogen sulfide on fish eggs and fry. *Water Research* 6:711-720.
531. Norup, B. 1972. Toxicity of chemicals in paper factory effluents. *Water Research* 6:1585-1588.
532. Hemens, J., and R. J. Wareick. 1972. The effects of fluoride on estuarine organisms. *Water Research* 6:1301-1308.
533. Sprague, J. B. 1964. Avoidance of copper-zinc solutions by young salmon in the laboratory. *Journal of Water Pollution Control Federation* 36(8):990-1004.
534. McLean, R. I. 1973. Chlorine and temperature stress on estuarine organisms. *Journal of Water Pollution Control Federation* 45(5): 837-841.
535. Okubo, K., and T. Okubo. 1962. Study on the bioassay method for the evaluation of water pollution. II. Use of the fertilized eggs of sea urchins and bivalves. *Bull. Tokai Reg. Fish. Res. Lab.* 32:131-140.

536. Perkins, E. J. 1968. The toxicity of oil emulsifiers to some inshore fauna in: The Biological Effects of Oil Pollution on Littoral Communities. U. of Strathclyde, p. 81-90.
537. Forsberg, C. G. 1972. Algal assay procedure. Journal of Water Pollution Control Federation 44(8):1623-1628.
538. Lewis, S. D., and W. M. Lewis. 1971. The effect of zinc and copper on the osmolality of blood serum of the channel catfish (Ictalurus punctatus Rafinesque) and golden shiner (Notemigonus chrysoleucas Mitchill). Trans. Amer. Fish. Soc. 100(4):639-643.
539. Moore, R. B. 1970. Effects of pesticides on growth and survival of Euglena gracilis. Bull. Envir. Contam. Toxicol. 5:226-230.
540. Holland, H. T., and D. L. Coppage. 1970. Sensitivity to pesticides in three generations of sheepshead minnows. Bull. Envir. Contam. Toxicol. 5:362-367.
541. Solon, J. M., and J. H. Nair III. 1970. The effect of a sublethal concentration of LAS on the acute toxicity of various phosphate pesticides to the fathead minnow (Pimephales promelas Rafinesque) Bull. Envir. Contam. Toxicol. 5:408-413.
542. Derby, S. B., and E. Ruber. 1970. Primary production: Depression of oxygen evolution in algal cultures by organophosphorus insecticides. Bull. Envir. Contam. Toxicol. 5:553-558.
543. Pickering, Q. H., and C. Henderson. 1966. The acute toxicity of some pesticides to fish. Oh. J. Sci. 66:508-513.
544. Egloff, D. A., and R. Partridge. 1972. Resistance to DDT of a freshwater alga. Oh. J. Sci. 72(1):6-10.
545. Kauss, P., T. C. Hutchinson, C. Soto, J. Helebust, and M. Griffiths. 1973. The toxicity of crude oil and its components to freshwater algae. Proc. Joint Conf. on Prevention and Control of Oil Spills. March 13-15, 1973, Washington, D. C.
546. Rice, S. D. 1973. Toxicity and avoidance tests with Prudhoe Bay oil and pink salmon fry. Proc. Joint Conf. on Prevention and Control of Oil Spills. March 13-15, 1973, Washington, D. C.
547. Gilfillan, E. S. 1973. Effects of seawater extracts of crude oil on carbon budgets in two species of mussels. Proc. Joint Conf. on Prevention and Control of Oil Spills. March 13-15, 1973, Washington, D. C.
548. Brocksen, R. W., and H. T. Bailey. 1973. Respiratory response of juvenile chinook salmon and striped bass exposed to benzene, a water soluble component of crude oil. Proc. Joint Conf. on Prevention and Control of Oil Spills. March 13-15, 1973, Washington, D. C.
549. Nuzzi, R. 1973. Effects of water soluble extracts of oil on phytoplankton. Proc. Joint Conf. on Prevention and Control of Oil Spills. March 13-15, 1973. Washington, D. C.

550. Nadeau, R. J., and T. H. Roush. 1973. A salt marsh microcosm: An experimental unit for marine pollution studies. Proc. Joint Conf. on Prevention and Control of Oil Spills. March 13-15, 1973, Washington, D. C.
551. Whitley, L. S. 1968. The resistance of tubificid worms to three common pollutants. *Hydrobiologia* 32:193-205.
552. Roberts, D. 1972. The assimilation and chronic effects of sub-lethal concentrations of endosulfan on condition and spawning in the common mussel (Mytilus edulis). *Marine Biology* 16:119-125.
553. Bellan, G., D. J. Reish, and J. P. Foret. 1972. The sublethal effects of a detergent on the reproduction, development, and settlement in the polychaetous annelid Capitella capitata. *Marine Biology* 14:183-188.
554. Sprague, J. B. 1964. Lethal concentrations of copper and zinc for young Atlantic salmon. *J. Fish. Res. Bd. Canada* 21(1):17-26.
555. Granmo, A. 1972. Development and growth of eggs and larvae of Mytilus edulis exposed to a linear dodecylbenzenesulphonate, LAS. *Marine Biology* 15:356-358.
556. Mandelli, E. F. 1969. The inhibitory effects of copper on marine phytoplankton. *Contributions in Marine Biology* 11:191-197.
557. Nimmo, D. R., R. R. Blackman, A. J. Wilson, Jr., and J. Forester. 1971. Toxicity and distribution of Aroclor 1254 in the pink shrimp (Penaeus duorarum). *Marine Biology* 11:191-197.
558. Renzoni, A. 1973. Influence of crude oil, derivatives and dispersants on larvae. *Marine Pollution Bulletin* 4(1):9-13.
559. Wells, P. G. 1972. Influence of Venezuelan crude oil on lobster larvae. *Marine Pollution Bulletin* 3(7):105-106.
560. Perkins, E. J., E. Gribbon, and J. W. M. Logan. 1973. Oil dispersant toxicity. *Marine Pollution Bulletin* 4(5):90-93.
561. Pybus, C. 1973. Effects of anionic detergent on the growth of Laminaria. *Marine Pollution Bulletin* 4(5):73-77.
562. Chia, F. 1973. Killing of marine larvae by diesel oil. *Marine Pollution Bulletin* 4(2):29-30.
563. Connor, P. M. 1972. Acute toxicity of heavy metals to some marine larvae. *Marine Pollution Bulletin* 3(12):190-192.
564. Lowe, J. I., P. R. Parrish, J. M. Patrick, Jr., and J. Forester. 1972. Effects of the polychlorinated biphenyl Aroclor 1254 on the American oyster (Crassostrea virginica). *Marine Biology* 17:209-214.
565. Lewis, A. G., P. H. Whitfield, and A. Ramnarine. 1972. Some particulate and soluble agents affecting the relationship between metal toxicity and organism survival in the calanoid copepod, Euchaeta japonica. *Marine Biology* 17:215-221.

566. Whitley, L. S., and R. A. Sikora. 1970. The effect of three common pollutants on the respiration rate of tubificid worms. Journ. Wat. Poll. Cont. Fed. 42(2):R57-R66.
567. Sanders, H. 1970. Toxicities of some herbicides to six species of freshwater crustaceans. Journ. Wat. Poll. Cont. Fed. 42(8): 1544-1550.
568. Epifanio, C. E. 1971. Effects of dieldrin in seawater on the development of two species of crab larvae, Leptodius floridanus and Panopeus herbstii. Marine Biology 11:356-362.
569. Epifanio, C. E. 1972. Effects of dieldrin contaminated food on the development of Leptodius floridanus larvae. Marine Biology 13:292-297.
570. Steed, D. L., and B. J. Copeland. 1967. Metabolic responses of some estuarine organisms to an industrial effluent. Contributions in Marine Science 12:143-159.
571. Mount, D. I., L. W. Vigor, and M. L. Schafer. 1966. Endrin: Use of concentration in blood to diagnose acute toxicity to fish. Science: 152(3727):1388-1390.
572. Ukeles, R. 1962. Growth of pure cultures of marine phytoplankton in the presence of toxicants. Applied Microbiology 10:532-537.
573. Natarajan, K. V. 1970. Toxicity of ammonia to marine diatoms. Journ. Wat. Poll. Cont. Fed. 42(5):184-190.
574. Pickering, Q. H., and T. O. Thatcher. 1970. The chronic toxicity of linear alkylate sulfonate (LAS) to Pimphales promelas Rafinesque. Journ. Wat. Poll. Cont. Fed. 42(2):243-254.
575. Erickson, S. J., T. E. Maloney, and J. H. Gentile. 1970. Effect of nitrilotriacetic acid on the growth and metabolism of estuarine phytoplankton. Journ. Wat. Poll. Cont. Fed. 42(8):R329-R335.
576. Erickson, S. J., N. Lackie, and T. E. Maloney. 1970. A screening technique for estimating copper toxicity to estuarine phytoplankton. Journ. Wat. Poll. Cont. Fed. 42(8):R270-R279.
577. Portmann, J. E. 1972. Results of acute toxicity tests with marine organisms, using a standard method. Marine Pollution and Sea Life: 212-217.
578. Griffith, D. de G. 1972. Toxicity of crude oil and detergents to two species of edible molluscs under artificial tidal conditions. Marine Pollution and Sea Life: 224-229.
579. Jensen, S., A. Jernelov, R. Lange, and K. H. Palmork. 1972. Chlorinated by-products from vinyl chloride production: A new source of marine pollution. Marine Pollution and Sea Life: 242-244.
580. Bellan, G., J. P. Foret, P. Foret-Montardo, and R. A. Kaim-Malka. 1972. Action in vitro of detergents on some marine species.

- Station Marine d'Endoume, 13, Marseille, France. Marine Pollution and Sea Life: 245-248.
581. Mann, H. G. W. 1972. Toxicity and degradation of tensides in sea water. Marine Pollution and Sea Life: 248-250.
  582. Engel, R. H., M. J. Neat, and R. E. Hillman. 1972. Sublethal chronic effects of DDT and Lindane on glycolytic and gluconeogenic enzymes of the quahog Mercenaria mercenaria. Marine Pollution and Sea Life: 257-260.
  583. Hirayama, K., and R. Hirano. 1970. Influences of high temperature and residual chlorine on marine phytoplankton. Marine Biology 7:205-213.
  584. Nielsen, E. Steeman, and S. Wium-Andersen. 1970. Copper ions as poison in the sea and in fresh water. Marine Biology 6:93-97.
  585. Davey, F. B., H. Kleerehoper, and J. H. Matis. 1973. Effects of exposure to sublethal DDT on the exploratory behavior of goldfish (Carassius auratus). Water Resources Res. 9(4):900-905.
  586. Winter, J. E. 1972. Long-term laboratory experiments on the influence of ferric hydroxide flakes on the filter-feeding behaviour, growth, iron content, and mortality in Mytilus edulis L. Marine Pollution and Sea Life: 315-318.
  587. Kuhnhold, W. W. 1972. The influence of crude oils on fish fry. Marine Pollution and Sea Life: 315-318.
  588. Wilson, K. W. 1972. Toxicity of oil-spill dispersants to embryos and larvae of some marine fish. Marine Pollution and Sea Life: 318-322.
  589. Hannan, P. J., and C. Patouillet. 1972. Nutrient and pollutant concentrations as determinants in algal growth rates. Marine Pollution and Sea Life: 340-342.
  590. Swedmark, M., B. Braaten, E. Emanuelson, and A. Granmo. 1971. Biological effects of surface active agents on marine animals. Marine Biology: 9:183-201.
  591. Theede, H., A. Ponat, K. Hiroi, and C. Schlieper. 1969. Studies on the resistance of marine bottom invertebrates to oxygen deficiency and hydrogen sulphide. Marine Biology 2:325-337.
  592. Raymont, J. E. G., and J. Shields. 1964. Toxicity of copper and chromium in the marine environment. Advances in Water Pollution Research 3:275-283.
  593. Zillich, J. A. 1972. Toxicity of combined chlorine residuals to freshwater fish. Journ. Wat. Poll. Cont. Fed. 44(2):212-220.
  594. Belle W. Baruch Coastal Research Institute. 1973. Bioassay Studies, Charleston Harbor, South Carolina. The effects of dredging sediments on plankton. Final report submitted to the U. S. Army Corps of Engineers, Charleston District.

595. Coppage, D. L. 1972. Organophosphate pesticides: Specific level of brain AChE inhibition related to death in sheephead minnows. Trans. Amer. Fish. Soc. 101(3):534-536.
596. Poon, C., and K. H. Bhayan. 1971. Metal toxicity to sewage organisms. Journ. of Sanitary Engineering Division 97(SA2):161-169.
597. Nuzzi, R. 1972. Toxicity of mercury to phytoplankton. Nature 237(5349):38-39.
598. Hendricks, C. W. 1971. Enteric bacterial metabolism of stream sediments eluates. Canadian Journal of Microbiology 17(4):551-556.
599. North, W. J. 1963. An investigation of the effects of discharged wastes on kelp. State Water Quality Control Board, Publ. #26.
600. Eller, L. L. 1971. Histopathologic lesions in cutthroat trout (Salmo clarki) exposed chronically to the insecticide Endrin. Amer. J. of Pathology 64:321-336.
601. Sprague, J. B. 1964. Lethal concentrations of copper and zinc for young Atlantic salmon. J. Fish. Res. Bd. Canada 21(1):17-26.
602. Sprague, J. B., and B. A. Ramsey. 1965. Lethal levels of mixed copper-zinc solutions for juvenile salmon. J. Fish. Res. Bd. Canada 22(2):425-432.
603. Saila, S. B. 1953. Bioassay procedures for the evaluation of fish toxicants with particular reference to rotenone. Trans. Amer. Fish. Soc.:104-114.
604. Davis, H. C., and H. Hidu. 1969. Effects of turbidity producing substances in sea water on eggs and larvae of three genera of bivalve mollusks. Veliger 11(4):316-323.
605. Hollister, T. A., and G. E. Walsh. 1973. Differential responses of marine plankton to herbicides.
606. Hansen, D. J., S. C. Schimmel, and J. Keltner, Jr. 1973. Avoidance of pesticides by grass shrimp (Palaemonetes pugio). Bull. Envir. Contam. and Toxicology 9(3):129-133.
607. Earnest, R. D., and P. Benville. 1972. Acute toxicity of four organochlorine insecticides to two species of surf perch. Calif. Fish Game 58(2):127-132.
608. Macek, K. J. 1968. Growth and resistance to stress in brook trout fed sublethal levels of DDT. J. Fish. Res. Bd. Canada 25(11):2443-2451.
609. Wilson, R. C. H. 1972. Acute toxicity of spent sulfite liquor to Atlantic salmon (Salmo salar). J. Fish. Res. Bd. Canada 29(8):1225-1228.
610. Otto, R. G. 1971. Effects of salinity on the survival and growth of pre-smolt Coho salmon (Oncorhynchus kisutch). J. Fish. Res. Bd. Canada 28:343-349.

611. Duke, T. W. 1967. Possible routes of zinc 65 from an experimental estuarine environment to man. *Journal of Water Pollution Control Federation* 39(4):536-542.
612. Lowe, J. I., R. R. Parrish, A. J. Wilson, Jr., P. D. Wilson, and T. W. Duke. 1971. Effects of mirex on selected estuarine organisms. *Trans. of 36th North American Wildlife and Natural Resources Conference*.
613. Maki, A. W., K. W. Stewart, and J. K. Silvey. 1973. The effects of dilrom on respiratory activity of the stonefly (Hydropsyche crosbyi), hellgrammite (Corydalis cornutus), and the golden shiner (Notemigonus crysoleucas). *Trans. American Fish. Soc.* 102(4): 806-815.
614. DeCoursey, P., and W. Vernberg. 1972. Effects of mercury on survival, metabolism and behavior of larval Uca pugnator (Brachyura). *Oikos* 23:241-247.
615. Kennedy, H., L. L. Eller, and D. F. Walsh. 1970. Chronic effects of methoxychlor on bluegills and aquatic invertebrates. *Technical Papers of Bureau of Sport Fishing and Wildlife*, Number 53.
616. Wilhm, J. L. 1970. Transfer of radioisotopes between detritus and benthic macroinvertebrates in laboratory microecosystems. *Health Physics* 18:277-284.
617. Reish, D. J., and W. M. Hetherington, III. 1969. The effects of hyper- and hypo-chlorinities on members of the wood-boring genus Limnoria. *Marine Biology* 2:137-139.
618. Hansen, D. J., P. R. Parrish, J. I. Lowe, A. J. Wilson, Jr., and P. D. Wilson. 1971. Chronic toxicity, uptake, and retention of anoclor 1254 in two estuarine fishes. *Bulletin of Environmental Control and Toxicity*, Volume 6, Number 2:113-119.
619. Kawatski, J. 1973. Acute toxicities of antimycin A, Bayer 73, and TFM to the ostracod Cyprina kawatai. *Trans. Amer. Fish. Soc.* 102(4):829-831.
620. Heimstra, N. W., D. K. Damkot, and N. Benson. 1969. Some effects of silt turbidity on behavior of juvenile largemouth bass and green sunfish. *Technical Papers of Bureau of Sport Fishing and Wildlife*, Number 20.
621. Bell, H. 1970. Effects of pH on the life cycle of the midge Tanytarsus dissimilis. *Canadian Entomologist* 102:636-639.
622. Scherer, E. 1971. Effects of oxygen depletion and of carbon dioxide buildup on the photic behavior of the walleye (Stizostedion vitreum vitreum). *Jour. Fish. Res. Bd. Canada* 28(9):1303-1307.
623. Brungs, W. A. 1971. Chronic effects of constant elevated temperature on the fathead minnow (Pimephales promelas). *Trans. Amer. Fish. Soc.* 100(4):659-664.



624. Allen, K. O., and K. Strawn. 1971. Rate of acclimation of juvenile channel catfish, Ictalurus punctatus, to high temperatures. Trans. Amer. Fish. Soc. 100(4):665-671.
625. Wilkes, F. G., and C. M. Weiss. 1971. The accumulation of DDT by the dragonfly nymph Tetragoneuria. Trans. Amer. Fish. Soc. 100(2):222-236.
626. Dunston, W. M., and D. W. Menzel. 1971. Continuous cultures of natural populations of phytoplankton in diluted, treated sewage effluent. Limnology and Oceanography 16(4-6):623-632.
627. Rosenthal, H., and R. Stelzer. 1970. Effects of 2,4- and 2,5-dinitrophenol on the embryological development of the herring, Clupea harengus. Marine Biology 5:325-336.
628. Nimmo, D. R., and R. R. Blackman. 1972. Effects of DDT on cations in the hepatopancreas of penaeid shrimp. Trans. Amer. Fish. Soc. 101(3):547-549.
629. Seymour, A. H., and V. A. Nelson. 1971. Biological half-lives for zinc and mercury in the Pacific oyster, Crassostrea gigas. Proceedings of the 3rd National Symposium on Radiology, Oak Ridge, TN.
630. Calabrese, A. 1972. How some pollutants affect embryos and larvae of American oyster and hard-shell clam. Marine Fisheries Review 34(11-12):66-77.
631. Calabrese, A. 1969. Effect of acids and alkalies on survival of bluegills and largemouth bass. Technical Papers of Bureau of Sport Fisheries and Wildlife, Number 42.
632. Calabrese, A., and H. C. Davis. 1957. Effects of "soft" detergents on embryos and larvae of the American oyster (Crassostrea virginica). Proceedings of National Shellfisheries Association, Volume 57.
633. Grice, G., P. H. Wiebe, and E. Hoagland. 1973. Acid-iron waste as a factor affecting the distribution and abundance of zooplankton in the New York blight. Estuarine and Coastal Marine Science 1:45-50.
634. Ukeles, R. 1968. Sulfonamide inhibition in Monochrysis lutheri. Journal of Protozoology 4(4):341-346.
635. Robertson, B., S. Arhelger, P. L. Kinney, and D. K. Button. 1973. Hydrocarbon biodegradation in Alaskan waters. Institute of Marine Sciences, University of Alaska.
636. Loosanoff, V. 1961. Effects of turbidity on some larval and adult bivalves. Proceedings of the Gulf and Caribbean Fisheries Institute, November 1961, p. 80-94.
637. Walsh, G., and T. Grow. 1971. Depression of carbohydrate in marine algae by urea herbicides. Weed Science 19(5):568-570.

638. Walsh, G. E. 1972. Effects of herbicides on photosynthesis and growth of marine unicellular algae. Hyacinth Control Journal, Volume 10.
639. Walsh, G., R. Barrett, G. Cook, and T. Hollister. 1973. Effects of herbicides on seedlings of the red mangrove, Rhizophora mangle L. BioScience 23(6):361-364.
640. Button, D. K., S. S. Dunker, and M. L. Morse. 1973. Continuous culture of Rhodotorula rubra: Kinetics of phosphate-arsenate uptake, inhibition and phosphate limited growth. Journal of Bacteriology 113(2):599-611.
641. Stadnyk, L., R. Campbell, and B. T. Johnson. 1971. Pesticide effect on growth and C-14 assimilation in freshwater algae. Bulletin of Environmental Control and Toxicology 6(1):1-8.
642. Ludke, J. L., M. T. Finley, and C. Lusk. 1971. Toxicity of mirex to crayfish Procambarus blandingi. Bulletin of Environmental Contamination and Toxicology 6(1):89-96.
643. deKoning, H. W., and D. C. Mortimer. 1971. DDT uptake and growth of Euglena gracilis. Bulletin of Environmental Control and Toxicity 6(3):244-248.
644. Young, R. G., L. St. John, and D. J. Lisk. 1971. Degradation of DDT by goldfish. Bulletin of Environmental Control and Toxicity 6(4):351-354.
645. Rehwoldt, R., G. Bida, and B. Nerrie. 1971. Acute toxicity of copper, nickel, and zinc ions to some Hudson River fish species. Bulletin of Environmental Control and Toxicity 6(5):445-448.
646. Shaw, T. L., and V. M. Brown. 1971. Heavy metals and the fertilization of rainbow trout eggs. Nature, Volume 230.
647. Lewin, J., and C. H. Chen. 1971. Available iron: A limiting factor for marine phytoplankton. Limnology and Oceanography 16:(4-6):670-675.
648. Stanley, J. G., and P. J. Colby. 1971. Effects of temperature on electrolyte balance and osmoregulation in the alewife (Alosa pseudoharengus) in fresh and sea water. Trans. Amer. Fish. Soc. 100(4):624-638.
649. Loosanoff, V. L., and F. D. Tommers. 1948. Effect of suspended silt and other substances on rate of feeding of oysters. Science, Vol. 107, Jan. 1948.
650. Ukeles, R. 1961. The effect of temperature on the growth and survival of several marine algal species. Biol. Bull. 120(2): 255-264.
651. Corner, E. D. S., and B. W. Sparrow. 1956. The modes of action of toxic agents. I. Observations on the poisoning of certain crustaceans by copper and mercury. J. Mar. Biol. U.K. 35:531-548.

652. Grzenda, A. R., W. J. Taylor, and D. F. Paris. 1971. The uptake and distribution of chlorinated residues by goldfish (Carossius auratus) fed a  $^{14}\text{C}$  Dieldrin contaminated diet. Trans. Amer. Fish. Soc. 100(2):215-221.
653. Sills, J. B., and J. L. Allen. 1971. The influence of pH on the efficacy and residues of Quinaldine. Trans. Amer. Fish. Soc. 100(3):544-545.
654. Kawatski, J. A., and J. C. Schmulback. 1971. Accumulation of insecticide in freshwater ostracods exposed continuously to sub-lethal concentrations of Aldrin or Dieldrin. Trans. Amer. Fish. Soc. No. 3.
655. Post, G., and T. R. Schroeder. 1971. The toxicity of four insecticides to four salmonid species. Bull. Envir. Cont. & Toxic. 6(2):144-155.
656. Ebel, W. J., E. M. Dawley, and B. H. Monk. 1971. Thermal tolerance of juvenile Pacific salmon and steelhead trout in relation to supersaturation of nitrogen gas. Fish. Bull. 69(4):833-843.
657. Duke, T. W., J. P. Baptist, and D. E. Hoss. 1966. Bioaccumulation of radioactive gold used as a sediment tracer in the estuarine environment. U. S. Fish & Wildlife Serv., Fish. Bull. 65:427-436.
658. Bidgood, B. F., and A. H. Berst. 1969. Lethal temperatures for Great Lakes rainbow trout. J. Fish. Res. Bd. Canada 26(2):246-459.
659. Bloom, S. A. 1970. An oil dispersant's effect on the microflora of beach sand. J. Mar. Biol. Ass. U.K. 50:919-923.
660. Reed, P. H. 1969. Culture methods and effects of temperature and salinity on survival and growth of Dungeness crab (Cancer magister) larvae in the laboratory. J. Fish. Res. Bd. Canada 26(2):389-397.
661. Peterson, R. H., and J. M. Anderson. 1969. Influence of temperature change on spontaneous locomotion activity and oxygen consumption of Atlantic salmon (Salmo salar) acclimated to two temperatures. J. Fish. Res. Bd. Canada 26:93-109.
662. Arai, M. N. 1973. Behavior of the planktonic coelenterates Sarsia tubulosa, Phialidium gregarium, and Pleurobrachia pileus, in salinity discontinuity layers. J. Fish. Res. Bd. Canada 30(8):1105-1110.
663. Davis, J. C., and B. J. Mason. 1973. Bioassay procedures to evaluate acute toxicity of neutralized bleached Kraft pulp mill effluent to Pacific salmon. J. Fish. Res. Bd. Canada 30(10):1565-1573.
664. Burrows, E. M. 1971. Assessment of pollution effects by the use of algae. Proc. Royal Soc. Lond. B. 177:295-306.
665. Chittenden, M. E. 1973. Salinity tolerance of young American shad, Alosa sapidissima. Ches. Science 14(3):207-210.

666. Kennedy, U. S., and J. A. Mihursky. 1971. Upper temperature tolerances of some estuarine bivalves. Ches. Science 12(4):193-204.
667. Salazar, M. H. 1973. Evaluation of fuel biocide toxicity. Dept. of Navy, Pasadena, CA.
668. Salazar, M. H., S. Yamamoto, W. H. Shipman, and A. R. Zirino. 1972. Informal report on lead and chromium pollution study. Dept. of Navy, Pasadena, CA.
669. Coutant, C. C. 1973. Effect of thermal shock on vulnerability of juvenile salmonids to predation. J. Fish. Res. Bd. Canada 30(7):965-973.
670. Brenko, M. H., and A. Calabrese. 1969. The combined effects of salinity and temperature on larvae of the mussel Mytilus edulis. Mar. Biol. 4:224-226.
671. Kenny, R. 1969. Temperature tolerance of the polychaete worms Diopatra cuprea and Clymonella torquata. Mar. Biol. 4:219-223.
672. Roberts, D. 1972. The assimilation and chronic effects of sub-lethal concentrations of Endosulfan on condition and spawning in the common mussel, Mytilus edulis. Mar. Biol. 16:119-125.
673. Brown, B. E., and R. C. Newell. 1972. The effect of copper and zinc on the metabolism of the mussel Mytilus edulis. Mar. Biol. 16:108-118.
674. Portmann, J. E. 1968. Progress report on a programme of insecticide analysis and toxicity testing in relation to the marine environment. Helgolander wiss Meeresunters 17:247-256.
675. Otto, R. G., and J. E. McInerney. 1970. Development of salinity preference in pre-smolt Coho salmon, Oncorhynchus kisutch. J. Fish. Res. Bd. Canada 27(4):793-800.
676. McLeese, D. W. 1970. Behavior of lobsters exposed to bleached Kraft mill effluent. J. Fish. Res. Bd. Canada 27(4):731-739.
677. Woelke, C. E. 1962. Environmental requirements of marine invertebrates. Bioassays of pulp mill wastes with oysters. Biological Problems in Water Pollution, 3rd Seminar, U.S. HEW Pub. 999-WP-25.
678. Costlow, J. D., and C. G. Bookhout. 1962. The effect of environmental factors on larval development of crabs. Biological Problems in Water Pollution, 3rd Seminar, U.S. HEW Pub. 999-WP-25.
679. Kinne, O. 1962. Salinity requirements of the fish Cyprindon maculorius. Biological Problems in Water Pollution, 3rd Seminar, U.S. HEW Pub. 999-WP-25.
680. Price, T. J. 1962. Accumulation of radionuclides and the effects of radiation on molluscs. Biological Problems in Water Pollution, 3rd Seminar, U.S. HEW Pub. 999-WP-25.

681. Bridges, W. R. 1962. Effects of time and temperature on the toxicity of Heptachlor and Kepone to redear sunfish. Biological Problems in Water Pollution, 3rd Seminar, U.S. HEW Pub. 999-WP-25.
682. Chipman, W. A., and P. S. Goltsoff. Undated. Effects of oil mixed with carbonized sand on aquatic animals. U.S. Dept. of Interior, Special Scientific Report - Fisheries #1.
683. Hoffman, C. H., and E. W. Surber. 1948. Effects of feeding DDT-sprayed insects to freshwater fish. U.S. Dept. of Interior, Special Scientific Report - Fisheries #3.
684. Surber, E. W., and C. Hoffman. Undated. Effects of various concentrations of DDT on several species of fish of different sizes. U.S. Dept. of Interior, Special Scientific Report - Fisheries #4.
685. Davidson, R. C., W. P. Breeze, C. E. Warren, and P. Doudoroff. 1959. Experiments on the dissolved oxygen requirements of cold water fishes. Sewage & Ind. Wastes 31:950-966.
686. Pringle, B. H., D. E. Hissong, E. L. Katz, and S. T. Mulawka. 1968. Trace metal accumulation by estuarine molluscs. J. Sanit. Eng. Div. SA3:455-475.
687. Burton, D. T., and P. R. Abell. 1973. Ventilation rate changes in the flatfish Trinectes maculatus, subjected to rapid thermal increases and environmental hypoxia. Assoc. S.E. Biol. Bull. 20(2):43.
688. Abell, P. R., and D. T. Burton. 1973. Changes in oxygen consumption of the mud crab, Rhithropanopeus harrisii, following rapid thermal stress. Assoc. S.E. Biol. Bull. 20(2):35.
689. Lowe, C. H., D. S. Hinds, and E. A. Halpern. 1967. Experimental catastrophic selection and tolerances to low oxygen concentration in native Arizona freshwater fishes. Ecology 48(6):1013-1017.
690. Moore, W. G., and A. Burn. 1968. Lethal oxygen thresholds for certain temporary pond invertebrates and their applicability to field stations. Ecology 49(2):349-351.
691. Lutz, P. 1968. Effects of temperature and photoperiod on larval development in Lestes eurinus. Ecology 49(4):637-644.
692. Wells, M. M. 1918. The reactions and resistance of fishes to carbon dioxide and carbon monoxide. Bull. Ill. State Lab. of Nat. Hist. 11(8):557-571.
693. Scheier, A., and J. Cairns. 1968. An apparatus for estimating the effects of toxicants on the critical flicker frequency response of the bluegill sunfish. Proc. 23rd Ind. Waste Conf., Purdue, Univ. Eng. Ext. Serv. 132.
694. Hoff, J., M. E. Chittenden, and J. R. Westman. 1966. Oxygen requirements of some marine and anadromous fishes with particular reference to problems of measurement. Proc. 21st Ind. Waste Conf., Purdue Univ. Eng. Ext. Ser. 121.

695. Grosch, D. S. 1973. Reproduction tests: The toxicity for Artemia of derivatives from non-persistent pesticides. Biol. Bull. 145: 340-351.
696. Brungs, W. A., and G. W. Bailey. 1966. Influence of suspended solids on the acute toxicity of Endrin to fathead minnows. Proc. 21st Ind. Waste Conf., Purdue, Univ. Eng. Ext. Ser. 121.
697. Cairns, J., Jr., and A. Scheier. 1958. The effects of periodic low oxygen upon the toxicity of various chemicals to aquatic organisms. Proc. 12th Ind. Waste Conf., 1957, Purdue Univ. Eng. Ext. Serv. 94.
698. Anderson, B. G. 1948. The apparent thresholds of toxicity to Daphnia magna for chlorides of various metals when added to Lake Erie water. Trans. Amer. Fish. Soc. 78:96-113.
699. Gilfallin, E. Undated. Reactions of Euphasia pacifica Hansen (Crustacea) from oceanic, mixed oceanic-coastal, and coastal waters of British Columbia, to experimental changes in temperature and salinity. J. Exp. Mar. Biol. Ecol. 10:29-40.
700. Allen, K. O., and K. Strawn. 1968. Heat tolerance of channel catfish. Proc. 21st Conf. S.E. Game Fish Comm. 1967:399-411.
701. Burdick, G. E., and M. Lipschuetz. 1948. Toxicity of ferro- and ferric cyanide solutions to fish, and determinations of the cause of mortality. Trans. Amer. Fish. Soc. 78:192-202.
702. Van Horn, W. M., J. B. Anderson, and M. Katz. 1949. The effect of Kraft pulp mill wastes on some aquatic organisms. Trans. Amer. Fish. Soc. 79:55-63.
703. Whitmore, C. M., C. E. Warren, and P. Doudoroff. 1960. Avoidance reactions of Salmonid and Centrarchid fishes to low oxygen concentrations. Trans. Amer. Fish. Soc. 89:17-26.
704. Katz, M., A. Pritchard, and C. E. Warren. 1959. Ability of some salmonids and a centrarchid to swim in water of reduced oxygen content. Trans. Amer. Fish. Soc. 88:88-95.
705. Marking, L. L. 1960. Salicylanilide I, an effective, nonpersistent candidate piscicide. Trans. Amer. Fish. Soc. 101(3):526-533.
706. Nebeker, A. W. 1960. Effect of low oxygen concentration on survival and emergence of aquatic insects. Trans. Amer. Fish. Soc. 101:675-679.
707. Bell, H. L., and A. U. Nebeker. 1969. Preliminary studies on the tolerance of aquatic insects to low pH. J. Kans. Entomol. Soc. 42:230-236.
708. Hughes, J. S. 1969. Toxicity of some chemicals to striped bass (Roccus saxatilis). Proc. 22nd Conf. S. E. Game Fish. Comm. 1968:230-234.

709. Hughes, J. S. 1967. Use of the red crawfish, Procambarus clarki (Girard), for herbicidal assays. Proc. 20th Conf. S. E. Game Fish. Comm. 1966:437-439.
710. Allen, K. O. 1969. Heat tolerance of albino vs. normal channel catfish. Proc. 22nd Conf. S. E. Game Fish. Comm. 1968:267-270.
711. Bacon, E. J., Jr., W. H. Neill, Jr., and R. V. Kilambi. 1968. Temperature selection and heat resistance of the mosquito-fish (Gambusia affinis). Proc. 21st Conf. S. E. Game Fish. Comm. 1967: 411-416.
712. Lozacano, H. 1968. Some effects of salinity on two populations of red swamp crawfish, Procambarus clarki (Girard). Proc. 21st Conf. S. E. Game Fish. Comm. 1967:423-435.
713. Shannon, E. H., and W. B. Smith. 1968. Preliminary observations of the effect of temperature on striped bass eggs and sac fry. Proc. 21st Conf. S. E. Game Fish. Comm. 1967:257-260.
714. Angelovic, J. W., and D. W. Engel. 1970. Effects of radiation on estuarine organisms. Mar. Poll. Bull. 1(7):103-105.
715. Scott, D. M., and C. W. Major. 1972. The effect of copper (II) on survival, respiration, and heart rate in the common blue mussel (Mytilus edulis). Biol. Bull. 143:679-688.
716. Williams, L. G., and Q. Pickering. 1961. Direct and food chain uptake of cesium<sup>137</sup> and strontium<sup>85</sup> in bluegill fingerlings. Ecol. 42:205-206.
717. Kennedy, V. S., and J. A. Mihursky. 1972. Effects of temperature on the respiratory metabolism of three Chesapeake Bay bivalves. Ches. Sci. 13(1):1-22.
718. Brown, B., and M. Ahsanullah. 1971. Effect of heavy metals on mortality and growth. Mar. Poll. Bull. 2(12):182-187.
719. Saksena, V. P., and E. P. Joseph. 1972. Dissolved oxygen requirements of newly hatched larvae of the striped blenny, (Chasmodes bosquianus), the naked goby, (Gobiosoma bosci), and the skillet-fish (Gobiesox strumosus).
720. D'Aoust, B. G. 1969. Hyperbaric oxygen: toxicity to fish at pressures present in their swimbladders. Sci. 163:576-578.
721. Larsen, F. C., and R. J. Staub. 1972. Effects of industrial effluents on primary phytoplankton indicators. Water Resources Research Center. Research Report No. 26.
722. McLean, R. I. 1972. Chlorine tolerance of the colonial hydroid, Bimeria franciscana. Ches. Sci. 13(3):229-230.
723. Moore, D. J. 1971. The uptake and concentration of fluoride by the blue crab, Callinectes sapidus. Ches. Sci. 12(1):1-13.
724. Tagatz, M. E. 1971. Osmoregulatory ability of blue crabs in different temperature-salinity combinations. Ches. Sci. 12(1): 14-17.

725. Haefner, P. A., Jr. 1971. Avoidance of anoxic conditions by the sand shrimp, Crangon septemspinosa. Ches. Sci. 12(1):50-51.
726. Bonn, E. W., and B. G. Follis. 1967. Effects of hydrogen sulfide on channel catfish (Ictalurus punctatus). Proc. 20th Conf. S. E. Game Fish. Comm. 1966:424-432.
727. Bohle, B. 1970. Effects of adaptation to reduced salinity on filtration activity and growth of mussels (Mytilus edulis L.). J. Exp. Mar. Biol. Ecol. 10:41-47.
728. Schulz-Baldes, M. 1972. Toxicity and accumulation of lead in the common mussel (Mytilus edulis), in laboratory experiment. Marine Biology 16:226-229.
729. Pyefinch, K. A., and J. C. Mott. 1948. The sensitivity of barnacles and their larvae to copper and mercury. J. Exp. Biol. 25: 276-298.
730. Jones, J. R. E. 1948. A further study of the reactions of fish to toxic solutions. J. Exp. Biol. 25:22-34.
731. Brungs, W. A. 1969. Chronic toxicity of zinc to the fathead minnow, Pimephales promelas Rafinesque.
732. Miettinen, J. K., M. Heyrand and S. Keckes. 1972. Mercury as a hydrospheric pollutant. II. Biological half-time of methyl mercury in four Mediterranean species: a fish, a crab, and two molluscs.
733. Miettinen, V., E. B. Blankenstein, K. Rissanen, M. Tillander, J. K. Miettinen, and M. Valtonen. 1972. Preliminary study on the distribution and effects of two chemical forms of methyl mercury in pike and rainbow trout. In: Marine Pollution and Sea Life, FAO, Fishing News Ltd., England.
734. Bouch, G. R., and R. C. Bull. 1965. Influence of a diurnal oxygen pulse on fish serum proteins. Trans. Amer. Fish. Soc. 94: 363-370.
735. Hannan, P. J., and C. Patouillet. 1972. Nutrient and pollutant concentrations as determinants in algal growth rates. In: Marine Pollution and Sea Life, FAO, Fishing News LTD, England.
736. Winter, J. E. 1972. Long-term laboratory experiments on the influence of ferric hydroxide flakes on the filter feeding behaviour, growth, iron content, and mortality in Mytilus edulis L. In: Marine Pollution and Sea Life, FAO, Fishing News Ltd., England.
737. Westfall, B. A. 1945. Coagulation film anoxia in fishes. Ecol. 26:283-287.
738. King, S. F. 1964. Uptake and transfer of cesium-137 by Chlamydomonas, Daphnia, and bluegill fingerlings. Ecol. 45:852-859.



739. Hubbs, C. 1964. Effects of thermal fluctuations on the relative survival of greenthroat darter young from stenothermal and eurythermal waters. *Ecol.* 45:376-379.
740. Claffey, F. J., and J. E. Ruck. 1967. The effect of rotenone on certain fish-food organisms. *Proc. 20th Conf. S. E. Game Fish. Comm.* 1966:278-283.
741. Gatz, A. J., Jr., V. S. Kennedy, and J. A. Mihursky. 1973. Effects of temperature on activity and mortality of the scyphozoan medusa (Chrysaora quinquecirrha). *Ches. Sci.* 14(3):171-180.
742. Burrows, E. M., and C. Pybus. 1971. Laminaria saccharina and marine pollution in north east England. *Mar. Poll. Bull.* 2(4): 53-56.
743. Armitage, K. B., and L. J. Olund. 1962. Salt tolerance of the brook stickleback. *Amer. Midl. Nat.* 68:274-277.
744. Malone, C. R., and B. G. Blaylock. 1970. Toxicity of insecticide formulations to carp embryos reared in vitro. *J. Wildlife Manage.* 34(2):460-463.
745. Parker, B. L., J. E. Deney, and C. A. Bache. 1970. Carbamate bioassay using Daphnia magna. *J. Econ. Entomol.* 63(3):710-714.
746. Finley, M. T., D. E. Ferguson, and J. L. Ludke. 1970. Possible selective mechanisms in the development of insecticide-resistant fish. *Pestic. Monit. J.* 3(4):212-218.
747. Hilden, S., and A. C. Giese. 1969. Effect of salt concentration on regeneration rate in Blepharisma acclimated to high salt levels. *J. Protozool.* 16(3):419-422.
748. Roy, A. W., and P. H. Johansen. 1970. The temperature selection of small hypophysectomized goldfish (Carassius auratus L.) *Can. J. Zool.* 48:323-326.
749. Jackson, D. A., J. M. Anderson, and D. R. Gardner. 1970. Further investigations of the effect of DDT on learning in fish. *Can. J. Zool.* 48:577-580.
750. Arnold, D. C. 1970. A tidal rhythm in the response of the barnacle, Balanus balanoides, to water of diminished salinity. *J. Mar. Biol. Assn. U. K.* 50:1045-1055.
751. Bardach, J. E., M. Fujya, and A. Hall. 1965. Detergents: effects on the chemical senses of the fish Ictalurus natulis (le Sueur). *Science* 148:1605-1607.
752. Woelke, C. E. 1967. Measurement of water quality with the Pacific oyster embryo bioassay. *ASTM Spec. Tech. Publication* 416.
753. Smith, L. L., Jr., and D. M. Oseid. 1970. Toxic effects of hydrogen sulfide to juvenile fish and fish eggs. *Proc. 25th Ind. Waste Conf., Purdue Univ., Eng. Ext. Ser.* 137:739-744.

754. Jamnaback, H. 1962. An electric method of testing the effectiveness of chemicals in killing blackfly larvae (Simuliida dysteri). Mosq. News 22:384-389.
755. Lewallen, L. L., and W. H. Wilder. 1962. Toxicity of certain organophosphorus and carbamate insecticides to rainbow trout. Mosq. News 22:369-372.
756. Lewallen, L. L. 1962. Toxicity of certain insecticides to hydrophilid larvae. Mosq. News 22:112-113.
757. Hilsenhoff, W. L. 1962. Toxicity of granular malathion to walleyed pike fingerlings. Mosq. News 22:14-15.
758. Sullivan, C. M., and K. C. Fisher. 1954. The effects of light on temperature selection in speckled trout (Salvelinus fontinalis [Mitchill]). Biol. Bull. 107:278-288.
759. Trama, F. B. 1965. The acute toxicity of phenol to the common bluegill. Not. Nat. 269, 10 p.
760. Trama, F. B. 1954. The acute toxicity of copper to the common bluegill (Lepomis macrochirus Rafinesque). Not. Nat. 257, 13 p.
761. Trama, F. B. 1954. The pH tolerance of the common bluegill (Lepomis macrochirus Rafinesque). Not. Nat. 256, 13 p.
762. Haefner, P. A., Jr. 1970. The effect of low dissolved oxygen concentrations on temperature-salinity tolerance of the sand shrimp, Crangon septemspinosa Say. Physiol. Zool. 388-397.
763. Haefner, P. A., Jr. 1969. Temperature and salinity tolerance of the sand shrimp, Crangon septemspinosa Say. Physiol. Zool. 388-397.
764. Teal, J. M., and F. G. Carey. 1967. The metabolism of marsh crabs under conditions of reduced oxygen pressure. Physiol. Zool. 40:83-91.
765. Augenfeld, J. M. 1967. Effects of oxygen deprivation on aquatic midge larvae under natural and laboratory conditions. Physiol. Zool. 40:149-156.
766. Whitten, B. K., and C. J. Goodnight. 1967. The accumulations of Sr-89 and Ca-45 by an aquatic oligochaete. Physiol. Zool. 40:371-385.
767. Allen, H. 1971. Effects of petroleum fractions on the early development of a sea urchin. Mar. Poll. Bull. 2(9):138-140.
768. McCanley, R. W., and L. A. A. Read. 1973. Temperature selection by juvenile and adult yellow perch (Perca flavescens) acclimated to 24C. J. Fish. Res. Bd. Canada 20:1253-1255.
769. Wisely, B., and R. A. P. Blick. 1967. Mortality of marine invertebrate larvae in mercury, copper and zinc solutions. Aust. J. Mar. E-W Res. 18:63-72.

770. Doudoroff, P. 1938. Reactions of marine fishes to temperature gradients. Biol. Bull. 75:494-509.
771. Rucker, R. R., and D. F. Amend. 1969. Absorption and retention of organic mercurials by rainbow trout and chinook and sockeye salmon. Prog. Fish-Cult. 31:197-201.
772. Watabe, N., and K. M. Wilbur. 1966. Effects of temperature on growth, calcification, and coccolith form in Coccolithus huxleyi (Coccolithineae). Limnol. and Oceanogr. 11(4):567-575.
773. Amend, D. F. 1970. Retention of mercury by salmon. Prog. Fish-Cult. 32:192-194.
774. Elliot, J. W. 1969. The oxygen requirements of chinook salmon. Prog. Fish-Cult. 31:67-73.
775. Wellborn, T. L., Jr. 1971. Toxicity of some compounds to striped bass fingerlings. Prog. Fish-Cult. 33:32-35.
776. Boyd, C. E., and D. E. Ferguson. 1964. Susceptibility and resistance of mosquito fish to several insecticides. J. Econ. Entomol. 57(4):430-431.
777. Jones, J. R. E. 1939. The relation between the electrolytic solution pressures of the metals and their toxicity to the stickleback (Gasterosteus aculeatus L.). J. Exp. Biol. 16:425-437.
778. Rodgers, E. O., B. H. Hagen, S. B. Friddle, and S. F. Snieszko. 1951. The toxicity of pyridylmercuric acetate technical (PMA) to rainbow trout (Salmo gairdneri). Prog. Fish-Cult. 12:71-73.
779. Lawrence, J. M. 1950. Toxicity of some new insecticides to several species of pondfish. Prog. Fish-Cult. 12:141-146.
780. Wells, G. P., and Isabel C. Ledingham. 1940. Physiological effects of a hypotonic environment. I. The action of hypotonic salines on isolated rhythmic preparations from polychaete worms (Arenicola marina, Nereis diversicolor, Perineris cultifera) J. Exp. Biol. 17:337-352.
781. Jones, J. R. E. 1947. The oxygen consumption of Gasterosteus aculeatus L. in toxic solutions. J. Exp. Biol. 23:298-311.
782. Parry, G. D. R., and J. Hayward. 1973. The uptake of <sup>65</sup> Zn by Dunaliella tertiolecta Butcher. J. Mar. Biol. Assn. U. K. 53: 915-922.
783. Bryan, G. W., and L. G. Hummerstone. 1973. Adapting of the polychaete Nereis diversicolor to manganese in estuarine sediments. J. Mar. Biol. Assn. U. K. 859-872.
784. Bryan, G. W., and L. G. Hummerstone. 1973. Adapting of the polychaete Nereis diversicolor to estuarine sediments containing high concentrations of zinc and cadmium. J. Mar. Biol. Assn. U. K. 53:839-857.

785. Keenan, J. D. 1973. Response of Anabaena to pH, carbon, and phosphorus. J. Envir. Eng. Div., Proc. ASCE, EE5:607-620.
786. Howard, T. E., and C. C. Walden. 1965. Pollution and toxicity characteristics of Kraft pulp mill effluents. TAPPI 48:136-141.
787. Gould, W. R., and T. C. Dorris. 1961. Toxicity changes of stored oil refinery effluents. Sew. Ind. Wastes 33:1107-1111.
788. Lee, E. L., and J. C. Buzzell, Jr. 1969. Measurements of pesticide toxicity by fish respiration rate. Proc. 24th Ind. Waste Conf., Purdue Univ., Eng. Ext. Ser. 135:595-609.
789. Steeman, N. E., L. Kamp-Nielsen, and S. Winn-Anderson. 1969. The effect of deleterious concentrations of copper on the photosynthesis of Chlorella pyrenoidosa. Physiol. Plant. 22:1121-1133.
790. Pickering, D. C., and I. C. Puia. 1969. Mechanism for the uptake of zinc by Fontinalis pantipyretica. Physiol. Plant. 22:653-661.
791. Gramlich, J. V., and R. E. Frans. 1964. Kinetics of Chlorella inhibition by herbicides. Weeds 12(3):184-189.
792. Price, C. A., and M. T. G. Estrada. 1964. Chlorophyll formation in Euglena as a test for herbicides. Weeds 12(3):234-235.
793. Funderburk, H. H., Jr., and J. M. Lawrence. 1964. Mode of action and metabolism of diquat and paraquat. Weeds: 12(4):259-264.
794. Walker, C. M. 1964. Dichlobenil as a herbicide in fish habitats. Weeds 12(4):267-269.
795. Blackburn, R. D., and L. W. Weldon. 1965. The sensitivity of duckweeds (Lemnaceae) and Azolla to diquat and paraquat. Weeds 13(2):147-149.
796. Walker, C. M. 1964. Diuron, tenuron, monuron, neburon, and TCA mixtures as aquatic herbicides in fish habitats. Weeds 13(4):297-301.
797. Haight, J. J., and R. Y. Morita. 1966. Some physiological differences in Vibrio marinus grown at environmental and optimal temperatures. Limnol. and Oceanogr. 11(4):470-474.
798. Nunogawa, J. N., N. C. Burbank, and R. H. F. Young. 1970. The relative toxicities of selected chemicals to several species of tropical fish. Advances in Water Pollution Research, Vol. 2, 1970.
799. Johnson, J. M., O. R. Ruschmeyer, T. O. Odlang, and Y. A. Olson. 1970. Algal bioassay potential primary productivity studies of the lower St. Louis River, Minnesota. Advances in Water Pollution Research, Vol. 2, 1970.
800. Sanders, H. O., and O. B. Cope. 1968. The relative toxicities of several pesticides to naiads of three species of stoneflies. Limnol. and Oceanogr. 13(1):112-117.

801. Jensen, A. L. 1972. Variation in results of identical bioassays of minnows subjected to instant temperature increase. Trans. Amer. Fish. Soc. 101(3):403-407.
802. Goodyear, C. P. 1972. A simple technique for detecting effects of toxicants or other stresses on a predator-prey interaction. Trans. Amer. Fish. Soc. 101(2):367-370.
803. Beamish, R. J. 1972. Lethal pH for the white sucker, Catostomus commersoni (Lacepede). Trans. Amer. Fish. Soc. 101(2):355-358.
804. Hansen, D. J. 1972. DDT and malathion: effect on salinity selection by mosquitofish. Trans. Amer. Fish. Soc. 101(2):346-350.
805. Derr, D. K., and M. J. Zabik. 1972. Biologically active compounds in the aquatic environment: the uptake and distribution of [1, 1-dichloro-2, 2-bis (p-chlorophenyl) ethylene] DDE by Chironomus tentans (Diptera: Chironomidae). Trans. Amer. Fish. Soc. 101(2):323-329.
806. McCann, J. A., and R. L. Jasper. 1972. Vertebral damage to bluegills exposed to acutely toxic levels of pesticides. Trans. Amer. Fish. Soc. 101(2):317-322.
807. Merna, J. W., M. E. Bender, and J. R. Nory. 1972. The effects of methoxychlor on fishes. I. Acute toxicity and breakdown studies. Trans. Amer. Fish. Soc. 101(2):298-301.
808. Chittenden, M. E., Jr. 1972. Responses of young American shad, Alosa sapidissima, to low temperatures. Trans. Amer. Fish. Soc. 101(4):680-685.
809. Saksena, V. P., C. Steinmetz, Jr., and E. D. Honde. 1972. Effects of temperature on growth and survival of laboratory-reared larvae of the scaled sardine, Harengula pensacolatae. Goode and Bann. Muskingum College, New Concord, OH.
810. Chittenden, M. E., Jr. 1972. Salinity tolerance of young blue-back herring, Alosa aestivalis. Trans. Amer. Fish. Soc. 101(1):123-125.
811. Wilson, W. J. 1974. The effects of concentration and particle size of suspended materials on growth and conditioning of the Pacific oyster (Crassostrea gigas). M. S. Thesis, Oregon State Univ., Corvallis, Oregon.
812. Poorman, A. E. 1973. Effects of pesticides on Euglena gracilis. I. Growth studies. Bull. Envir. Contam. Toxicol. 10(1):25-28.
813. Petrocelli, S. R., A. R. Hanks, and J. Anderson. 1973. Uptake and accumulation of an organochlorine insecticide (Dieldrin) by an estuarine mollusc, Rangia cuneata. Bull. Envir. Contam. Toxicol. 10(5):315-320.
814. Naqui, S. M., and A. A. de la Cruz. 1973. Mirex incorporation in the environment. Toxicity in selected fresh-water organisms. Bull. Envir. Contam. Toxicol. 10(5):305-308.

815. Slonim, C. B., and A. R. Slonim. 1973. Effect of water hardness on the tolerance of the guppy to beryllium sulfate. Bull. Envir. Contam. Toxicol. 10(5):295-301.
816. Rehwoldt, R., L. Lasko, C. Shaw, and E. Wirhowski. 1973. The acute toxicity of some heavy metal ions toward benthic organisms. Bull. Envir. Contam. Toxicol. 10(5):291-294.
817. Zitko, P., W. V. Carson, and W. G. Carson. 1973. Prediction of incipient lethal levels of copper to juvenile Atlantic salmon in the presence of humic acid by cupric electrode. Bull. Envir. Contam. Toxicol. 10(5):265-271.
818. Marvin, D. E., and D. T. Burton. 1973. Cardiac and respiratory responses of rainbow trout, bluegills, and brown bullhead catfish during rapid hypoxia and recovery under normoxic conditions. Comp. Biochem. Physiol. 46A:755-765.
819. Rogers, B. A. 1969. Tolerance levels of 4 species of estuarine fishes to suspended mineral solids. M. S. Theses, Univ. of Rhode Island, Kingston, R. I.
820. Kamp-Nielsen, L. 1971. The effect of deleterious concentrations of mercury on the photosynthesis and growth of Chlorella pyrenoidosa. Physiol. Plant. 24:556-561.
821. Steeman, N. E., and S. Wiium-Andersen. 1971. The influence of Cu on photosynthesis and growth in diatoms. Physiol. Plant. 24:480-484.
822. Chaston, I. 1969. Anaerobiosis in Cyclops varians. Limnol. Oceanogr. 14(2):298-301.
823. Paasche, E. 1968. The effect of temperature, light intensity, and photoperiod on coccolith formation. Limnol. Oceanogr. 13(1):178-181.
824. Phinney, H. K., and C. D. McIntire. 1965. Effect of temperature on metabolism of periphyton communities developed in a laboratory stream. Limnol. Oceanogr. 10(3):341-344.
825. Jones, G. E. 1967. Growth of Escherichia coli in heat- and copper treated synthetic sea water. Limnol. Oceanogr. 13(1):167-172.
826. Kelley, J. W. 1968. Effects of incubation temperature on survival of largemouth bass eggs. Prog. Fish-Cult. 30:159-163.
827. Johnson, M. W. 1967. Some observations on the hatching of Tortanus discaudatus eggs subjected to low temperatures. Limnol. Oceanogr. 12(3):405-410.
828. Mount, D. I., and C. E. Stephan. 1967. A method for detecting cadmium poisoning in fish. J. Wildlife Manage. 31(1):168-172.
829. Lloyd, R., and D. H. M. Jordan. 1964. Some factors affecting the resistance of rainbow trout (Salmo gairdneri Richardson) to acid waters. Int. J. Air Wat. Poll. 8:393-403.

830. Cairns, J. Jr., W. T. Waller, and J. C. Smrchek. 1969. Fish bioassays contrasting constant and fluctuating input of toxicants. *Rev. de Biologie* 7(1-2):75-91.
831. Environmental Protection Agency/Corps of Engineers. 1977. Ecological Evaluation of Proposed Discharge of Dredged Material into Ocean Waters, July 1977, U. S. Army Engineer Waterways Experiment Station, CE, Vicksburg, MS.
832. Emerson, R. R. 1974. Preliminary Investigations of the Effects of Resuspended Sediment on Two Species of Benthic Polychaetes from Los Angeles Harbor, Marine Studies of San Pedro Bay, California, Part III, pp. 97-110.
833. Hoss, D. E., et al. 1974. Effects of Seawater Extracts of Sediments from Charleston Harbor, South Carolina, on Larval Estuarine Fishes, Estuarine and Coastal Marine Science, Vol. 2, pp. 223-238.
834. Shuba, et al. 1978. Biological Assessment Methods to Predict the Potential Environmental Impact of Open-Water Disposal of Dredged Material, Technical Report (In preparation), U. S. Army Engineer Waterways Experiment Station, CE, Vicksburg, MS.
835. Shuba, P. J., J. H. Carroll, and K. L. Wong. 1977. Biological Assessment of the Soluble Fraction of the Standard Elutriate Test, Technical Report D-77-3, March 1977, U. S. Army Engineer Waterways Experiment Station, CE, Vicksburg, MS.
836. Corps of Engineers. 1976. Ecological Evaluation of Proposed Discharge of Dredged or Fill Material into Navigable Waters, Miscellaneous Paper D-76-17, May 1976, U. S. Army Engineer Waterways Experiment Station, CE, Vicksburg, MS.
837. Lee, G. F., et al. 1975. Research Study for the Development of Dredged Material Disposal Criteria, Contract Report D-75-4, U. S. Army Engineer Waterways Experiment Station, CE, Vicksburg, MS.
838. Lee, G. F., et al. 1978. The Development of Criteria for Dredged Material Disposal, Contract Report (In preparation), U. S. Army Engineer Waterways Experiment Station, CE, Vicksburg, MS.
839. Wright, T. 1978. Task 1A - Synthesis Report: Aquatic Dredged Material Disposal Impacts, Technical Report (In preparation), U. S. Army Engineer Waterways Experiment Station, CE, Vicksburg, MS.
840. Holliday, B. W. 1978. Task 1B - Synthesis Report: Processes Affecting the Fate of Dredged Materials, Technical Report (In preparation), U. S. Army Engineer Waterways Experiment Station, CE, Vicksburg, MS.
841. Gannon, J. E., and Beeton, A. M. 1969. Studies on the Effects of Dredged Materials from Selected Great Lakes Harbors on Plankton and Benthos, Special Report No. 8, Center for Great Lakes Studies, Milwaukee, Wisconsin.

842. Swartz, R. C., W. A. DeBen, and F. A. Cole. In preparation. A Bioassay for the Toxicity of Sediment to the Marine Macrobenthos. Submitted to Jour. Water. Poll. Control Fed.



Table 1  
Commonly Used Species, Reported in Bioassay Literature

Species	Habitat	Reference Numbers
<b>PLANTS</b>		
<u>Chlamydomonas</u> sp. (green)	Marine, planktonic	458, 549, 583, 597, 605, 650, 721
<u>Chlorella pyrenoidosa</u> (green)	Freshwater, planktonic	480, 584, 589, 735, 789, 791, 820, 821
<u>Chlorella</u> sp. (green)	Marine, planktonic	549, 572, 605, 650
<u>Chlorococcum</u> sp. (green)	Marine, planktonic	605, 637, 638, 650
<u>Cyclotella nana</u> (diatom)	Marine, planktonic	542, 556, 573, 575, 576, 605, 647
<u>Dunaliella tertiolecta</u> (green)	Marine, planktonic	556, 576, 577, 605, 637, 638, 668, 782
<u>Euglena gracilis</u> (green)	Freshwater, planktonic	539, 643, 792, 812
<u>Isochrysis galbana</u> (diatom)	Estuarine, planktonic	575, 576, 605, 637, 638, 647, 650
<u>Monochrysis lutheri</u> (diatom)	Marine, planktonic	542, 572, 605, 634, 637, 650
<u>Nitzschia closterium</u> ( <u>Phaeodactylum tricornutum</u> ) (diatom)	Marine, planktonic	285, 458, 542, 549, 556, 572, 589, 597, 605, 638, 650, 735
<u>Skeletonema costatum</u> (diatom)	Marine, planktonic	285, 542, 549, 556, 573, 575, 576, 583, 626, 647
<b>MOLLUSCS</b>		
<u>Crassostrea gigas</u> (oyster)	Marine, estuarine	301, 485, 535, 558, 562, 629, 677, 686, 752, 811
<u>C. virginica</u> (oyster)	Marine, estuarine	442, 470, 487, 521, 564, 604, 630, 632, 636, 649, 657, 680, 682, 686
<u>Mercenaria mercenaria</u> (quahog)	Marine	286, 442, 471, 487, 516, 521, 582, 604, 629, 636, 680, 682, 686
<u>Mya arenaria</u> (soft-shell clam)	Marine, estuarine	381, 470, 590, 591, 666, 686, 717
<u>Mytilus edulis</u> (mussel)	Marine, estuarine	286, 301, 381, 414, 512, 515, 535, 536, 547, 552, 555, 560, 578, 579, 586, 590, 591, 667, 668, 670, 672, 673, 715, 727, 728, 736
<u>Physa heterostrophia</u> (snail)	Freshwater	386, 438, 616, 697
<b>ANNELIDS</b>		
<u>Limnodrilus hoffmeisteri</u> (tubificid)	Freshwater	551, 566, 616, 766
<b>ARTHROPODS (Crustaceans)</b>		
<u>Artemia salina</u> (brine shrimp)	Estuarine	492, 581, 651, 695, 718, 769
<u>Balanus balanoides</u> (barnacle)	Marine	512, 515, 536, 601, 682, 729, 750
<u>Callinectes sapidus</u> (blue crab)	Marine, estuarine	468, 470, 612, 657, 678, 714, 723, 724
<u>Crangon septemspinosa</u> (sand shrimp)	Marine, estuarine	381, 512, 725, 762, 763
<u>Daphnia magna</u> (waterflea)	Freshwater	350, 440, 463, 500, 501, 567, 698, 738, 745
<u>D. pulex</u> (waterflea)	Freshwater	297, 353, 524, 738
<u>Gammarus pseudolimnaeus</u> (amphipod)	Freshwater	348, 359, 379, 432
<u>Homarus americanus</u> (lobster)	Marine	286, 341, 482, 483, 559, 676
<u>Palaemonetes kadiakensis</u> (shrimp)	Freshwater	343, 463, 567, 814
<u>P. pugio</u> (grass shrimp)	Marine, estuarine	534, 594, 606, 612, 714
<u>P. vulgaris</u> (grass shrimp)	Marine, estuarine	286, 333, 381, 612
<u>Penaeus aztecus</u> (brown shrimp)	Marine	570, 612, 628
<u>P. duorarum</u> (pink shrimp)	Marine	557, 570, 612, 628
<b>ARTHROPODS (Insects)</b>		
<u>Acroneuria lyctorias</u> (stonefly)	Freshwater	486, 706, 707
<u>A. pacifica</u> (stonefly)	Freshwater	327, 384, 385, 396, 488
<u>Ephemerella subvaria</u> (mayfly)	Freshwater	486, 706, 707
<u>Hydropsyche betteni</u> (caddisfly)	Freshwater	486, 706, 707
<u>Pteronarcys californica</u> (stonefly)	Freshwater	327, 384, 385, 396, 488, 800
<b>FISHES</b>		
<u>Carassius auratus</u> (goldfish)	Freshwater	303, 304, 321, 328, 335, 342, 353, 354, 358, 363, 407, 411, 443, 448, 449, 452, 453, 495, 497, 502, 503, 517, 529, 585, 644, 652, 705, 748, 779, 788, 830,
<u>Catostomus commersoni</u> (white sucker)	Freshwater	280, 295, 342, 349, 352, 380, 495, 530, 705, 753, 803
<u>Cyprinodon variegatus</u> (sheepshead minnow)	Estuarine	305, 311, 314, 368, 381, 458, 540, 570, 595, 819
<u>Cyprinus carpio</u> (carp)	Freshwater	321, 326, 334, 342, 361, 375, 380, 495, 525, 645, 705, 744
<u>Esox lucius</u> (Northern pike)	Freshwater	323, 342, 346, 349, 353, 705, 733, 753
<u>Fundulus heteroclitus</u> (mummichog)	Estuarine, marine	286, 291, 315, 330, 381, 455, 475, 491, 714, 819
<u>Gambusia affinis</u> (mosquitofish)	Freshwater, estuarine	311, 325, 339, 351, 361, 363, 392, 459, 522, 527, 603, 711, 746, 776, 798, 802, 804

(Continued)

Table 1 (Concluded)

Species	Habitat	Reference Numbers
FISHES (Continued)		
<u>Gasterosteus aculeatus</u> (threespine stickleback)	Estuarine, freshwater	301, 351, 356, 399, 465, 730, 777, 781
<u>Ictalurus melas</u> (black catfish)	Freshwater	304, 321, 334, 342, 380, 392, 705
<u>I. natalis</u> (yellow bullhead)	Freshwater	296, 352, 734, 746, 751
<u>I. punctatus</u> (channel catfish)	Freshwater	300, 304, 321, 324, 342, 392, 398, 538, 571, 624, 700, 705, 710, 725, 726
<u>Lagodon rhomboides</u> (pinfish)	Marine	499, 570, 612, 618
<u>Leiostomus xanthurus</u> (spot)	Marine	311, 355, 464, 507, 618
<u>Lepomis cyanellus</u> (green sunfish)	Freshwater	299, 304, 324, 334, 342, 354, 380, 392, 444, 448, 453, 620, 692, 705, 746
<u>L. gibbosus</u> (pumpkinseed)	Freshwater	295, 436, 525, 645, 794
<u>L. macrochirus</u> (bluegill)	Freshwater	281, 288, 299, 301, 309, 312, 313, 320, 321, 324, 328, 334, 336, 342, 349, 351, 352, 353, 354, 357, 362, 365, 366, 369, 370, 386, 388, 389, 392, 398, 408, 411, 433, 434, 435, 437, 438, 440, 443, 448, 451, 452, 457, 481, 489, 490, 497, 503, 509, 517, 518, 529, 543, 615, 631, 683, 684, 693, 697, 703, 705, 716, 726, 734, 737, 759, 760, 761, 779, 794, 796, 806, 818, 819
<u>Micropterus dolomieu</u> (smallmouth bass)	Freshwater	299, 352, 503, 683, 684, 705
<u>M. salmoides</u> (largemouth bass)	Freshwater	295, 321, 324, 334, 352, 354, 357, 358, 369, 375, 380, 392, 452, 494, 503, 631, 653, 683, 684, 703, 704, 705, 725, 734, 779, 794, 796, 802, 826
<u>Morone saxatilis</u> (striped bass)	Freshwater, marine	286, 412, 525, 548, 708, 713, 775
<u>Mugil cephalus</u> (mullet)	Marine	311, 315, 330, 455, 532
<u>Notemigonus crysoleucas</u> (golden shiner)	Freshwater	324, 334, 352, 358, 367, 392, 452, 508, 517, 529, 538, 613, 684, 746, 779, 830
<u>Oncorhynchus kisutch</u> (silver or coho salmon)	Freshwater, marine	316, 321, 351, 372, 387, 391, 399, 421, 426, 441, 485, 610, 655, 656, 663, 675, 685, 703, 704, 705, 773
<u>O. nerka</u> (sockeye salmon)	Freshwater, marine	382, 439, 663, 771, 773, 786
<u>O. tshawytscha</u> (chinook salmon)	Freshwater, marine	351, 372, 391, 461, 548, 656, 669, 703, 704, 774
<u>Perca flavescens</u> (yellow perch)	Freshwater	295, 321, 352, 380, 705, 768, 807
<u>Pimephales promelas</u> (fathead minnow)	Freshwater	298, 317, 319, 321, 331, 349, 352, 354, 357, 358, 376, 379, 401, 403, 411, 419, 420, 443, 448, 451, 453, 484, 489, 497, 502, 541, 543, 574, 593, 623, 696, 705, 731, 779, 787, 801, 807
<u>Poecilia reticulata</u> (guppy)	Freshwater	292, 293, 445, 524, 798, 815, 830
<u>Salmo gairdneri</u> (rainbow trout)	Freshwater	280, 295, 304, 306, 318, 321, 326, 337, 342, 349, 351, 352, 353, 363, 372, 373, 377, 380, 387, 389, 392, 395, 398, 402, 404, 411, 421, 422, 423, 424, 425, 427, 428, 431, 432, 467, 476, 477, 478, 479, 483, 496, 504, 528, 530, 646, 655, 656, 658, 669, 684, 705, 733, 753, 755, 771, 778, 818, 829
<u>S. salar</u> (Atlantic salmon)	Freshwater, marine	289, 341, 374, 378, 397, 400, 406, 482, 483, 520, 523, 533, 554, 601, 602, 609, 661, 749, 817
<u>Salvelinus fontinalis</u> (brook trout)	Freshwater, marine	282, 329, 338, 345, 347, 371, 372, 400, 409, 411, 430, 462, 495, 608, 655, 684, 705, 749, 758
<u>Stizostedion v. vitreum</u> (walleye)	Freshwater	280, 301, 310, 332, 349, 352, 353, 359, 411, 415, 530, 622, 705, 753, 757

Table 2  
Infrequently Used Species

Species	Reported Bioassay Habitat	Reference Numbers
BACTERIA/YEASTS/FUNGI		
<u>Arizona arizonae</u> (bacteria)		598
<u>Bacteria</u> (unidentified)		447, 454
<u>Enterobacter aerogenes</u> (bacteria)		598
<u>Escherichia coli</u> (bacteria)		450, 598, 825
<u>Geotrichum candidum</u> (bacteria)		596
<u>Proteus rettgeri</u> (bacteria)		598
<u>Rhodotorula rubra</u> (yeast)		640
<u>Salmonella senftenberg</u> (bacteria)		598
<u>Shigella flexneri</u> (bacteria)		598
<u>Staphylococcus aureus</u> (bacteria)		460
<u>Vibrio marinus</u> (bacteria)		797
PLANTS (Algae)		
<u>Achnanthes brevipes</u> (diatom)	Marine, planktonic	605
<u>Actinocyclus</u> sp. (algae)	Marine, planktonic	650
<u>Amphidinium carteri</u> (diatom)	Marine, planktonic	575, 576, 647
<u>Amphiprora</u> sp. (algae)	Marine, planktonic	650
<u>Amphora exigua</u> (diatom)	Marine, planktonic	605
<u>Amphora</u> sp. (diatom)	Marine, planktonic	650
<u>Anabaena flos-aquae</u> (bluegreen)	Freshwater, planktonic	785
<u>Asterionella japonica</u> (diatom)	Marine, planktonic	626
<u>Blepharisma intermedium</u> (algae)	Freshwater, planktonic	747
<u>Chaetoceros compressus</u> (diatom)	Marine, planktonic	626
<u>C. debilis</u> (diatom)	Marine, planktonic	626
<u>C. decipiens</u> (diatom)	Marine, planktonic	285
<u>C. pelagicus</u> (diatom)	Marine, planktonic	668
<u>Chlamydomonas moewusii</u> (green)	Freshwater, planktonic	738
<u>C. reinhardtii</u> (green)	Freshwater, planktonic	524, 544
<u>Chlorella vanniellii</u> (green)	Freshwater, planktonic	473
<u>C. vulgaris</u> (green)	Freshwater, planktonic	413, 545
<u>Chlorococcum macrostrigatum</u> (green)	Freshwater, planktonic	493
<u>Cladophora</u> sp.	Marine, freshwater, benthic	460, 532
<u>Coccochloris elabans</u> (bluegreen)	Marine, planktonic	556
<u>Cyclotella</u> sp.	Freshwater, planktonic	721
<u>Cylindrotheca closterium</u> (diatom)	Marine, planktonic	626
<u>Dicrateria inornata</u> (green)	Marine, planktonic	637
<u>Dunaliella euchlora</u> (green)	Marine, planktonic	542, 572, 650
<u>Dunaliella</u> sp. (green)	Marine, planktonic	579
<u>Eucampa zodiacus</u> (algae)	Marine, planktonic	626
<u>Euglena viridas</u> (green)	Freshwater, planktonic	413
<u>Exuviella</u> sp. (red)	Marine, planktonic	556, 668
<u>Fontinalis antipyretica</u> (green)	Freshwater, planktonic	790
<u>Fragilaria</u> sp. (diatom)	Freshwater, planktonic	721
<u>Glenodinium foliaceum</u> (red)	Marine, planktonic	556
<u>Glenodinium</u> sp. (red)	Marine, planktonic	556
<u>Guinardia flaccida</u> (algae)	Marine, planktonic	626
<u>Gyrodinium resplendens</u> (algae)	Marine, planktonic	668
<u>Laminaria saccharina</u> (brown)	Marine, benthic	561, 664, 742
<u>Leptocylindricus minimus</u> (diatom)	Marine, planktonic	626
<u>Leptocylindricus</u> sp. (diatom)	Marine, planktonic	573
<u>Licmophra lyngbyei</u> (diatom)	Marine, planktonic	285, 626
<u>Macrocystis pyrifera</u> (brown)	Marine, benthic	599
<u>Mishococcus</u> sp. (diatom)	Marine, planktonic	458
<u>Nannochloris</u> sp. (diatom)	Marine, planktonic	637
<u>Navicula inserta</u> (diatom)	Marine, planktonic	605
<u>Neochloris</u> sp. (green)	Marine, planktonic	605
<u>Nitzschia laevis</u> (diatom)	Marine, planktonic	650
<u>N. linearis</u> (diatom)	Freshwater, planktonic	386
<u>N. seriata</u> (diatom)	Marine, planktonic	626
<u>Nitzschia</u> sp. (diatom)	Marine, planktonic	536
<u>Olisthodiscus luteus</u> (algae)	Marine, planktonic	576
<u>Paralia sulcata</u> (algae)	Marine, planktonic	626
<u>Pediastrum tetras</u> (green)	Freshwater, planktonic	413
<u>Periphyton</u> (unidentified)	Freshwater, benthic	418, 824
<u>Platymonas suecia</u> (green)	Marine, planktonic	647
<u>Platymonas</u> sp. (green)	Marine, planktonic	458, 605, 650
<u>Pleurosigma angulatum</u> (algae)	Marine, planktonic	626
<u>Porphyridium cruentum</u> (red)	Marine, planktonic	458, 605
<u>Protococcus</u> sp. (green)	Marine, planktonic	572, 650
<u>Rhizosolenia fragilissima</u> (diatom)	Marine, planktonic	626
<u>Scenedesmus quadricaudata</u> (green)	Freshwater, planktonic	641
<u>Scenedesmus</u> sp. (green)	Freshwater, planktonic	721

(Continued)

Table 2 (Continued)

Species	Reported Bioassay Habitat	Reference Numbers
PLANTS (Algae) (Continued)		
<i>Selenastrum capricornutum</i> (green)	Freshwater, marine, planktonic	513, 537, 635
<i>S. gracile</i> (green)	Freshwater, planktonic	799
<i>Stauroneis amphoroides</i> (diatom)	Marine, planktonic	605
<i>Synedra</i> sp. (diatom)	Freshwater, planktonic	721
<i>Tetrahymena pyriformis</i> (ciliate)	Freshwater, planktonic	514
<i>T. vorax</i> (ciliate)	Freshwater, planktonic	524
<i>Thalassionema nitzschoides</i> (algae)	Marine, planktonic	626
<i>Thalassiosira fluviatilis</i> (diatom)	Marine, planktonic	556, 605
<i>T. levanderi</i> (diatom)	Marine, planktonic	626
<i>T. nordenskioldii</i> (diatom)	Marine, planktonic	626
<i>Vaucheria</i> sp.	Freshwater, benthic	460
PLANTS (Flowering Plants)		
<i>Alternanthera philoxeroides</i> (alligator-weed)	Freshwater	793
<i>Azolla caroliniana</i> (azolla)	Freshwater	795
<i>Ceratophyllum demersum</i> (coontail)	Freshwater	794
<i>Chara vulgaris</i> (muskgrass)	Freshwater	794
<i>Distichlis spicata</i> (marsh grass)	Estuarine	550
<i>Lemna minor</i> (duckweed)	Estuarine	429, 793, 795
<i>Najas flexilis</i> (pondweed)	Estuarine	794
<i>Potamogeton diversifolius</i> (pondweed)	Estuarine	794
<i>P. foliosus</i> (pondweed)	Estuarine	794
<i>P. pectinatus</i> (pondweed)	Estuarine	794
<i>P. pusillus</i> (pondweed)	Estuarine	794
<i>Rhizophora mangle</i> (mangrove)	Marine	639
<i>Spartina alterniflora</i> (marsh grass)	Estuarine	550, 611
<i>S. patens</i> (marsh grass)	Estuarine	550
<i>Spirodela polyrrhiza</i> (giant duckweed)	Estuarine	795
<i>Wolffia columbiana</i> (watermeal)	Estuarine	795
<i>Wolffiella floridana</i>	Estuarine	795
PROTOZOANS		
<i>Coccolithus huxleyi</i> (coccolith)	Marine	647, 772, 823
COELENTERATES		
<i>Bimera franciscana</i> (hydroid)	Marine	722
<i>Chrysaora quinquecirrha</i> (medusa)	Marine	741
<i>Hydra</i> sp. (hydroid)	Freshwater	302
<i>Phialidium gregarium</i> (medusa)	Marine	662
<i>Sarsia tubulosa</i> (medusa)	Marine	662
CTENOPHORES		
<i>Pleurobrachia pileus</i> (comb jelly)	Marine	662
BRYOZOANS		
<i>Bugula neritina</i> (erect)	Marine	769
<i>Watersipora cucullata</i> (encrusting)	Marine	769
PLATHELMINTHES		
<i>Dugesia</i> sp.	Freshwater	440
<i>Turbellarian</i> (unidentified)	Freshwater	302
ECHINODERMS		
<i>Asterias forbesi</i> (starfish)	Marine	286, 381
<i>Crossaster papposus</i> (starfish)	Marine	562
<i>Dendraster exocutricus</i> (starfish)	Marine	562
<i>Luidia foliata</i> (starfish)	Marine	562
<i>Pisaster ochraceus</i> (starfish)	Marine	562
<i>Strongylocentrotus purpuratus</i> (urchin)	Marine	767
MOLLUSKS		
<i>Aequipecten irradians</i> (scallop)	Marine	680
<i>Amnicola</i> sp.	Freshwater	816
<i>Arctica islandica</i> (clam)	Marine	512
<i>Cameloma decisum</i> (snail)	Freshwater	348, 432
<i>Clinocardium nuttalli</i> (cockle)	Marine	301, 344
<i>Gemma gemma</i> (clam)	Marine	666
<i>Coniobasis</i> sp. (snail)	Freshwater	364
<i>Haminoea virescens</i> (limpet)	Marine	562
<i>Katharina tunicata</i> (chiton)	Marine	562
<i>Lymnaea</i> sp. (snail)	Freshwater	440
<i>Macoma balthica</i> (clam)	Marine, estuarine	390, 666, 717
<i>M. phenax</i> (clam)	Marine, estuarine	390
<i>Melibe lemina</i> (limpet)	Marine	562
<i>Modiolus demissus</i> (mussel)	Estuarine	547

(Continued)

Table 2 (Continued)

Species	Reported Bioassay Habitat	Reference Numbers
MOLLUSKS (Continued)		
<u>Mulinia lateralis</u> (clam)	Marine	666, 717
<u>Mytilus californianus</u> (mussel)	Marine	667, 668
<u>Nassa obsoleta</u> (snail)	Marine	333
<u>Nassarius obsoletus</u> (snail)	Marine	286, 381
<u>Physa integra</u> (snail)	Freshwater	348, 432
<u>Physa</u> sp. (snail)	Freshwater	302
<u>Rangia cuneata</u> (clam)	Marine, estuarine	390, 813
<u>Spisula solidissima</u> (clam)	Marine	512
<u>Unionidae</u> (clam)	Freshwater	302
<u>Urosalpinx cinerea</u> (oyster drill)	Estuarine	381
ANNELIDS		
<u>Capitella capitata</u> (polychaete)	Marine	417, 553, 580
<u>Clymenella torquata</u> (polychaete)	Marine	671
<u>Diopatra cuprea</u> (polychaete)	Marine	671
<u>Dorvillea articulata</u> (polychaete)	Marine	417
Leeches	Freshwater	302
<u>Limnodrilus udekemianus</u> (tubificid)	Freshwater	766
<u>Nais</u> sp. (oligochaete)	Freshwater	816
<u>Neanthes arenaceodentata</u> (polychaete)	Marine	417
<u>Nereis branti</u> (sandworm)	Marine, estuarine	562
<u>N. grubbei</u> (sandworm)	Marine, estuarine	417
<u>N. diversicolor</u> (sandworm)	Marine	283, 780, 783, 784
<u>N. vexillosa</u> (sandworm)	Marine	562
<u>N. virens</u> (sandworm)	Marine	286, 592
<u>Polydora</u> sp. (polychaete)	Marine	594
<u>Serpula vermicularis</u> (tubeworm)	Marine	562
<u>Tubifex tubifex</u> (tubificid)	Freshwater	551, 566
ARTHROPODS (Crustaceans)		
<u>Acartia clausii</u> (copepod)	Estuarine, marine	651
<u>A. tonsa</u> (copepod)	Estuarine, marine	390, 534, 594
<u>Asellus breviceaudus</u> (isopod)	Freshwater	567
<u>A. militaris</u> (isopod)	Freshwater	416, 690
<u>Asellus</u> sp. (isopod)	Freshwater	302
<u>Balanus cariosus</u> (barnacle)	Marine	562
<u>B. eburneus</u> (barnacle)	Estuarine, marine	472
<u>B. improvisus</u> (barnacle)	Estuarine	534
<u>Calanus finmarchicus</u> (copepod)	Marine	633
<u>Callinasa californianus</u> (ghost shrimp)	Marine	301
<u>Cancer irroratus</u> (crab)	Marine	512
<u>C. magister</u> (crab)	Marine	301, 340, 660
<u>Carcinus maenas</u> (crab)	Marine	381, 536, 560, 563, 590, 674, 732
<u>Chlamydotheca arcuata</u> (ostracod)	Freshwater	287, 654
<u>Cyclops varicans</u> (copepod)	Freshwater	822
<u>Cyprretta kawatai</u> (ostracod)	Freshwater	619
<u>Cypridopsis vidua</u> (ostracod)	Freshwater	524, 567
<u>Daphnia melanogaster</u> (cladoceran)	Freshwater	745
<u>Eubranchipus moorei</u> (crustacean)	Freshwater	690
<u>Euchaeta japonica</u> (copepod)	Marine	410, 565
<u>Eulimnadia inflecta</u> (conchostracan)	Freshwater	690
<u>Euphausia pacifica</u> (euphausiid)	Marine	699
<u>Eurytemora affinis</u> (copepod)	Estuarine, Marine	390
<u>Gammarus</u> sp. (crayfish)	Freshwater	302
<u>Gammarus fasciatus</u> (amphipod)	Freshwater	416, 567
<u>G. figrinus</u> (amphipod)	Estuarine	534, 581
<u>G. gammarus</u> (amphipod)	Freshwater	463
<u>G. lacustris</u> (amphipod)	Freshwater	323
<u>G. oceanicus</u> (amphipod)	Marine	289, 341, 454
<u>Gammarus</u> sp. (amphipod)	Freshwater	302, 816
<u>Hemigrapsus oregonensis</u> (crab)	Marine	301
<u>Hepatus epheliticus</u> (crab)	Estuarine, marine	678
<u>Hyalella azeteca</u> (amphipod)	Freshwater	383, 814
<u>Leptodius floridanus</u> (crab)	Marine	568, 569
<u>Limnoria lignorum</u> (gribble)	Marine, estuarine	617
<u>L. quadripunctata</u> (gribble)	Marine, estuarine	617
<u>L. tripunctata</u> (gribble)	Marine, estuarine	617
<u>Melita nitida</u> (amphipod)	Estuarine	534
<u>Orconectes nais</u> (crayfish)	Freshwater	394, 463, 567
<u>O. rusticus</u> (crayfish)	Freshwater	364
<u>Pagurus longicarpus</u> (hermit crab)	Marine	286, 381
<u>P. pollicaris</u> (hermit crab)	Marine	512
<u>Paleomonetes</u> sp. (grass shrimp)	Estuarine	458

(Continued)

Table 2 (Continued)

Species	Reported Bioassay Habitat	Reference Numbers
ARTHROPODS (Crustaceans) (Continued)		
<i>Panopeus herbstii</i> (crab)	Marine	568, 678
<i>Pontoporeia affinis</i> (amphipod)	Freshwater	279, 405
<i>Procambarus acutus</i> (crayfish)	Freshwater	526
<i>P. blandingi</i> (crayfish)	Freshwater	642
<i>P. clarki</i>	Freshwater	360, 709, 712
<i>P. simulans</i> (crayfish)	Freshwater	394
<i>Pseudocalanus</i> sp. (copepod)	Marine	633
<i>Pseudodiaptomus cornutus</i> (copepod)	Marine	285
<i>Pseudodiaptomus</i> sp. (copepod)	Marine	594
<i>Rhithropanopeus harrisi</i> (crab)	Estuarine, marine	678, 688
<i>Scottolana canadensis</i> (copepod)	Estuarine	390
<i>Sesarma cinereum</i> (crab)	Estuarine, marine	678, 764
<i>Simoecephalus serratulatus</i> (cladoceran)	Freshwater	297
<i>Streptocephalus seali</i> (anostracan)	Freshwater	690
<i>Temora longicornis</i> (copepod)	Marine	633
<i>Tigriopus californianus</i> (copepod)	Marine	519
<i>Tortanus discaudatus</i> (copepod)	Marine	827
<i>Uca minax</i> (fiddler crab)	Estuarine	311
<i>Uca pugilator</i> (fiddler crab)	Marine, estuarine	612, 614, 764
<i>Uca pugnax</i> (fiddler crab)	Estuarine	764
<i>Upogebia pugettensis</i> (mud shrimp)	Marine	301
ARTHROPODS (Insects)		
<i>Agrion</i> sp. (damselfly)	Freshwater	740
<i>Anax</i> sp. (dragonfly)	Freshwater	740
<i>Anopheles quadrimaculatus</i> (mosquito)	Freshwater	498, 508
<i>Arctopsyche grandis</i> (caddisfly)	Freshwater	327, 488
<i>Atherix</i> sp. (snipefly)	Freshwater	302
<i>Baetisca laurentina</i> (mayfly)	Freshwater	706
<i>Boyeria vinuosa</i> (dragonfly)	Freshwater	707
<i>Boyeria</i> sp. (dragonfly)	Freshwater	302
<i>Brachycentrus americanus</i> (caddisfly)	Freshwater	707
<i>Callibaetis</i> sp. (mayfly)	Freshwater	383
<i>Cheumatopsyche</i> sp. (caddisfly)	Freshwater	302, 364
<i>Chironomus plumosus</i> (midge)	Freshwater	765
<i>C. staegeri</i> (midge)	Freshwater	765
<i>C. tentans</i> (midge)	Freshwater	805
<i>Chironomus</i> sp. (midge)	Freshwater	463, 816
<i>Chloroperia</i> sp. (stonefly)	Freshwater	302
<i>Claasenia sabulosa</i> (stonefly)	Freshwater	327, 488, 800
<i>Corixidae</i> (water boatman)	Freshwater	302
<i>Corydalus cornutus</i> (hellgrammite)	Freshwater	613
<i>Culex pipiens</i> (mosquito)	Freshwater	463
<i>Culex</i> sp. (mosquito)	Freshwater	440
<i>Dineutes americanus</i> (beetle)	Freshwater	814
<i>Enallagma</i> sp. (damselfly)	Freshwater	383
<i>Ephemera simulans</i> (mayfly)	Freshwater	486, 706, 707
<i>Ephemerella</i> sp. (mayfly)	Freshwater	302
<i>Gomphus</i> sp. (dragonfly)	Freshwater	302
<i>Hexagenia bilineata</i> (mayfly)	Freshwater	463
<i>H. limbata</i> (mayfly)	Freshwater	706
<i>Hexagenia</i> sp. (mayfly)	Freshwater	302, 308
<i>Hydroperla crosbyi</i> (stonefly)	Freshwater	613
<i>Hydropsyche californica</i> (caddisfly)	Freshwater	327
<i>Ischnura verticalis</i> (mayfly)	Freshwater	463
<i>Isogenus tontinalis</i> (stonefly)	Freshwater	707
<i>Isonychia bicolor</i> (mayfly)	Freshwater	364
<i>Isonychia</i> sp. (mayfly)	Freshwater	302
<i>Isoperla</i> sp. (stonefly)	Freshwater	302
<i>Leptophlebia nebulosa</i> (mayfly)	Freshwater	706
<i>Lestes curinus</i> (damselfly)	Freshwater	691
<i>Lestes</i> sp. (damselfly)	Freshwater	690
<i>Lethocercus</i> sp. (giant waterbug)	Freshwater	302
<i>Libellula</i> sp. (dragonfly)	Freshwater	383, 463
<i>Limnephilus</i> sp. (caddisfly)	Freshwater	383
<i>Macromia</i> sp. (dragonfly)	Freshwater	814
<i>Ophiogomphus rupinsulensis</i> (dragonfly)	Freshwater	707
<i>Ophiogompha</i> sp. (dragonfly)	Freshwater	302
<i>Paragnetina</i> sp. (stonefly)	Freshwater	302
<i>Phryganea</i> sp. (caddisfly)	Freshwater	740
<i>Procladius</i> sp. (insect)	Freshwater	616
<i>Pteronarcella badia</i> (stonefly)	Freshwater	800
<i>Pteronarcys dorsata</i> (stonefly)	Freshwater	706, 707

(Continued)

(Sheet 4 of 6)

Table 2 (Continued)

Species	Reported Bioassay Habitat	Reference Numbers
ARTHROPODS (Insects) (Continued)		
<u>Simuliidae</u> (blackfly)	Freshwater	302
<u>Simulium tuberosum</u> (blackfly)	Freshwater	754
<u>S. venustum</u> (blackfly)	Freshwater	754
<u>S. verecundum</u> (blackfly)	Freshwater	754
<u>Siphonurus</u> sp. (mayfly)	Freshwater	463, 740
<u>Stenonema ares</u> (mayfly)	Freshwater	364
<u>S. heterotarsale</u> (mayfly)	Freshwater	364
<u>Stenonema rubrum</u> (mayfly)	Freshwater	707
<u>Stenonema</u> sp. (mayfly)	Freshwater	302
<u>Stictochironomus annuliferus</u> (bloodworm)	Freshwater	616
<u>Taeniopteryx maura</u> (stonefly)	Freshwater	707
<u>Tanytarsus dissimilis</u> (midge)	Freshwater	621, 706
<u>Tanytarsus</u> sp. (midge)	Freshwater	765
<u>Tendipedidae</u> sp. (bloodworm)	Freshwater	302, 383
<u>Tetragoneuria cynosura</u> (dragonfly)	Freshwater	625
<u>T. semiagua</u> (dragonfly)	Freshwater	625
<u>Trienoides</u> sp. (caddisfly)	Freshwater	302
<u>Tropiternus lateralis</u> (beetle)	Freshwater	756
FISHES		
<u>Adinia xenica</u> (killifish)	Estuarine	311
<u>Agosia chrysogaster</u> (longfin dace)	Freshwater	763
<u>Alosa aestivalis</u> (blueback herring)	Freshwater	810
<u>A. pseudoharengus</u> (alewife)	Marine, freshwater	648
<u>A. sapidissima</u> (American shad)	Freshwater	469, 665, 694, 808
<u>Ameiurus melas</u> (black bullhead)	Freshwater	692
<u>A. nebulosus</u> (brown bullhead)	Freshwater	466, 503, 684
<u>Anguilla rostrata</u> (American eel)	Freshwater, marine	315, 455, 525, 645
<u>Apeltes quadracus</u> (stickleback)	Marine	819
<u>Camptostoma anomalum</u> (stoneroller)	Freshwater	299
<u>Carassius carassius</u> (goldfish)	Freshwater	440
<u>Catostomus clarki</u> (gila sucker)	Freshwater	689
<u>Chasmodes bosquianus</u> (striped blenny)	Estuarine	719
<u>Clupea harengus</u> (herring)	Marine	341, 587, 588, 627
<u>Cottus bairdii</u> (sculpin)	Freshwater	295
<u>C. perplexus</u> (sculpin)	Freshwater	685
<u>Ctenopharyngodon idellus</u> (grass carp)	Freshwater	705
<u>Cymatogaster aggregata</u> (shiner perch)	Estuarine, marine	290, 301, 607
<u>Cyprinodon macularius</u> (desert pupfish)	Freshwater	679, 689
<u>Engraulis mordax</u> (northern anchovy)	Marine	505
<u>Erimyzon sucetta</u> (lake chubsucker)	Freshwater	299
<u>Etheostoma lepidum</u> (greenthroat darter)	Freshwater	739
<u>Eucalia inconstans</u> (stickleback)	Freshwater	743
<u>Fundulus diaphanus</u> (banded killifish)	Freshwater	525, 645, 705
<u>F. grandis</u> (killifish)	Estuarine	311
<u>F. majalis</u> (killifish)	Estuarine	315, 333, 381
<u>F. similis</u> (killifish)	Estuarine	311, 458
<u>Fundulus</u> sp. (killifish)	Estuarine	311
<u>Gadus morhua</u> (Atlantic cod)	Marine	397, 515, 579, 587, 590
<u>Girella nigricans</u> (greenfish)	Marine	770
<u>Gobiesox strumosus</u> (skilletfish)	Estuarine	719
<u>Gobionellus boleosoma</u> (goby)	Estuarine	311
<u>Gobiosoma bosci</u> (naked goby)	Estuarine	719
<u>Harengula pensacolatae</u> (sardine)	Marine	809
<u>Ictalurus nebulosus</u> (brown bullhead)	Freshwater	295, 796, 818, 828
<u>Ictiobus cyprinellus</u> (bigmouth buffalo)	Freshwater	324
<u>Jordanella floridae</u> (flagfish)	Freshwater	456, 610
<u>Kuhlia sandvichensis</u> (aholehole)	Marine	798
<u>Lepomis humilis</u> (orangespotted sunfish)	Freshwater	692
<u>L. microlophus</u> (redear sunfish)	Freshwater	321, 681, 794, 796
<u>Lucania parva</u> (killifish)	Estuarine	311
<u>Megastomatobus cyprinella</u> (buffalo)	Freshwater	779
<u>Menidia beryllina</u> (silverside)	Estuarine	311
<u>M. menidia</u>	Estuarine, marine	315, 330, 455, 694
<u>Micrometrus minimus</u> (dwarf shiner)	Marine, estuarine	339, 607
<u>Micropogon undulatus</u> (croaker)	Marine, estuarine	507, 570, 657
<u>Moxostoma aureoleum</u> (redhose)	Freshwater	692
<u>Notropis a. atherinoides</u> (minnow)	Freshwater	702
<u>N. blennius</u> (minnow)	Freshwater	692
<u>N. cornutus</u> (minnow)	Freshwater	295
<u>N. heterolepis</u> (bluntnose shiner)	Freshwater	352
<u>N. rubellus</u> (minnow)	Freshwater	702

(Continued)

(Sheet 5 of 6)

Table 2 (Concluded)

Species	Reported Bioassay Habitat	Reference Numbers
FISHES (Continued)		
<u>N. spilopterus</u> (minnow)	Freshwater	702, 725
<u>Oligocottus snyderi</u> (fluffy sculpin)	Marine	393
<u>Oncorhynchus gorbuscha</u> (pink salmon)	Freshwater, marine	546, 663
<u>Opsanus tau</u> (toadfish)	Marine	657, 682
<u>Osmorus mordax</u> (smelt)	Marine	347
<u>Parophrys vetulus</u> (English sole)	Marine	301
<u>Percina caprodes</u> (logperch)	Freshwater	295
<u>Petromyzon marinus</u> (sea lamprey)	Freshwater	295, 352
<u>Pimephales notatus</u> (bluntnose minnow)	Freshwater	510, 692, 702, 796
<u>Poecilia latipinna</u> (silfin molly)	Freshwater, estuarine	305, 311, 322, 440
<u>Pomatomus saltatrix</u> (bluefish)	Marine	315
<u>Pomoxis annularis</u> (white crappie)	Freshwater	392, 503
<u>P. nigromaculatus</u> (black crappie)	Freshwater	683, 684, 779
<u>Pseudopleuronectes americanus</u> (flounder)	Marine	315, 455, 694
<u>Ptychocheilus oregonensis</u> (squawfish)	Freshwater	372
<u>P. umpquae</u> (squawfish)	Freshwater	372
<u>Rhinichthys a. atratulus</u> (bluntnose dace)	Freshwater	701
<u>R. scutellus</u> (speckled dace)	Freshwater	689
<u>Salmo clarki</u> (cutthroat trout)	Freshwater	511, 600, 655
<u>S. trutta</u> (brown trout)	Freshwater	306, 321, 684, 705
<u>Salvelinus namaycush</u> (lake trout)	Freshwater	705
<u>Sebastodes miniatus</u> (Pacific rockfish)	Marine	705, 720
<u>Semotilus a. atromaculatus</u> (creek chub)	Freshwater	701
<u>Sphaeroides maculatus</u> (northern puffer)	Marine	315, 694
<u>Stenotomus chrysops</u> (scup)	Marine	286
<u>Stolephorus purpurea</u> (nehu)	Marine	798
<u>Syngnathus fuscus</u> (northern pipefish)	Marine	315
<u>Tautoglabrus adspersus</u> (cunner)	Marine, estuarine	512, 819
<u>Tilapia mossambica</u> (tilapia)	Freshwater	798
<u>Tilapia</u> sp. (tilapia)	Freshwater	334
<u>Trinectes maculatus</u> (hogchoker)	Marine, estuarine	390, 474, 687
<u>Xiphophorus maculatus</u> (platyfish)	Freshwater	403
AMPHIBIANS		
<u>Rana catesbeiana</u> (bullfrog)	Freshwater	324, 334
<u>Rana pipiens</u> (leopard frog)	Freshwater	334
<u>Rana</u> sp. (frog)	Freshwater	440
BIRDS		
<u>Anas platyrhynchos</u> (mallard duck)	Freshwater	284



Table 3

## Test Species Previously Used in Turbidity and Turbidity-Related Bioassays

Species	Life Cycle Stage	Type of Turbidity	Reference Numbers
<u>Acartia tonsa</u>	Adults	Sludge	594
Bacteria			
<u>Escherichia coli</u>	Not applicable	Stream sediment eluates	598
<u>Enterobacter aerogenes</u>	Not applicable	Stream sediment eluates	598
<u>Proteus rettgeri</u>	Not applicable	Stream sediment eluates	598
<u>Arizona arizonae</u>	Not applicable	Stream sediment eluates	598
<u>Shigella flexneri</u>	Not applicable	Stream sediment eluates	598
<u>Salmonella senftenberg</u>	Not applicable	Stream sediment eluates	598
<u>Balanus balanoides</u>	Adults	Oil and carbonized sand	682
<u>Callinectes sapidus</u>	Adults	Sediment-sorbed radioactive gold	657
<u>Catostomus commersoni</u>	Adult	Paper mill effluent	295
<u>Chaetoceros decipiens</u>	Adults	Acid-iron wastes	285
<u>C. pelagicus</u>	Adults	Chromium and lead in marine sewage outfall sediments	668
<u>Cottus bairdii</u>	Adult	Paper mill effluent	295
<u>Crassostrea gigas</u>	Larvae/adults	Pulp mill wastes	677
	Adults	Suspended silt, kaolin, fuller's earth	811
<u>C. virginica</u>	Adults	Turbidity, siltation from dredging operations	294
	Larvae	Turbidity-producing substances	604
	Adults/larvae	Silt and kaolin turbidity	636
		Suspended silt	649
	Adults	Oil and carbonized sand	682
<u>Cyprinodon variegatus</u>	Adult	Various types of silt	819
<u>Dunaliella tertiolecta</u>	Adults	Chromium and lead in marine sewage outfall sediments	668
<u>Euchaeta japonica</u>	Eggs/larvae	Copper and marine sediment extracts	410
	Adults	Affect of particulate clay minerals, diatoms, ascorbic acid, humic acid, sewage effluent and soil extracts	565
<u>Exuviella</u> sp.	Adults	Chromium and lead in marine sewage outfall sediments	668
<u>Fundulus heteroclitus</u>	Adult	Various types of silt	819
<u>Gammarus pseudolimnaeus</u>		Paper fiber sludge	359
<u>Gasterosteus aculeatus</u>	Adults	Harbor sediment contaminants	399
<u>Gyrodinium resplendens</u>	Adults	Chromium and lead in marine sewage outfall sediments	668
<u>Homarus americanus</u>	Larvae/adults	Kraft pulp mill effluent	482, 676
<u>Hyalella azteca</u>	Adult	Herbicides and pond mud	383
<u>Ictalurus nebulosus</u>	Adults	Paper mill effluent	295
<u>Laminaria saccharina</u>	Adults	Silt/various pollutants	742
<u>Lebistes reticulatus</u>	Adults	Paper factory effluents	531
	Adults	Kraft pulp mill effluents	786
<u>Leiostomus xanthurus</u>	Adults	Sediment-sorbed radioactive gold	657
<u>Lepomis cyanellus</u>	Juveniles	Silt, turbidity	620
<u>L. gibbosus</u>	Adults	Paper mill effluent	295
<u>Licmorpha lyngbyei</u>	Adults	Acid-iron wastes	285
<u>Limnodrilus hoffmeisteri</u>	Adults	Detritus-sorbed radioisotopes, transfer to macroinvertebrates	616
<u>Limnephilus</u> sp.	Larvae	Herbicides and pond mud	383
<u>Mercenaria mercenaria</u>	Eggs/larvae	Suspended silt	516
	Larvae	Turbidity-producing substances	604
	Adults/larvae	Silt and kaolin turbidity	636
	Adults	Oil and carbonized sand	682
Microflora	Not applicable	Oil dispersant in beach sand	659
<u>Micropterus salmoides</u>	Adults	Paper mill effluent	295
	Juveniles	Suspended silt	620
<u>Mytilus edulis</u>	Embryos	Kraft pulp mill effluent	414
	Adults	Ferric hydroxide turbidity	586, 736
<u>Nereis diversicolor</u>	Adults	Manganese in estuarine sediments	783
	Adults	Zinc and cadmium in estuarine sediments	784
<u>Nitzschia closterium</u>	Adults	Acid-iron wastes	285
<u>Notropis a. atherinoides</u>	Adults	Kraft pulp mill effluent	702
<u>N. cornutus</u>	Adults	Paper mill effluent	295
<u>N. notatus</u>	Adults	Kraft pulp mill effluent	702
<u>N. rebellus</u>	Adults	Kraft pulp mill effluent	702
<u>N. spilopterus</u>	Adults	Kraft pulp mill effluent	702
<u>Oligocottus snyderi</u>	Adults	Kraft pulp mill effluent	393
<u>Oncorhynchus gorbuscha</u>		Kraft pulp mill effluent	663
<u>O. kisutch</u>	Adults/juveniles	Kraft pulp mill effluent	316, 382, 421, 426, 663
<u>O. nerka</u>		Kraft pulp mill effluent	439, 663, 786

(Continued)

Table 3 (Concluded)

Species	Life Cycle Stage	Type of Turbidity	Reference Numbers
<u>Opsanus tau</u>	Adults	Sediment-sorbed radioactive gold	657
	Embryos	Oil and carbonized sand	682
<u>Ostrea edulis</u>	Larvae	Turbidity-producing substances	604
<u>Paleomonetes pugio</u>	Adults	Sludge	594
<u>Perca flavescens</u>	Adults	Paper mill effluent	295
<u>Percina caprodes</u>	Adults	Paper mill effluent	295
<u>Petromyzon marinus</u>	Adults	Paper mill effluent	295
<u>Physa heterostropha</u>	Adults	Detritus-sorbed radioisotopes, transfer to macroinvertebrates	616
<u>Pimephales promelas</u>	Adults	Wood fibers	317
	Adults	Ferric hydroxide turbidity	401
	Adults	Turbidity and endrin	696
<u>Poecilia reticulata</u>	Adults	Mercury-containing sediments	292, 293
<u>Polydora</u> sp.	Larvae	Sludge	594
<u>Pontoporeia affinis</u>	Adults	Harbor and Great Lakes sediments	279
	Adults	Dredged sediments	405
<u>Pseudodiaptomus coronatus</u>	Adults	Acid-iron wastes	285
<u>Pseudodiaptomus</u> sp.	Adults	Sludge	594
<u>Salmo gairdneri</u>		Paper mill effluent	295
	Eggs/fry	Wood fibers	306
	Juveniles	Kraft pulp mill effluent	421
<u>S. salar</u>		Kraft pulp mill effluent	482, 483
<u>S. trutta</u>	Eggs/fry	Wood fibers	306
<u>Salvelinus fontinalis</u>	Juveniles	Turbidity and neutralized Fe(OH) <sub>2</sub>	282
<u>Skeletonema costatum</u>	Adults	Acid-iron wastes	285
<u>Stictochironomus annulicrus</u>	Adults	Detritus-sorbed radioisotopes, transfer to macroinvertebrates	616
<u>Stizostedion v. vitreum</u>		Wood fibers	307, 310
	Eggs/fry	Wood fibers	332
	Eggs/fry	Paper fiber sludge	359
	Eggs	Wood fibers/slime bacterium	415
<u>Tautoglabrus adspersus</u>	Adults	Various types of silt	819
<u>Tubularia crocea</u>	Adults	Oil and carbonized sand	682
Viruses	Not applicable	Clay particles	446

Table 4

Test Species Which Have Been Used Successfully in Bioassays but Present Some Special Problems

(A brief explanation of problems encountered and reported follows each accession number)

Species	Reference Numbers	Problems
<b>PLANTS</b>		
<u>Laminaria saccharina</u> (brown algae)	664	Can grow too large for laboratory conditions
<b>MOLLUSKS</b>		
<u>Crassostrea gigas</u> (oyster)	677	Some mortality problems among larvae and spat
<u>Clinocardium nuttalli</u> (cockle)	344	Unexplained mortalities
<u>Macoma balthica</u> (clam)	390	Poor experimental animal
<u>M. phenax</u> (clam)	390	Poor experimental animal
<u>Mercenaria mercenaria</u> (quahog)	471	Resistant to methoxychlor and malathion
<u>Mytilus edulis</u> (mussel)	536	Problems with laboratory mortality
	552	100 percent control mortality in summer months, feeding problems
	578	Difficult to determine point of death
	586	10 percent control mortality at 4 months, 20 percent at 5 months
	672	High control and experimental mortalities
<b>ARTHROPODS</b>		
<u>Acartia tonsa</u> (copepod)	534	Sensitive to mechanical handling
<u>Artemia salina</u> (brine shrimp)	492	Poor survival rate during development to adult stage due to crowding
<u>Balanus improvisus</u> (barnacle)	534	Larvae sensitive to mechanical handling
<u>Callinectes sapidus</u> (blue crab)	468	High mortality problem
	470	High mortality rate required confinement to cages
	723	Cannibalism problems during ecdysis
<u>Carcinus maenas</u> (crab)	536	Difficult to keep alive due to reluctance to feed in laboratory
<u>Chlamydotheca arcuata</u> (ostracod)	287	Death point difficult to determine
<u>Gammarus oceanicus</u> (amphipod)	341	Difficult to determine point of death
	506	Some cannibalism problems
<u>Homarus americanus</u> (lobster)	482	Artificial seawater not favorable for long survival of larvae
<u>Hydropsyche</u> sp. (caddisfly)	308	Larvae survival beyond 96 hours poor
<u>Lethocercus</u> sp. (giant waterbug)	302	High control mortality
<u>Penaeus duorarum</u> (pink shrimp)	557	4-26 percent control mortality in 15-53 days
<u>Procambarus acutus</u> (crayfish)	526	Cannibalism problems with more than one individual per container
<u>P. clarki</u> (crayfish)	360	Cannibalism problems with more than one individual per container
	709	High mortality during transporting of young
<u>Scototolana canadensis</u> (copepod)	390	Contamination of culture medium caused high mortality
<u>Tropisternus lateralis</u> (beetle)	756	Larvae require separation due to cannibalism problems
<b>FISH</b>		
<u>Alosa aestivalis</u> (blueback herring)	810	Some mortality due to handling stress
<u>A. pseudoharengus</u> (alewife)	648	High mortality rate upon introduction of seawater, fragile
<u>A. sapidissima</u> (American shad)	665	Mortality problems during handling
	694	Death rate accelerated due to overexcitation when placed in clear plexiglass containers
<u>Carassius auratus</u> (goldfish)	335	Problems with <u>Gyrodactylus</u> skin flukes, requiring repeated methylene blue treatments
<u>Cyprinodon macularius</u> (desert pupfish)	689	Pugnacious behavior requires separation from other species
<u>Cyprinus carpio</u> (carp)	366	Field experiments did not work where histological examinations were required
<u>Gambusia affinis</u> (mosquitofish)	339	Treatments with 2.5 percent noniodized table salt required to control tail rot fungus
<u>Gasterosteus aculeatus</u> (threespine stickleback)	351	High mortality among larvae 4 to 7 days after hatching
<u>Leiostomus xanthurus</u> (spot)	387	High control mortality due to crowding and low seawater flow
<u>Lepomis macrochirus</u> (bluegill)	312	High control mortality
	369	Parasitism problems
	408	Dominance-submissive relations develop among fish held together in small containers and affect results
<u>Megalops atlanticus</u> (tarpon)	305	Difficult to keep alive in laboratory, small number available
<u>Micropterus salmoides</u> (largemouth bass)	369	Parasitism problems
<u>Mugil cephalus</u> (mullet)	352	Fish tend to jump out of holding tanks
<u>Notropis cornutus</u> (shiner)	295	Vulnerable to ordinary retention
<u>Osmerus mordax</u> (smelt)	347	Salinity of 33 o/oo caused some control mortality
<u>Pomoxis annularis</u> (white crappie)	620	Suitable for field experiments, but not suitable for laboratory tests due to inability to adapt to aquarium conditions
<u>Ptychocheilus oregonensis</u> (squawfish)	372	Much more sensitive to piscicide than salmonids
<u>P. umquae</u> (squawfish)	372	Much more sensitive to piscicide than salmonids
<u>Salmo clarki</u> (cutthroat trout)	511	Approximately 20 percent control mortality after 20 months
<u>S. gairdneri</u> (rainbow trout)	778	Problem with fish jumping out of holding tanks
<u>Salmo salar</u> (Atlantic salmon)	400	When trout are handled, air is lost from the swim bladder and must be replaced at the surface before they are able to swim without undue effort
<u>Salvelinus fontinalis</u> (brook trout)	282	High mortality due to outbreak of <u>Ichthyophthirius</u> fungus
	338	Fish established territories and attacked each other when more than one fish was placed in small (317 liter) container
	347	In marine tests salinity of 33 o/oo caused some control mortality
	400	When trout are handled, air is lost from the swim bladder and must be replaced at the surface before they are able to swim without undue effort
<u>Sphaeroides maculatus</u> (northern puffer)	315	Died within 24 hours in 20 liter jars without aeration

Table 5  
Rankings of Fish Species on a Scale of 1-8 as  
Suitable Bioassay Animals<sup>245</sup>

Species in Order of Suitability	Availability	Size	Health	Adaptability to Laboratory Life	Metabolic Rate	Acceptance of Food	Does not Jump from Effluent	Total
1. <u>Poecilia reticulata</u>	8	8	8	8	8	8	8	56
2. <u>Gambusia affinis</u>	7	8	8	8	8	8	8	55
3. <u>Pimephales notatus</u>	4	8	8	8	8	8	8	52
4. <u>Pimephales promelas</u>	4	8	8	8	8	8	8	52
5. <u>Pimephales vigilax</u>	4	8	8	8	8	8	8	52
6. <u>Lepomis macrochirus</u>	5	7	8	7	8	8	8	50
7. <u>Carassius auratus</u>	4	7	8	7	8	8	8	51
8. <u>Chromosmus erythrogaster</u>	4	8	8	6	8	8	8	50
9. <u>Lepomis cyanellus</u>	4	7	8	7	8	8	8	50
10. <u>Notropis boops</u>	4	8	8	6	8	8	8	50
11. <u>Notropis lutrensis</u>	4	8	8	7	8	8	7	50
12. <u>Semotilus atromaculatus</u>	4	8	8	6	8	8	8	50
13. <u>Tilapia nilotica</u>	4	6	8	8	8	8	8	50
14. <u>Camptostoma anomalum</u>	4	8	6	7	8	8	8	49
15. <u>Lepomis gulosus</u>	2	6	8	7	8	8	8	49
16. <u>Etheostoma spectabile</u>	4	8	8	7	8	6	8	49
17. <u>Fundulus olivaceus</u>	3	8	8	6	8	8	8	49
18. <u>Labidesthes sicculus</u>	4	8	8	7	8	6	8	49
19. <u>Notropis zonatus</u>	2	8	8	6	8	7	8	49
20. <u>Lepomis humilis</u>	2	7	8	7	8	8	8	48
21. <u>Hybognathus placitus</u>	4	7	7	6	8	7	8	47
22. <u>Lepomis megalotus</u>	3	6	8	6	8	8	8	47
23. <u>Notropis cornutus</u>	2	7	8	6	8	8	8	47
24. <u>Notropis (percobromis)</u> <u>atherinoides</u>	4	8	6	6	7	8	8	47
25. <u>Notropis stramineus</u>	2	8	7	6	8	8	8	47
26. <u>Cottus caroliniae</u>	2	8	6	6	8	8	8	46
27. <u>Cyprinus carpio</u>	4	4	8	6	8	8	8	46
28. <u>Culaea inconstans</u>	4	8	8	8	2	8	8	46
29. <u>Hybognathus hankinsoni</u>	2	8	6	6	8	8	8	46
30. <u>Hybopsis amblops</u>	2	8	8	6	8	6	8	46
31. <u>Ictalurus melas</u>	4	4	8	6	8	8	8	46
32. <u>Notemigonus crysoleucas</u>	4	8	8	6	8	8	6	46
33. <u>Notropis dorsalis</u>	2	8	8	4	8	8	8	46
34. <u>Notropis girardi</u>	3	8	7	6	8	6	8	46
35. <u>Notropis whipplei</u>	2	8	6	6	8	8	8	46
36. <u>Ambloplites rupestris</u>	2	6	8	8	8	5	8	45
37. <u>Lepomis microlophus</u>	4	7	8	6	8	4	8	45
38. <u>Notropis venustus</u>	2	8	6	6	8	8	7	45
39. <u>Micropterus salmoides</u>	2	4	8	6	8	8	8	44
40. <u>Menidia audens</u>	2	8	6	6	8	6	8	44
41. <u>Notropis camurus</u>	2	8	8	4	8	6	8	44
42. <u>Lepomis auritus</u>	3	6	7	6	8	6	8	43
43. <u>Ictalurus nebulosus</u>	2	4	8	4	8	8	8	42
44. <u>Noturus exilis</u>	2	8	4	6	8	6	8	42
45. <u>Ictalurus punctatus</u>	2	4	7	4	8	8	8	41
46. <u>Ictiobus cyprinellus</u>	4	4	6	1	8	8	8	41
47. <u>Notropis rubellus</u>	1	8	4	4	8	6	8	40
48. <u>Dionda nubila</u>	1	8	6	3	7	6	8	39
49. <u>Notropis spilopterus</u>	2	8	4	4	8	4	8	38
50. <u>Pomoxis nigromaculatus</u>	2	4	6	4	7	6	8	37
51. <u>Catostomus commersoni</u>	1	2	7	4	8	6	8	36
52. <u>Micropterus dolomieu</u>	1	1	6	2	8	8	8	34
53. <u>Fundulus kansae</u>	4	8	4	0	8	0	8	32
54. <u>Esox lucius</u>	1	2	5	3	7	6	7	31
55. <u>Salmo gairdneri</u>	1	1	2	2	2	8	8	24
56. <u>Salmo trutta</u>	1	1	1	1	1	8	8	21
57. <u>Dorosoma cepedianum</u>	1	1	1	1	4	4	8	20

Table 6  
Recommended Fish Species Previously  
Used in Bioassay Research

Species	Lennon & Walker (1964) <sup>249</sup>	Hunn et. al. (1968) <sup>142</sup>	Mount (1968) <sup>90</sup>	Battelle (1971) <sup>105</sup>
Atlantic salmon				X
<u>Salmo salar</u>				
Threadfin shad			1	
<u>Dorsoma pentenense</u>				
Sea lamprey				X
<u>Petromyzon marinus</u>				
Rainbow trout	X	X	1	X
<u>Salmo gairdneri</u>				
Brown trout		X		X
<u>Salmo trutta</u>				
Brook trout	X	X	1	X
<u>Salvelinus fontinalis</u>				
Lake trout			2	
<u>Salvelinus namaychus</u>				
Coho salmon			2	X
<u>Oncorhynchus kisutch</u>				
Chinook salmon				X
<u>Oncorhynchus tshawytscha</u>				
Lake herring			2	
<u>Coregonus artedii</u>				
Mountain whitefish			2	
<u>Prosopium williamsoni</u>				
White sucker	X	X	1	X
<u>Catostomus commersoni</u>				
Bigmouth buffalo	X			
<u>Ictiobus bubalus</u>				
Goldfish	X	X	1	X
<u>Carassius auratus</u>				
Emerald shiner			1	
<u>Notropis atherinoides</u>				
Fathead minnow	X	X	1	X
<u>Pimephales promelas</u>				
Bluntnose minnow				X
<u>Pimephales notatus</u>				
Golden Shiner	X			X
<u>Notemigonus chrysoleucas</u>				

(Continued)

Note: 1 = primary list, all pollutants.  
2 = special list, for selected pollutants.

Table 6 (Concluded)

Species	Lennon & Walker (1964) <sup>249</sup>	Hunn et. al. (1968) <sup>142</sup>	Mount (1968) <sup>90</sup>	Battelle (1971) <sup>105</sup>
Red-sided Shiner <u>Phoxinus phoxinus</u>				X
Creek chub <u>Semotilus atromaculatus</u>				X
Channel catfish <u>Ictalurus punctatus</u>	X	X	1	X
Black bullhead <u>Ictalurus melas</u>	X			X
Yellow bullhead <u>Ictalurus natalis</u>				X
Brown bullhead <u>Ictalurus nebulosus</u>	X	X		X
Mosquitofish <u>Gambusia affinis</u>				X
Guppy <u>Lebistes reticulatus</u>				X
White bass <u>Roccus chrysops</u>			1	
Yellow perch <u>Perca flavescens</u>	X	X	1	
Walleye <u>Styostedion vetreum</u>	X	X	2	X
Green sunfish <u>Lepomis cyanellus</u>	X	X		X
Bluegill <u>Lepomis macrochirus</u>	X	X	1	X
Pumpkinseed <u>Lepomis gibbosus</u>	X			X
Redear sunfish <u>Lepomis microlophus</u>				X
Longear sunfish <u>Lepomis megalotus</u>	X			
Smallmouth bass <u>Micropterus dolomieu</u>	X	X	2	X
Largemouth bass <u>Micropterus salmoides</u>		X	1	X
Northern pike <u>Esox lucius</u>	X	X	1	
American smelt <u>Osmerus mordax</u>			2	
Brook stickleback <u>Culaea inconstans</u>	X			
Threespine stickleback <u>Gasterosteus aculeatus</u>				X

Table 7

Species Recommended for Use in Bioassays  
(References 88, 89, 90, 93, 142, 245, 248, 249)

---

## Fish

Mummichog (Fundulus hereroclitus)  
 Winter flounder (Pseudopleuronectes americanus)  
 Scup (Stenotomus chrysops)  
 Atlantic salmon (Salmo salar)  
 Striped bass (Morone saxatilis)  
 Herring (Clupea harengus)  
 Milkfish (Chanos chanos)  
 Rainwater killfish (Lucania parva)  
 Atlantic menhaden (Brevoortia tyrannus)  
 Summer flounder (Paralichthys dentatus)  
 Croaker (Micropogon undulatus)  
 Striped mullet (Mugil cephalus)  
 Sailfin molly (Poecilia (Molliensia) latipinna)  
 Pinfish (Lagodon rhomboides)  
 Southern flounder (Paralichthys lethostigma)  
 Red snapper (Lutjanus campechanus)  
 Gulf menhaden (Brevoortia patronus)  
 Sheepshead minnow (Cyprinodon variegatus)  
 Bay anchovy (Anchoa (Anchoiella) mitchilli)  
 Pompano (Trachinotus carolinus)  
 Chinook salmon (Oncorhynchus tshawytscha)  
 Coho salmon (Oncorhynchus kisutch)  
 English sole (Parophrys vetulus)  
 Starry flounder (Platichthys stellatus)  
 Shiner perch (Cymatogaster aggregata)  
 Staghorn sculpin (Leptocottus armatus)  
 Pacific sardine (Sardinops caerulea)  
 Pink salmon (Oncorhynchus gorbuscha)  
 Pacific herring (Clupea pallasii)  
 Threespine stickleback (Gasterosteus aculeatus)  
 Ling cod (Ophiodon elongatus)  
 Mountain bass (Kuhlia sandvicensis)  
 Anchovy (Stolephorus purpureus)

## Crustacea

Green crab (Carcinides maenas)  
 Sand shrimp (Crangon septemspinosa)  
 Lobster (Homarus americanus)  
 Copepod (Pseudocalanus minutus)  
 Barnacles (Balanus sp.)  
 Blue crab (Callinectes sapidus)

(Continued)

(Sheet 1 of 3)

Table 7 (Continued)

Crustacea (Continued)

Calico crab (Ovalipes ocellatus)  
 Brown shrimp (Penaeus aztecus)  
 White shrimp (P. setiferus)  
 Pink shrimp (P. duorarum)  
 Spiny lobster (Panulirus argus, P. interruptus, P. japonicus,  
P. pencillatus)  
 Dungeness crab (Cancer magister)  
 Shore crabs (Hemigrapsus sp.)  
 Shrimp (Pandalus spp., Crangon spp.)  
 King crab (Paralithoides camtschatica)  
 Crab (Ranina serrata)  
 Crab (Portunus sanguinolentus)  
 Crab (Podophthalmus vigil)

Mollusca

Eastern oyster (Crassostrea virginica)  
 Soft-shell clam (Mya arenaria)  
 Ribbed mussel (Volvella demissa)  
 Quahog (Mercenaria mercenaria)  
 Surf clam (Spisula solidissima)  
 Bay scallop (Aequipecten irradians)  
 Calico scallop (A. gibbus)  
 Abalone (Haliotis spp.)  
 Little neck clam (Protothaca staminea)  
 Razor clam (Siliqua patula)  
 Pismo clam (Tivela stultorum)  
 Bay mussel (Mytilus edulis)  
 Sea mussel (M. californianus)  
 Pacific oyster (Crassostrea gigas)  
 Ribbed limpet (Siphonaria normalis)  
 Saltwater limpets (Helcioniscus exaratus, H. argentatus)  
 Cone shells (Conus spp.)

Other Invertebrates

Bloodworm (Glycera dibranchiata)  
 Sandworms (Nereis virens, N. vexillosa)  
 Sea urchin (Arbacia punctulata)  
 Sea urchins (Lytechinus spp., Echinometra spp.)  
 Sea urchin (Strongylorentrotus purpuratus)  
 Sea anemone (Anthopleura elegantissima)

Thalloid Algae

Sea moss (Chondrus crispus)  
 Sea lettuce (Ulva spp.)

(Continued)



Table 7 (Concluded)

---

Thalloid Algae (Continued)

Atlantic kelp (Laminaria digitata, L. agardhii)

Red seaweed (Gracilaria verrucosa, G. foliifera)

Bullwhip kelp (Nereocystis luetkeana)

Giant kelp (Macrocystis pyrifera)

Unicellular Algae

Local species of genera: Cyclotella, Navicula, Nitzschia, Chaetoceros  
and others

In accordance with letter from DAEN-RDC, DAEN-ASI dated 22 July 1977, Subject: Facsimile Catalog Cards for Laboratory Technical Publications, a facsimile catalog card in Library of Congress MARC format is reproduced below.

Rosenberger, David R

Considerations in conducting bioassays / by David R. Rosenberger ... [et al.], Bioassay Laboratory, WAPORA, Inc., Charleston, Illinois. Vicksburg, Miss. : U. S. Waterways Experiment Station ; Springfield, Va. : available from National Technical Information Service, 1978.

127, [17] p. : 27 cm. (Technical report - U. S. Army Engineer Waterways Experiment Station ; D-78-23)

Prepared for Office, Chief of Engineers, U. S. Army, Washington, D. C., under Contract No. DACW 39-73-C-0134 (DMRP Work Unit No. 1D02)

References: p. 69-127.

1. Bioassay. 2. Freshwater fishes. 3. Toxicity. I. United States. Army. Corps of Engineers. II. Wapora, Inc. Bioassay Laboratory. III. Series: United States. Waterways Experiment Station, Vicksburg, Miss. Technical report ; D-78-23.  
TA7.W34 no.D-78-23